

YEAR END TECHNICAL REPORT

May 18, 2012 to August 17, 2013

Chemical Process Alternatives for Radioactive Waste

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Addendum:

This document represents one (1) of five (5) reports that comprise the Year End Reports for the period of May 18, 2012 to July 17, 2013 prepared by the Applied Research Center at Florida International University for the U.S. Department of Energy Office of Environmental Management (DOE-EM) under Cooperative Agreement No. DE-EM0000598.

The planned period of performance for FIU Year 3 under the Cooperative Agreement was May 18, 2012 to May 17, 2013. However, two no-cost extensions have been executed by DOE-EM. The first no-cost extension was received from DOE on 05/17/13 to extend the end of the period of performance for a period of two months (until 07/17/13). Another two months no-cost extension was received from DOE on 07/10/13 to extend the end of the period of performance to 9/16/13. The activities described in this report are for the FIU Year 3 period of performance from May 18, 2012 to August 17, 2013.

The complete set of FIU's Year End Reports for this reporting period includes the following documents:

1. Chemical Process Alternatives for Radioactive Waste
Document number: FIU-ARC-2013-800000393-04b-213
2. Rapid Deployment of Engineered Solutions for Environmental Problems at Hanford
Document number: FIU-ARC-2013-800000438-04b-217
3. Remediation and Treatment Technology Development and Support
Document number: FIU-ARC-2013-800000439-04b-219
4. Waste and D&D Engineering and Technology Development
Document number: FIU-ARC-2013-800000440-04b-216
5. DOE-FIU Science & Technology Workforce Development Initiative
Document number: FIU-ARC-2013-800000394-04b-072

Each document will be submitted to OSTI separately under the respective project title and document number as shown above.

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PROJECT 1 OVERVIEW

The Department of Energy's (DOE's) Office of Environmental Management (EM) has a mission to clean up the contaminated soils, groundwater, buildings and wastes generated over the past 60 years by the R&D and production of nuclear weapons. The nation's nuclear weapons complex generated complex radioactive and chemical wastes. This project is focused on tasks to support the safe and effective storage, retrieval and treatment of high-level waste (HLW) from tanks at the Hanford and Savannah River sites. The objective of this project is to provide the sites with modeling, pilot-scale studies on simulated wastes, technology assessment and testing, and technology development to support critical issues related to HLW retrieval and processing. Florida International University (FIU) engineers work directly with site engineers to plan, execute and analyze results of applied research and development.

During FIU Year 3, Project 1, titled "Chemical Process Alternatives for Radioactive Waste", focused on four tasks related to HLW research at FIU. These tasks are listed below and this report contains a detailed summary of the work accomplished for FIU Year 3.

Task 2 - Waste Slurry Transport Characterization: The objective of this task is to qualify (test & evaluate) pipeline unplugging technologies for deployment at the DOE sites. Additionally, FIU has worked closely with engineers from Hanford's Tank Farms and Waste Treatment and Immobilization Plant on developing alternative pipeline unplugging technologies. After extensive evaluation of available commercial unplugging technologies in the previous years, two novel approaches are being developed at FIU including a peristaltic crawler and an asynchronous pulsing method.

Task 12 - Multiple-Relaxation-Time Lattice Boltzmann Model for Multiphase Flows: The objective of this task is to develop stable computational models based on the multiple-relaxation-time lattice Boltzmann method. The computational modeling will assist site engineers with critical issues related to HLW retrieval and processing that involves the analysis of gas-fluid interactions in tank waste.

Task 15 - Evaluation of Advanced Instrumentation Needs for HLW Retrieval: This task has evaluated the maturity and effectiveness of commercial and emerging technologies capable of addressing several instrumentation needs for HLW feed mixing and retrieval. Promising candidate technologies have been evaluated for their functional and operational capabilities and an ultrasonic spectroscopy system was selected for additional testing to determine its effectiveness in the harsh chemical tank environments.

Task 16 - Computational Simulation and Evolution of HLW Plugs: The objective of this task is to develop a model that is aimed at predicting the formation of plugs in HLW lines with an emphasis on the effects of pipeline geometry. The goal is to develop a tool that can assist in providing guidelines to plug prevention.

TASK 2: WASTE SLURRY TRANSPORT CHARACTERIZATION

EXECUTIVE SUMMARY

In previous years, Florida International University (FIU) has tested and evaluated a number of commercially available pipeline unplugging technologies. Based on the lessons learned from the evaluation of the technologies, two alternative approaches have been developed by FIU. These are an asynchronous pulsing system (APS) and a peristaltic crawler. The APS is based on the principle of creating pressure waves in the pipeline filled with water from both ends of the blocked section in order to break the bonds of the blocking material with the pipe wall via forces created by the pressure waves. The waves are created asynchronously in order to shake the blockage as a result of the unsteady forces created by the waves. The peristaltic crawler is a pneumatically operated crawler that propels itself by a sequence of pressurization/depressurization of cavities (inner tubes). The changes in pressure result in the translation of the vessel by peristaltic movements.

For this performance period, experiments were conducted to validate the asynchronous pulsing system's ability to unplug a large-scale pipeline testbed and compare the performance of the APS to the data obtained from the testing conducted on small-scale testbeds.

The unplugging experiments consisted of placement of 3-ft kaolin-plaster plugs within a test pipeline loop which consists of two 135-foot runs on either side of a plug and using the system to unplug the pipeline. The results obtained during the experimental phase of the project are presented which include pressures and vibration measurements that capture the propagation of the pulses generated by the system.

During this performance period, efforts also focused on the engineering scale testing of the peristaltic crawler system (PCS). Design improvements were implemented to increase the navigational speed of the crawler. The pneumatic valves were relocated from the control box to a trailing capsule 1 ft from the rear of the crawler and the outer bellow was replaced with a thinner wall bellow of similar dimensions. These changes improved the speed of the crawler from 1 ft/hr to 38 ft/hr. Bench scale navigational tests conducted using a 90° elbow showed a time of 11 min for the crawler to clear the elbow. Force tests conducted to determine the forward force generated by the crawler yielded a maximum force of 108 lb with a supply pressure of 50 psi.

An engineering scale testbed, with a total length of 430 ft, was assembled and navigational tests were conducted to evaluate the PCS. It was observed that the inner diameter of schedule 10 pipe sections causes the flexible cavities of the crawler to over expand resulting in a drastic decrease in the fatigue life of the cavities. Options to increase the fatigue life of the cavities were evaluated and it was found that increasing the distance between the clamps to 1 in provides a total of 15,000 cycles prior to failure. It is estimated that a total of approximately 3,600 cycles is required for the crawler to completely navigate in the 430 ft testbed. Pull force tests conducted by manually routing the tether through the pipeline showed that a force of 8 lb is required for every 21 ft straight section of pipeline after the first 21 ft section. A stainless steel wire was wound around the tether to decrease the friction force and contact area between the tether and the pipeline.

INTRODUCTION

Pumping high-level waste (HLW) between storage tanks or treatment facilities is a common practice performed at Department of Energy (DOE) sites. Changes in the chemical and/or physical properties of the HLW slurry during the transfer process may lead to the formation of blockages inside the pipelines. Current commercially available pipeline unplugging technologies do not provide results that are cost-effective and reliable. As part of the research objectives at FIU, novel pipeline unplugging technologies that have the potential to efficiently remediate cross-site and transfer line plugging incidents are being developed. These are an asynchronous pulsing system (APS) and a peristaltic crawler. Both technologies been developed over the past few performance periods and details of the operational principles and history of the project can be obtained in (Roelant, 2012) This report presents details of the devices and procedures used for the experimental testing of the two technologies for FIU Year 3. The first section pertains to the experimental testing of the APS followed by the experimental testing of the peristaltic crawler.

Initially, a brief background on the principles on which the APS technology is based is provided. Previous studies have demonstrated on a lab scale testbed how operational process parameters can be optimized [1]. During FIU Year 3, additional parameterization was conducted as well as testing on plugged lines. An engineering scale testbed was designed and fabricated with modifications that provided a more realistic test bed.

In addition, this report presents the design improvements and experimental testing results of the Peristaltic Crawler System (PCS). It provides the experimental results that show the necessary requirements that will enable the PCS to conduct unplugging and inspection operations in a 430-ft pipeline. The general configuration of the system remained similar to that of the unit tested previously [1] but changes were implemented to improve the crawler's navigational speed and also to equip the system with a camera that provides visual feedback of the conditions inside the pipeline. Additionally, a 430-ft engineering testbed was designed and assembled and the system was placed in a compact platform for easy deployment. Preliminary navigational tests in the engineering scale testbed were conducted and a number of issues associated with operating the system in longer pipelines were revealed. Proposed solutions to the issues are presented and conclusions and recommendations for further improvement of the system are provided.

EXPERIMENTAL TESTING OF THE ASYNCHRONOUS PULSING SYSTEM

Background

In order to clear plugged radioactive waste transfer lines, non-invasive techniques can have significant advantages since problems such as contamination clean-up and exposure to radioactive waste of invasive devices can be avoided. During previous work, FIU evaluated two technologies that fall into this category, namely, NuVision's wave erosion method and AIMM Technologies' Hydrokinetics method. These technologies fill the plugged pipeline with water up to an operating pressure level and induce a pressure variation at the inlet of the pipeline to dislodge the plug. Using the experience obtained during experimental evaluations of both technologies, FIU has developed a non-invasive unplugging technology called the Asynchronous Pulsing System (APS) that combines the attributes of previously tested technologies. A pipeline unplugging technology using similar principles for generating pressure pulses in pipelines has previously been tested at the Idaho National Laboratory (INL) by Zollinger and Carney [2]. The most relevant difference of the current technology from the unplugging method developed at INL is that both sides of the pipeline are used to create the asynchronous pulsing in the current technology. Figure 1 shows a sketch of how this technology can be utilized for a typical plugging scenario. During last year's work, the asynchronous pulsing system's ability to unplug a pipeline on a small-scale testbed was validated. Part of this year's work concentrated on applying the information obtained from last year's work to unplug a large scale pipeline. The experiments consisted of placement of 3-ft kaolin-plaster plugs within a test pipeline loop and using the system to unplug the pipeline.



Figure 1. Pipeline unplugging scenario in a horizontal pipe.

The APS is based on the idea of creating pressure waves in the pipeline filled with water from both ends of the blocked section in order to dislodge the blocking material via forces created by the pressure waves. The waves are generated asynchronously in order to break the mechanical bonds between the blockage and the pipe walls as a result of the vibration caused by the unsteady forces created by the waves.

General Description

The asynchronous pulsing system uses a hydraulic pulse generator to create pressure disturbance that dislodge blockages within the pipeline. The test pipeline loop contains two identical 135-foot pipeline sections with a 3-foot plug between them. This test loop allows for control of the individual pipeline section pulse characteristics to determine how each pulse influences the total plug dynamic loading. The pipes used for the loop are 3-inch diameter schedule-40 threaded carbon-steel pipes. The heart of the asynchronous pulsing system is two hydraulic piston pumps that are powered by the pulse generation unit which is comprised of hydraulic power unit and two electronically controlled high-speed valves. The pipeline is instrumented with accelerometers and pressure transducers located at strategic locations to capture the changes of the induced disturbances inside the pipeline.

RESULTS - ASYNCHRONOUS PULSING SYSTEM

During the previous year’s work, it was observed that the system’s performance was hindered if any amount of air was present in the pipeline. This year’s work began with investigating, eliminating or minimizing the effect of air in the pipeline by utilizing vacuum or high static pressure. Vacuum tests were executed to determine the relationship between vacuum pressure and residual air in the pipeline. These tests were useful in determining the necessary vacuum required in order to mitigate any entrained air in the pipeline prior to flooding. As can be seen in Figure 2, the longer that vacuum is applied to the pipeline, the amount of compression required to increase the pressure in the pipeline is reduced. This is due to air being removed from the pipeline. The 8-hour test however shows an increase in the amount of compression required when compared to the 2 & 4 hours tests. It is believed that this is due to the residual water remaining in the pipeline being boiled off and thus filling the pipeline with water vapor which is compressible.

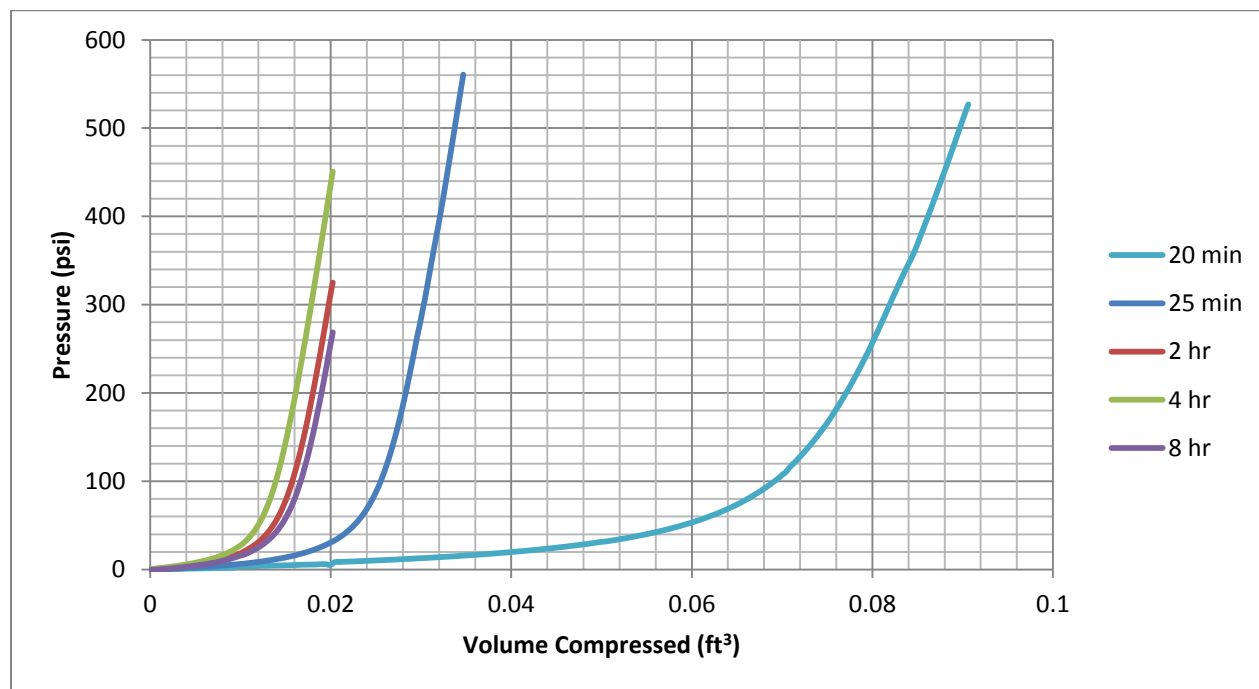


Figure 2. Vacuum test results.

Upon evaluating the vacuum tests, it became apparent that pre-vacuuming a pipeline is not the most efficient method of mitigating the effects of air in the pipeline. Therefore, pre-pressurization of the pipeline method was investigated to determine how effective increasing the static pressure inside the pipeline is in mitigating the effects of air in the pipeline. Tests were conducted on various quantities of air in the pipeline varying from 0.0 – 5.0%. Calibration curves were developed for the various air quantities. The test results indicated that the effects of an air bubble entrained in the system on a transient pulse can be effectively mitigated by increasing the static pressure of the water above atmospheric pressure. As is visible in Figure 3, the effect of air being present in the pipeline caused very limited pressurization of the loop as the air is compressed. When that bubble is pre-compressed to a certain density, the system begins to

respond to a volume reduction in a similar manner to a no-air system. This behavior can be exploited to ensure larger pressure changes (i.e. more force loading) at the plug face. The observed relationship can now be used to determine the minimum static pressure required so that the air/water mixture can act as a water-only system. The pre-pressurized value should be the minimum at which the compressibility of the system resembles the compressibility of water. These relationships can now be used to establish how much air can remain in longer pipe loops, and what will be the minimum static pressure required to cause significant dynamic loading of the plug during pulsing.

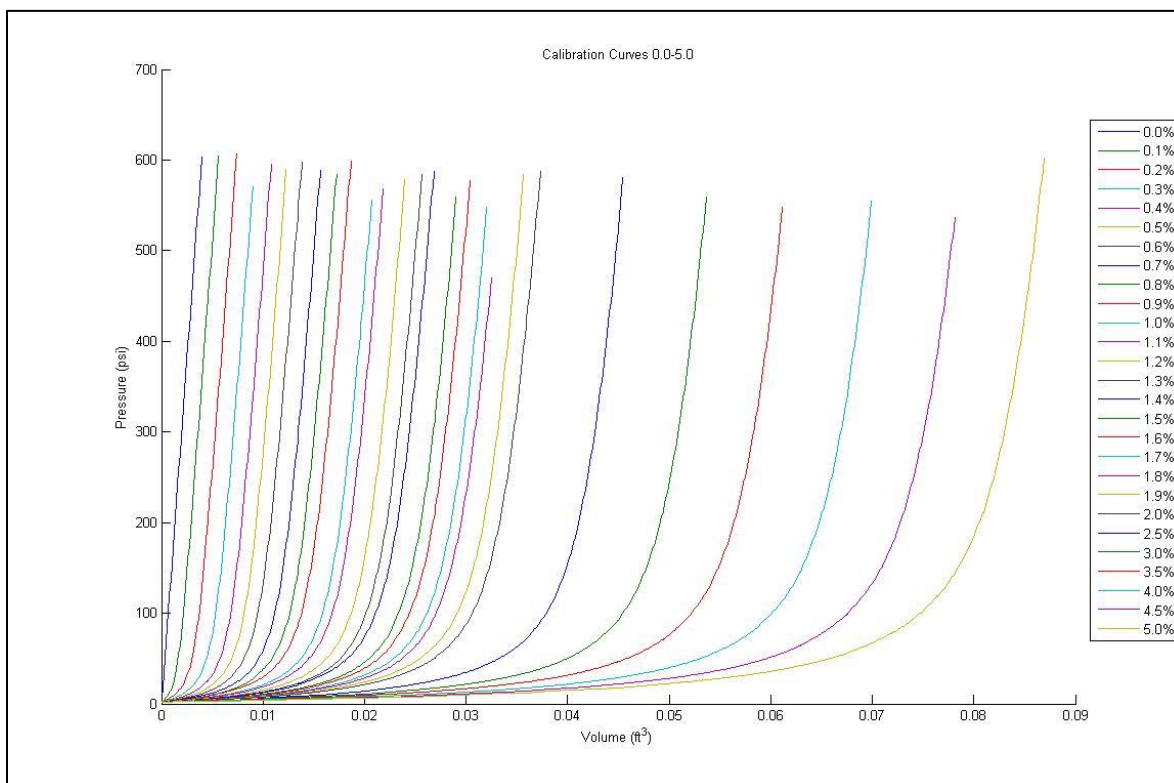


Figure 3. Calibration curves for various air quantities.

Alternate pulsing system

Research began on the development of an alternate pulsing system. Initial tests were conducted on a small positive displacement pump arranged with two solenoid valves; one on the discharge side of the pump and another on the suction side. By sequentially operating the valves, a positive pressure pulse is delivered to the pipeline that is then followed by a suction phase. The benefit of utilizing a pump as the pressure source instead of a piston is that the pump will provide more flexibility in the volumetric flow during the pressure and suction phases. Results from the initial tests proved to be successful and work began on configuring a large progressive cavity pump to perform large scale testing on the pipeline. A large progressive cavity pump was plumbed to be used as an alternative pressure source to perform asynchronous pulsing tests on the pipeline. A side-by-side comparative test was conducted on both the alternative pressure source (large progressive cavity pump) and the current asynchronous pulse unit. It was determined that they behave similarly; however, as can be seen by Figure 4, the alternative pressure source was

limited to only 150 psig. This can be overcome by utilizing a larger capacity pump. We investigated upgrading the current pump to increase its capacity and it was determined that it would cost the same as purchasing a new larger capacity pump.

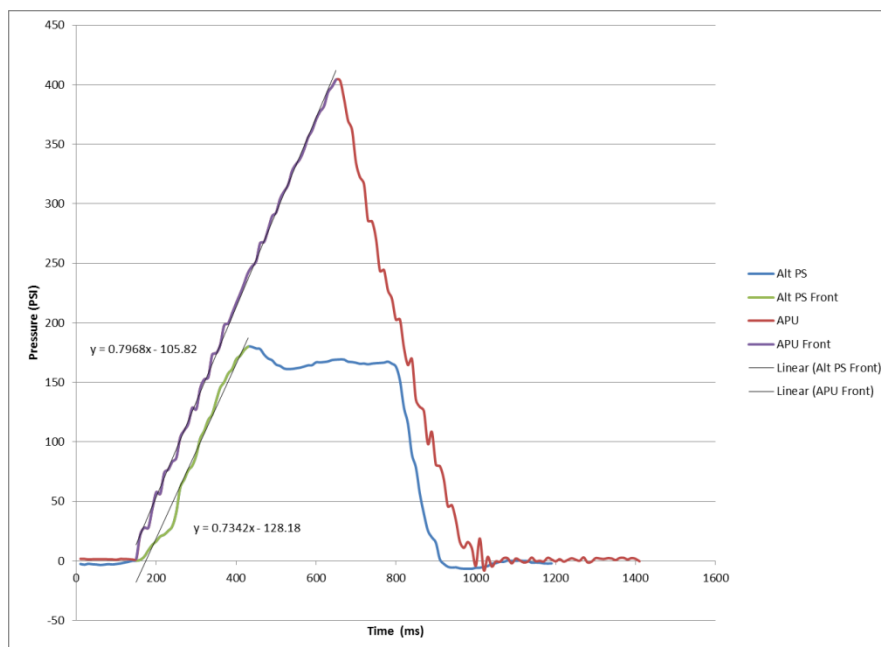


Figure 4. APU and alternative pressure source pulse test.

Plug Development and Characterization

The main objective behind the plug development was to create a high shear strength 3-ft plug that could withstand a maximum static pressure of 300 psi. In the past, several plugs made from k-mag, kaolin-bentonite, sodium-aluminum-silicate, etc. were tested. In some cases a longer length plug had to be made (for example kaolin-bentonite, chemical plug) to withstand the desired maximum static pressure. The k-mag plug could meet the 300 psi limit; however, was very difficult to make and test. Hence a need existed to develop a plug that could withstand the maximum pressure and was only 3-ft long. The kaolin-plaster plug of various compositions [3] was reviewed that mimicked the physical behavior of actual waste. The kaolin clay though being a low strength material was selected as it has excellent wall adherence property. On the other hand, the plaster of paris provided the needed high strength at low solids concentration. When combined together at varying weight percent (wt%), the kaolin-plaster plug yielded shear strengths ranging from 5-250 kPa. The plug was also easy to make and had the texture of a whipping cream which made it easier to scoop it in the pipe before the initiation of the curing process of the plug.

Several 1-L samples were created in a laboratory setting for a combination of kaolin and plaster ratios. The samples were cured for 24-36 hours and tested for shear strength by a hand held shear vane (

Table 1). The shear vane was inserted into the prepared sample before the curing process started to get accurate shear strength results.

Table 1. Plug Composition and Properties

No.	Material	Composition (wt%)	Shear Strength (kPa)
1	Kaolin	50	2±0.30
	Plaster	12	
	Water	38	
2	Kaolin	30.0	28±4
	Plaster	27.5	
	Water	42.5	
3	Kaolin	40.0	135±25
	Plaster	22.5	
	Water	37.5	
4	Kaolin	35	225±25
	Plaster	30	
	Water	35	
5	Kaolin	30	>250 ¹
	Plaster	35	
	Water	35	

It was observed that the success of making this plug was highly affected by the plug procedure and conditions. For example the mixing time yielded different shear strengths for the same composition. Hence, best practices relevant to the manufacturing of the plug were discussed with an AEM consulting engineer and the recommendations were incorporated while developing the plug.

Pressure blowout tests were conducted on a variety of 3-ft kaolin-plaster plugs made from different recipes and with different curing times (Figure 5). The purpose of these tests was to verify that the plugs could withstand a maximum static pressure of 300 psi. The optimal plug was found to be 30% kaolin, 35% plaster and 35% water (by weight) with a 4-day curing time. This combination provided a blowout pressure between 400 and 420 psi.

¹ The hand held shear vane was limited to 250 kPa shear strength measurements.



Figure 5. Pressure blowout tests.

Engineering scale asynchronous pulsing unplugging trials

A test loop was designed to test the APS and its functionality on a large scale, not previously tested. Figure 6 shows the piping and instrumentation diagram of the engineering scale loop which consists of two 135-foot runs on either side of a plug.

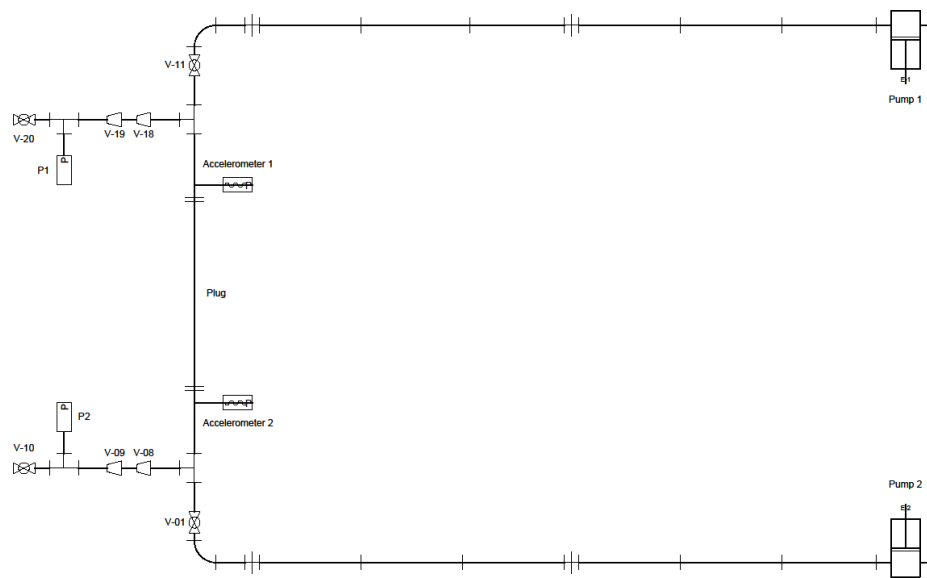


Figure 6. Engineering scale asynchronous pulsing test loop piping and instrumentation diagram.

The outdoor site was cleared and leveled and the pipeline supports were fabricated and anchored to the ground. The elevations of the pipeline supports were surveyed and adjusted to provide the proper pipeline slope. As can be seen in Figure 7, the hydraulic power unit was placed inside a shed to protect it from the rain.



Figure 7. Engineering scale testbed images for asynchronous pulsing system.

To prepare for the parametric testing of the system, data acquisition software changes were made. During the initial testing, some hardware problems were encountered concurrent with placing the system in an outdoor wet environment. Even though the new pressure transducers were verified to be working properly after installation, during initial system tests several transducers failed. It is believed that water intrusion is the issue since it had rained in the days leading up to the failures. Furthermore, since the transducer connectors had been previously used, they did not form a very tight seal around the transducer terminals; therefore, each connector was filled with non-dielectric silicone.

Additional issues related to the pump’s linear position transducers were also observed. Both linear position transducers failed, with the cause believed to be water damage. Both linear transducers were sent to the manufacturer for assessment and repair. It was determined that the damage was due to either “an external overvoltage [or] stray power applied to the ground”. During our failure analysis, the step-down transformer was also found to be damaged. It had failed, delivering only half of the designed voltage. The transformer was also equipped with a voltage rectifier, thus, eliminating the overvoltage scenario as a possible cause. Corrective

actions were taken to eliminate the step-down transformer from the control/data acquisition system circuit. In addition, the control/data acquisition system was directly connected to a grounded power supply with a surge protector. To eliminate any future water infiltration, new gaskets and connectors for the pressure transducers were also purchased and miniature enclosures were constructed for each location where the transducers are exposed to the elements. After the pumps' repaired linear position transducers were received, they were installed and the wiring was tested. Upon testing of the electrical wiring for the linear position transducers, a stray power-to-ground was discovered and repaired. This coincides with the findings of the manufacturer of the linear position transducers. Once repaired, the transducers were wired and found to be operating correctly.

Once all the electrical issues were addressed, test finally began on the engineering scale loop. Initial tests were a single pressure pulse tests that were intended to determine the pressure pulse amplification within the pipeline. Tests were conducted at both 0 and 50 psi static pressures.

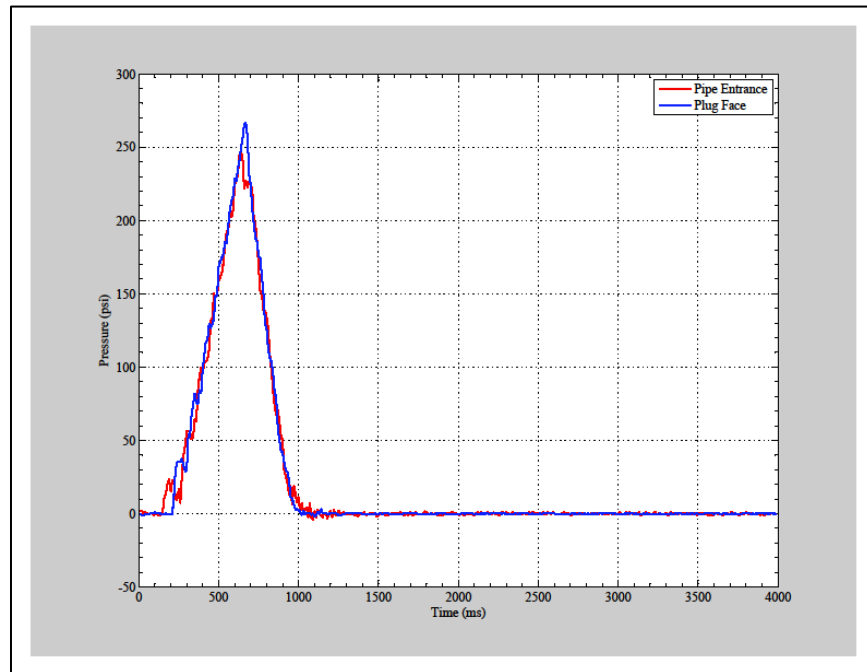


Figure 8. Single pressure pulse results with a 0 psi static pressure.

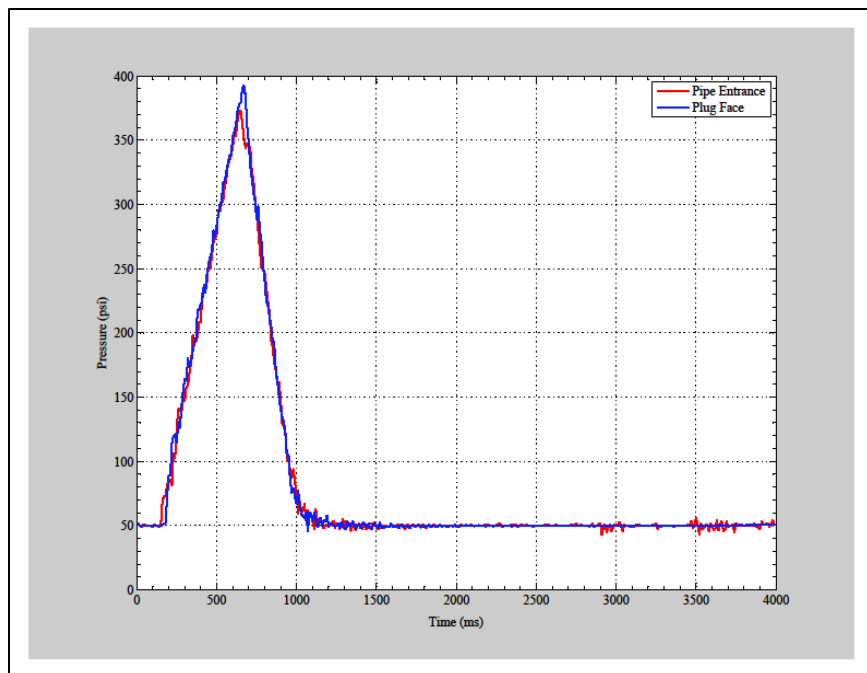


Figure 9. Single pressure pulse results with a 50 psi static pressure.

As can be seen in Figure 8, a 30-psi pressure amplification was observed with a 270-psi max pressure when the static pressure was set to zero. However when the static pressure was increased to 50-psi, as shown in Figure 9, only a 20-psi pressure amplification is obtained but the max pressure increased to 390-psi. This is due to the air in the pipeline being compressed. As observed in the previous tests, when the air is compressed, the damping effects of the air is minimized and thus the higher pressure pulse is obtained.

The next phase of tests involved the actual unplugging of 3-ft kaolin-plaster plugs. Figure 10 shows kaolin-plaster plug before being unplugged.



Figure 10. Kaolin-plaster plug before and after unplugging.

Figure 11 shows the pressure data from an unplugging of a 3-ft kaolin-plaster plug. P1 and P2 represent the pressures on each side of the plug face. For these tests, a static pressure of approximately 55 psi was utilized. The majority of air was removed from the pipeline, with residual amounts remaining due to limitations created by the plug. Frequencies were varied from

0.5 to 2 Hz, however, transducer issues arose at 2 Hz. Resulting pulse amplitudes were dependent on valve opening, frequency/piston speed and air entrainment. After analyzing the data it was observed that P2 had a smaller pressure range than P1 for the 1.5 Hz trial. This is believed to be due to the P2 side of the pipe loop containing more air than the P1 side. Another observation was that just before unplugging occurred P1 started to increase as P2 was decreasing. This is due to water starting to leak past the plug from the P2 side to the P1 side. The decrease in the water volume from one side will cause the pressure per stroke to decrease while the increase of water volume on the other side will cause the pressure per stroke to increase.

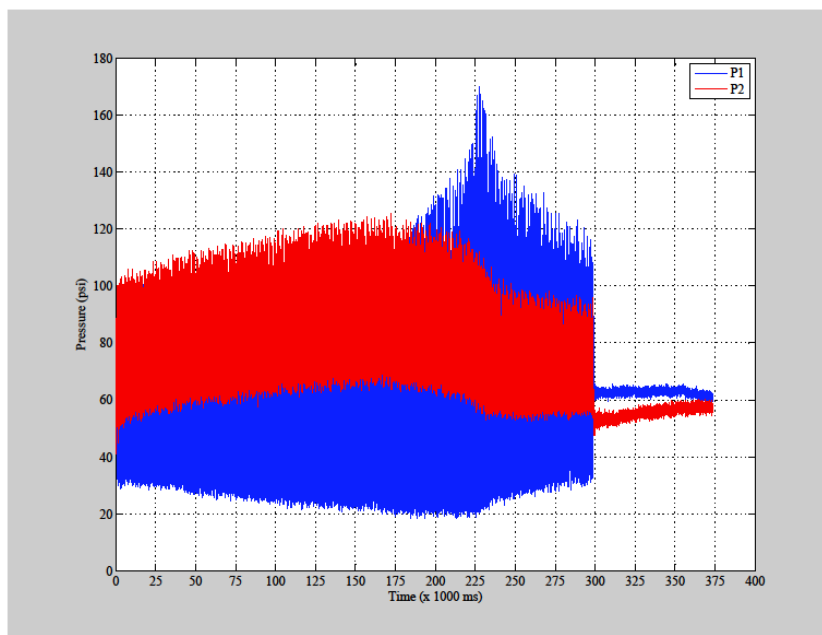


Figure 11. Pressure data from an unplugging of a kaolin-plaster plug at 1.5 Hz.

Figure 12 shows the pressures for a 0.5 Hz trial. As with the previous test, P2 had a larger pressure range than P1. This is also believed to be due to air still remaining in the P1 side of the loop. Also in this test, water started to leak past the plug from the P1 side to the P2 side which resulted in P2 starting to increase as P1 was decreasing just before unplugging occurred.

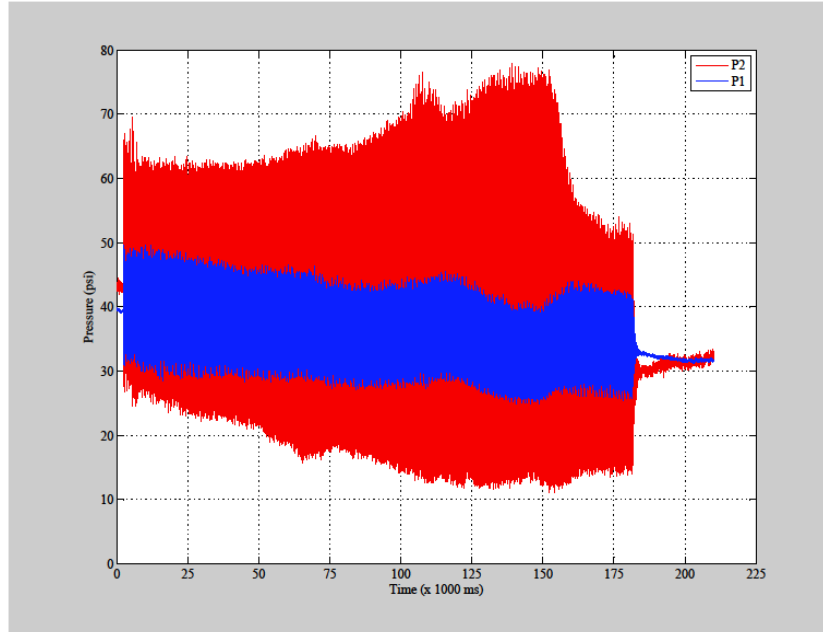


Figure 12. Pressure data from an unplugging of a kaolin-plaster plug at 0.5 Hz.

Figure 13 shows the data from the accelerometers mounted on the pipeline on both sides of the plug for the 0.5 Hz trial. As can be seen from the graph, the force applied to the pipeline started to increase as just before the plug was unplugged and reached its maximum when unplugging occurred.

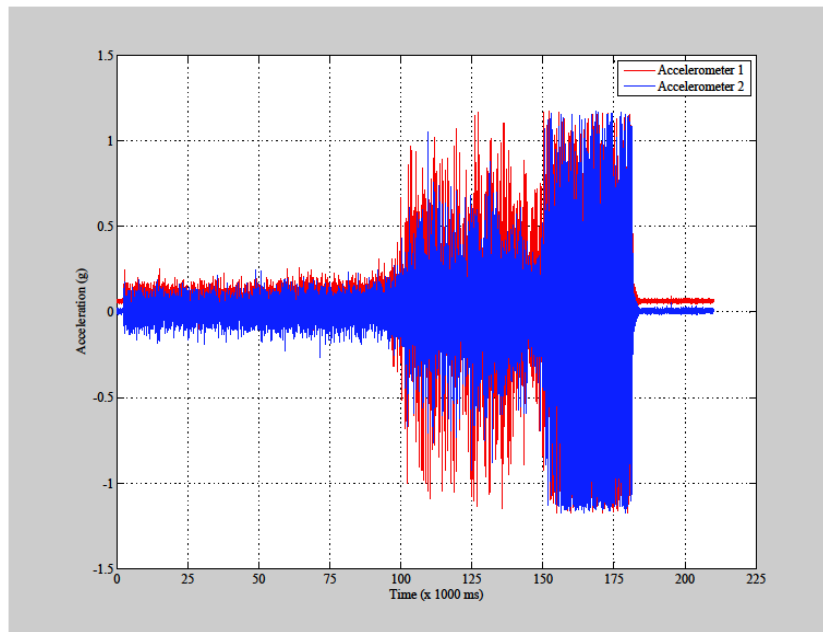


Figure 13. Pipeline accelerometer data during a 0.5 Hz unplugging.

Discussion

After analyzing the data of the APS on a large scale loop, several observations were made. As in the previous testing done on smaller loops, air in the system has a major effect on the system's performance. Just as in the tests conducted on smaller loops, the consequence of entrained air can be mitigated by either applying a vacuum to the pipeline or increasing the system's static pressure. The data obtained from this testing indicates that increasing the static pressure is a more effective method. During previous testing, the pulsing frequency played a major role in the system's unplugging capability. However, during these tests on the larger loop this was not observed which is believed to be due to the longer pipeline length damping the frequency.

ENGINEERING SCALE PIPELINE UNPLUGGING TESTING USING THE PERSITALTIC CRAWLER SYSTEM

Background

FIU's efforts to develop innovative pipeline unplugging technologies include the development of a crawler system that can conduct unplugging operations. The PCS is a pneumatic/hydraulic unit that can navigate inside a pipeline by pressurization/depressurization of flexible cavities. To date, three generations of the PCS have been designed, assembled, and tested each one having improvements to overcome the limitations previously observed. For each generation, the tests conducted on bench-scale testbed have provided information on the navigational speed, ability to negotiate thru a 90° elbow, pull force, and unplugging ability. Design improvements to bolster its durability and have also been implemented on each generation.

The current system consists of a crawler unit, a tether-reel assembly and a control station. The unplugging tool is attached to the front of the unit and powered by pressurized water. The crawler unit has a double walled bellow assembly and a front and back rim to which flexible sleeves are clamped forming the front and back cavities. The center of the double walled bellow assembly creates a passage that allow particles of the plug that are set loose during the unplugging process to travel to the back of the unit. Attached to the front rim is a nose cap designed to hold the unplugging tools (high pressure water nozzle) and the camera. Figure 14(a) shows a rendering of the crawler unit and (b) shows an exploded view of the crawler assembly.

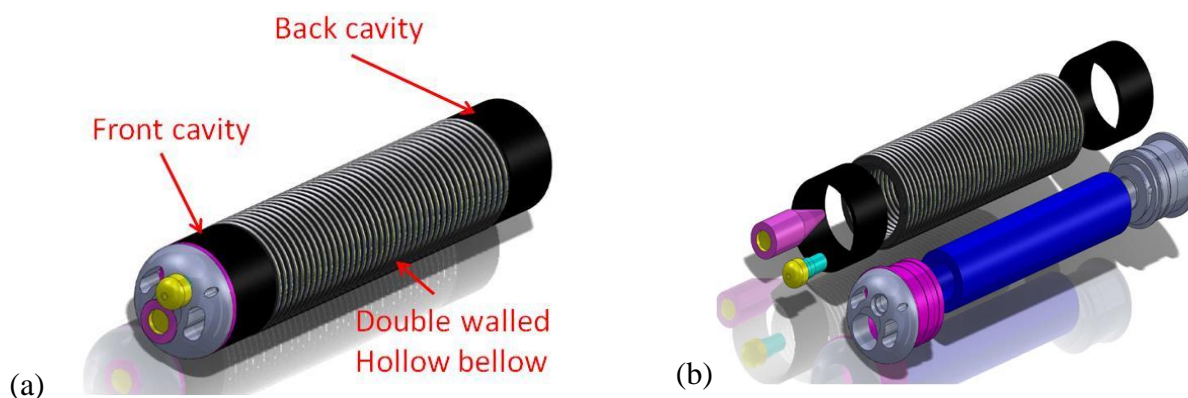


Figure 14. (a) Rendering of crawler, (b) exploded view of crawler assembly [4].

The tether connects the crawler to the control station and the power supply source. The tether consists of three pneumatic lines, one hydraulic line, and one multi-conductor cable jacketed together having a total length of 500 ft. The reel system was designed to accommodate the tether and provides rotating connections to the pneumatic, hydraulic, and electrical lines (Figure 15 (a)). The control station includes the pneumatic pressure regulators, a vacuum pump, vacuum chamber and a controller box containing a programmable logic controller (PLC) that controls the position (opened/closed) pneumatic valves of the crawler unit. By programming an appropriate sequence on the PLC, the desired motion is achieved (Figure 15 (b)).

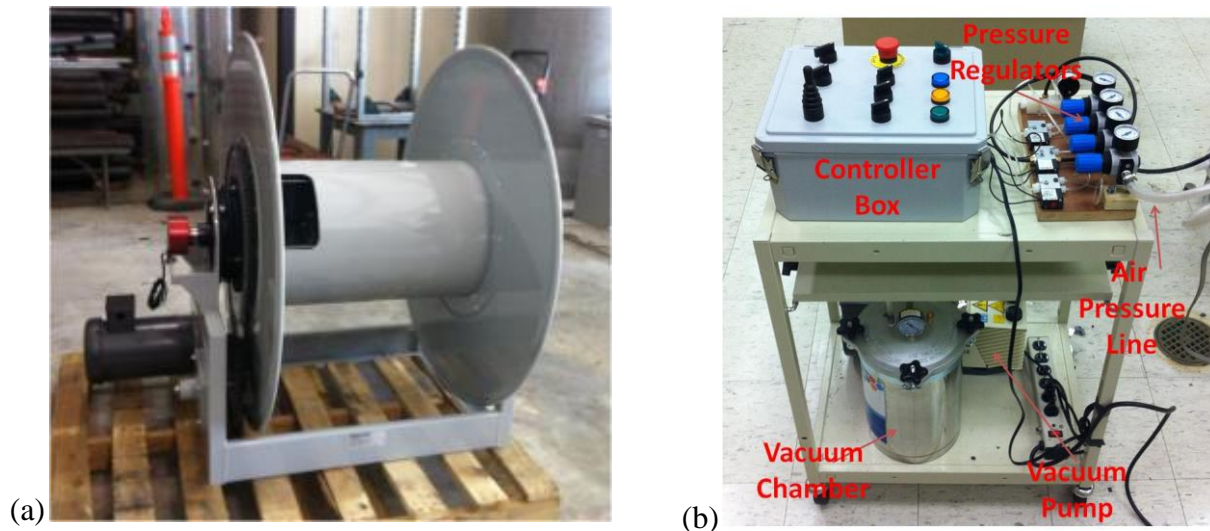


Figure 15. (a) Reel system, (b) control station.

Design Improvements

For this performance period, efforts focus on meeting the necessary requirements that will enable the PCS to conduct unplugging and inspection operations in a 430-ft pipeline. Improvements to the system include the re-design of the rims of the unit to accommodate a camera system to provide visual feedback of the conditions inside the pipeline, positioning of a pneumatic valve manifold system located in close proximity to the crawler and optimization of the double bellow assembly to reduce cycle time.

The tether was connected and wound to the reel system. The reel has rotating manifold/connectors for pneumatic, hydraulic and electrical lines. The rotating manifold consists of four 1/4" FNPT ports and 6 electrical conductors. The reel makes controlling the length of tether routed thru the inlet point easy to control and prevents the pneumatic and hydraulic lines from damage due to kinking. Figure 16 (a) shows the tether attached to the crawler, and Figure 16 (b) shows the tether wound to the reel system.

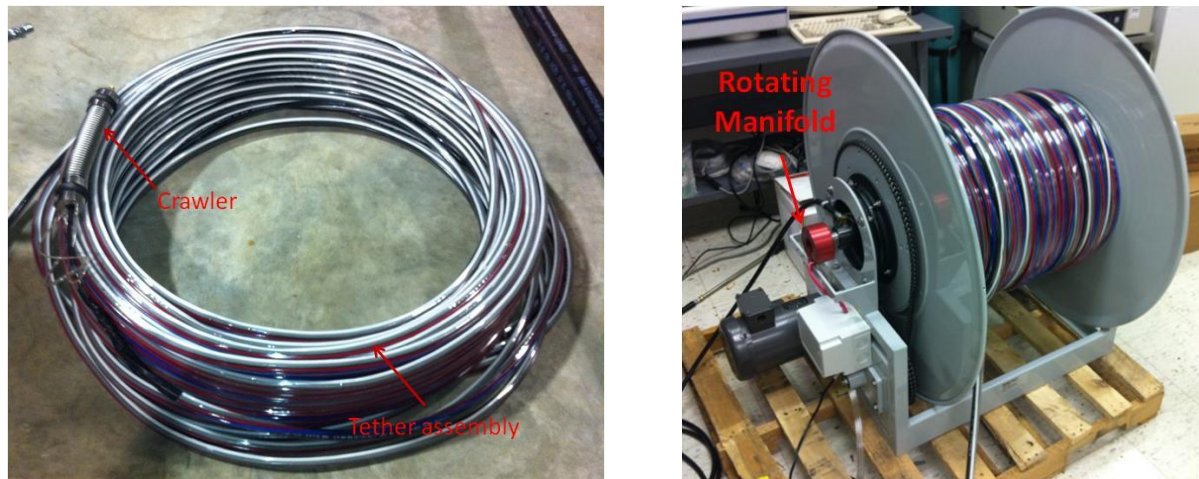


Figure 16. (a) Tether attached to the crawler, (b) tether-reel assembly.

Previous tests conducted using different tether lengths demonstrated that there is a significant reduction in speed of the crawler with increasing length of the tether. This decrease in speed is the result of the increase of volume required to be pressurized/depressurized inside the tether before the set pressure can be reached at the cavity in the crawler unit for each cycle. In previous tests, a speed of 1 ft/hr was recorded using a 500-foot tether which shows that in order for the PCS to be effective in longer pipelines, it would require design changes to reduce the cycle time.

A solution to eliminate pressurizing/depressurizing the tether with each cycle was implemented by relocating the pneumatic valves from the control to a trailing capsule located approximately 1 ft from the rear of the unit. This change allows the pneumatic lines of the tether assembly to remain at a constant pressure, making the cycle time independent of tether length. The capsule encases 3 12-volt pneumatic valves connected to the tether assembly. The valves used are single acting Bullet Valves manufacture by MAC Valves, Inc. Each of the valves is connected to the PLC using the electrical lines of the tether and each controls a separate cavity of the crawler. Figure 17 (a) shows a rendering of the design of the capsule and valves and Figure 17 (b) shows the assembled capsule manufactured using ABS (acrylonitrile butadiene styrene).

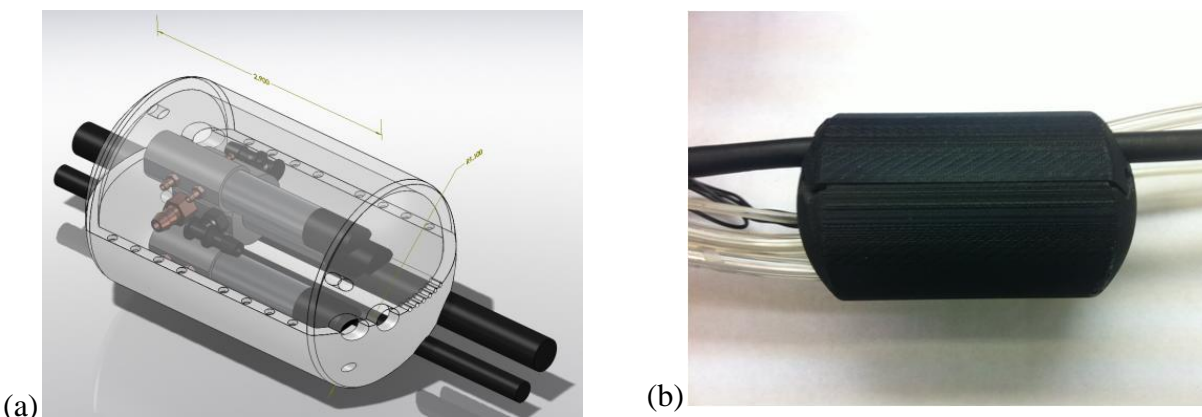


Figure 17. (a) Rendering of capsule, (b) capsule manufactured from ABS.

As noted in the introduction, an improvement to the crawler was the implementation of a visual feedback system that provides images of the conditions inside the pipeline. Several commercially available systems were evaluated based on their compatibility with the crawler requirements. The criteria for selection included that the visual feedback system should not hamper the navigational capabilities of the unit, must be small enough to be contained within the crawler, must be able to operate at a minimum distance of 1000 ft from the inlet point and must be able to operate in a HLW environment. The visual system selected was the INUKTUN Mini-Cristal Cam®. This system uses a camera that is 7/8" in diameter and connects to a display station using a fiber optic cable. It has an operational range of 1000 ft and can function in HLW environments up to 300 rad/hour. Additionally, the system is provided with a light source to illuminate dark environments. Figure 18 shows a close-up of the camera and the display station.



Figure 18. (a) Camera, (b) display station.

Using the dimension parameters from the camera, the front assembly of the unit was re-designed to accommodate both the unplugging tool and the camera. The front assembly consists of a hollow rim and a hemisphere that positions the camera and a high pressure hydraulic nozzle with enough clearance to fit inside the crawler. The high pressure nozzle selected is manufactured by Arthur Products (model AP0125RNH-01). The design process included the computer design and rapid prototyping of the front assembly to test for fit, form and function. Proper clearance between the parts of the assembly was calculated to prevent interference due to tolerance stack-up. The crawler assembly was then manufactured using 316 stainless steel. Figure 19 (a) shows the top view of the front cap, and Figure 19 (b) shows the side view of the front rim. Figure 20 (a) shows a side view of the front of the unit prior to assembly and Figure 20 (b) shows the assembled crawler.

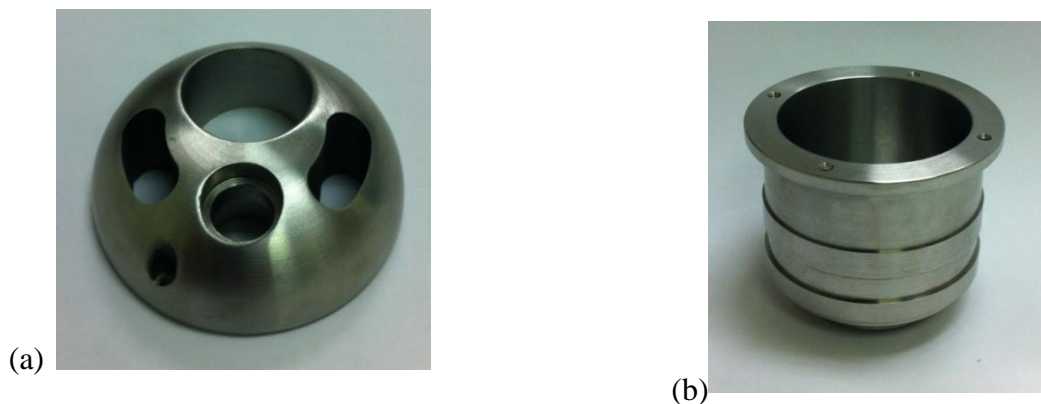


Figure 19. (a)Top view of front cap, (b) side view of front rim.

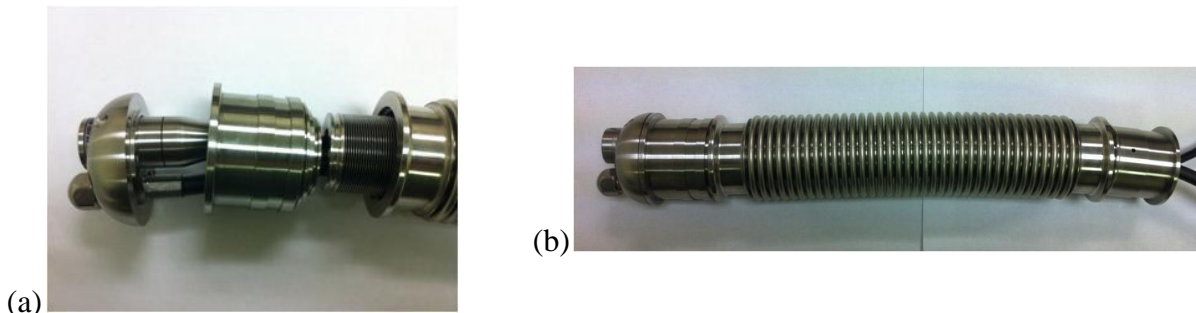


Figure 20. (a) Front assembly of the crawler, (b) side view of crawler.

In order to improve the navigational speed of the crawler, several outer bellow configurations were evaluated. One approach was to place a rubber sleeve over the outer bellow. The sleeve is

placed over the compressed bellow during assembly. This configuration provides a faster compression time of the bellow. Three sleeve thicknesses were evaluated 0.004, 0.0625, and 0.125 inches. Figure 21 shows the crawler with a 0.125 in thick sleeve. An additional configuration was to replace the outer hydroformed bellow with one having similar overall dimension but smaller thickness. A thinner wall bellow reduces the force required for the crawler to reach full compression.



Figure 21. Crawler with a 0.125 in thick sleeve.

Testbeds

The experimental testing of the PCS was conducted using a bench scale testbed and an engineering scale testbed. Similar to the testbed used previously, the bench scale testbed consisted of two 3-ft long clear PVC straight sections connected with a 90° elbow and a 3-ft long carbon steel section. All of the pipes are 3 inches in diameter. Figure 22 shows the testbed used to conduct the navigation tests.

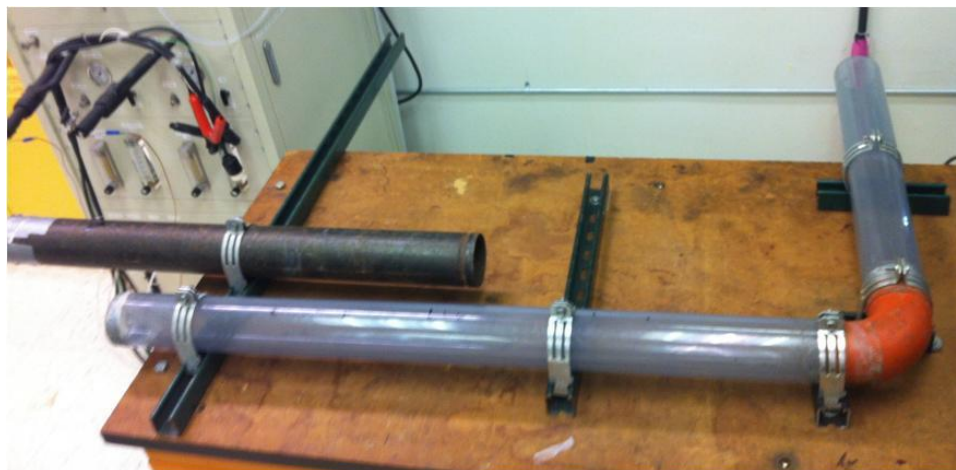


Figure 22. Bench scale testbed.

To experimentally determine the pull force that the crawler unit can generate, the crawler was placed inside a fixture consisting of a clear pipe section containing a spring of known stiffness (Figure 23). The testing procedure consisted on first pressurizing the back cavity to anchor the unit to the pipeline to then provide different pressure to the double wall assembly. The displacement of the front end of the crawler unit was recorded to calculate the force generated.



Figure 23. Fixture used to determine the crawler unit force.

The engineering scale experimental testbed consists of 24 straight sections and three 90° elbows assembled with couplings. The pipes sections used were grade 10 carbon steel pipes and had an inner diameter of 3.26 inches. The total length of the pipeline was adjusted based on the length of tether available. The final configuration was reduced from 500 ft to 430 ft to account for losses in tether length resulting from jacketing the pneumatic, hydraulic and electric lines together. The control station, air compressor, vacuum pump, tether-reel assembly, and feed-back camera monitor were placed in a container for easy deployment. Figure 24 shows the testbed layout and system configuration.

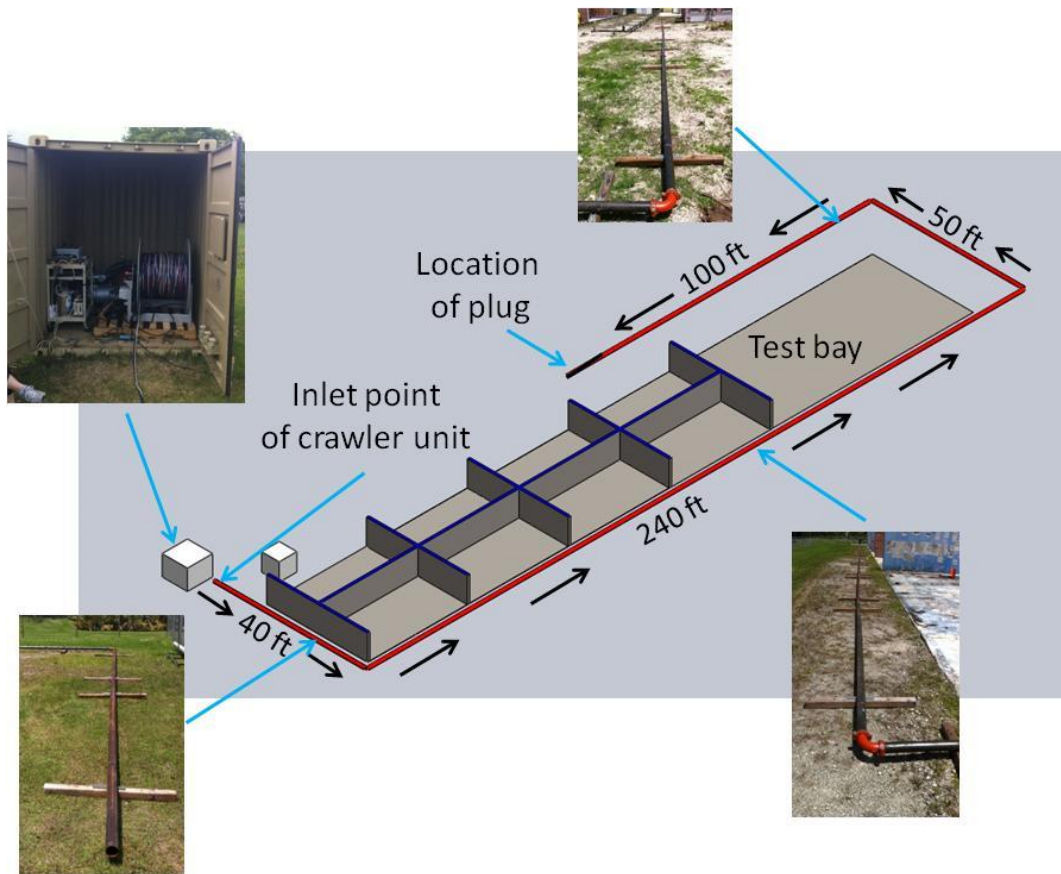


Figure 24. Engineering scale testbed configuration.

RESULTS

Navigational tests

For the speed tests, the PLC that controls the pneumatic valves was programmed for a total cycle time of 16 sec/cycle and the tether length was 500 ft. The tests were conducted in the bench scale testbed. The outside of the PVC pipes were marked with gradients 0.5 in a part and the time was recorded for the crawler to navigate a total length of 2 feet. A total of six speed tests are presented. Relocation of the pneumatic valves from the control box to the trailing capsule increased the navigational speed from 1 ft/hr to 3.5 ft/hr. Three tests were conducted with a rubber sleeve of various thickness clamped over the outer bellow to aid in decreasing the cycle time. Using the 0.125 in thick sleeve yielded a speed of 19 ft/hr. The fastest speed recorded of 38 ft/hr was achieved using the thinner bellow configuration. Figure 25 shows a summary of the speed tests conducted using different outer bellow configurations.

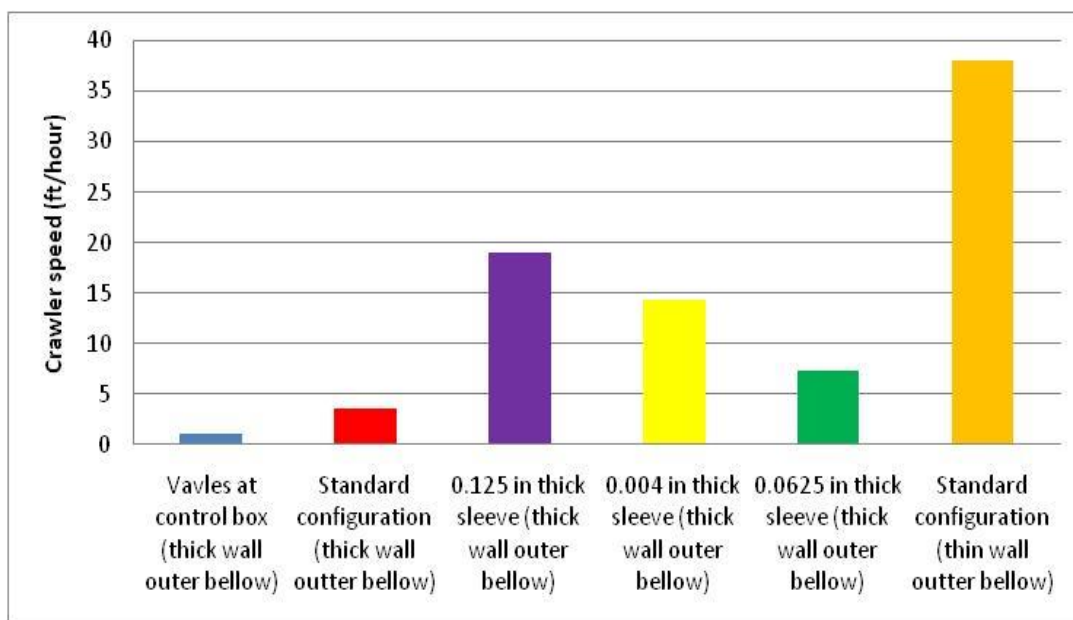


Figure 25. Speed test using different bellow configurations.

Navigational tests were conducted to determine the crawler’s ability to negotiate through a 90° elbow. After manually timing an ideal cycle sequence that provided the shortest time for the crawler to navigate through an elbow, a separate program was loaded into the PLC. The control box was adapted to provide the operator with the capability to alternate between the straight and the turning cycle sequence as needed. Using the turning cycle sequence, the crawler was able to successfully navigate through a 90° elbow in 11 minutes.

Pull force tests were conducted using the thick and thin wall outer bellow configuration. For the thick outer wall below the supply pressure was varied from 0 to 80 psi and generated a maximum force of 166 lb. The crawler assembled using the thinner bellow configuration was provided with a maximum pressure of 50 psi and generated a force of 108 lb. For both cases there was a liner relation between supply pressure and force generated was observed. Figure 26 shows the results of the force test conducted.

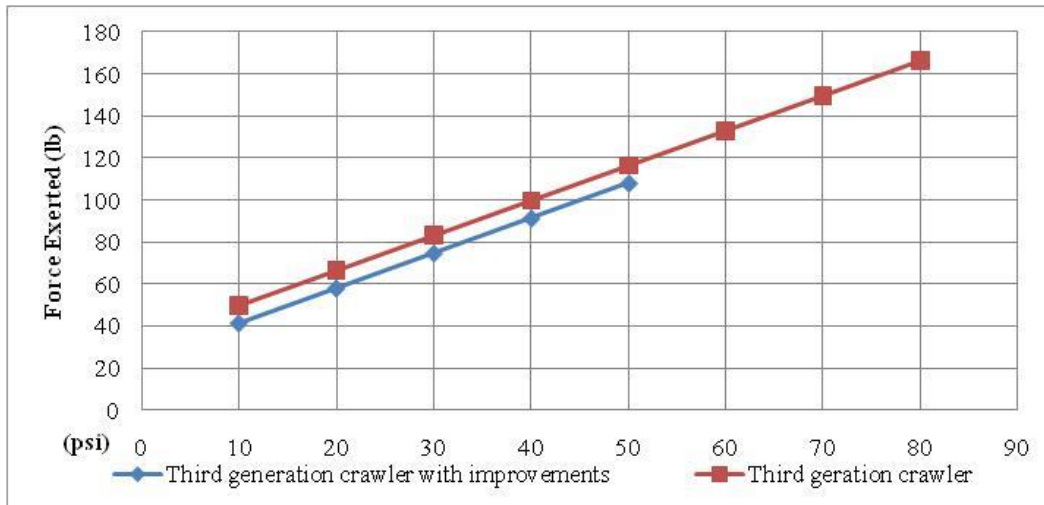


Figure 26. Force tests results.

Engineering scale tests

Using the engineering scale testbed, tests were conducted to determine the pull force required to route the tether through the pipeline. The tests consisted by manually pulling the tether and recording the force using a spring scale. The force recorded to pull the tether in a 21 ft straight pipe section was 18 lb. The force required to pull the tether increased dramatically once the tether was routed through a straight pipe coupled to a 90° elbow. This large force requirement would hinder the crawler unit from successfully navigating longer distances. In order to decrease the friction force and also reduce the contact area between the tether and the pipeline, a 0.051 in stainless steel wire was coiled onto the tether (shown in Figure 27). Pull force tests conducted using this configuration showed that the force required to pull the tether using a 21-ft section was reduced to 4 lb. The force required to pull the tether thru two 42-ft straight sections coupled with a 90° elbow was 28 lb. Figure 28 shows the results of the all manual pull force tests.



Figure 27. Tether with stainless steel coil.

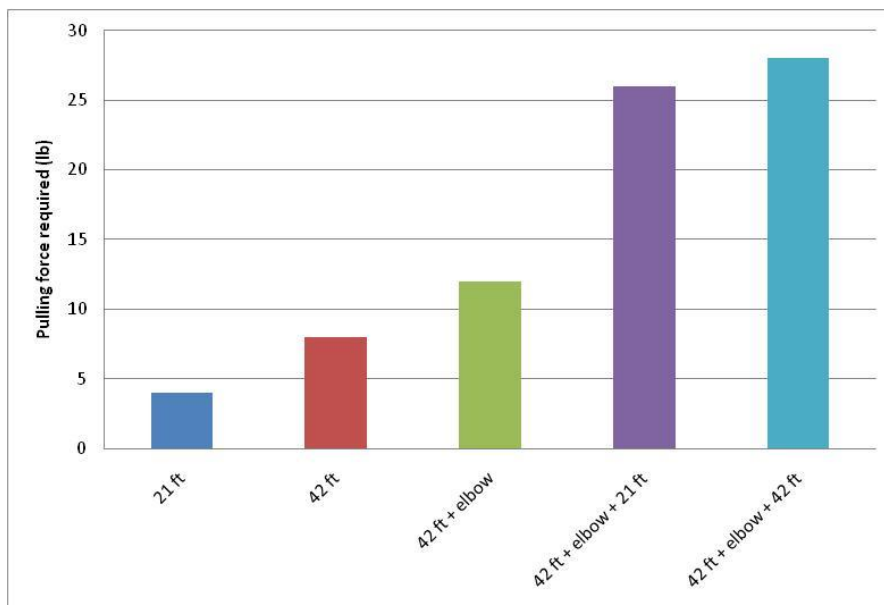


Figure 28. Manual tether pulling force for different pipeline lengths.

Rims fatigue tests

Previous bench scale navigational tests were conducted using PVC pipe sections. The inner diameter of the PVC pipes was 2.9 inches. The carbon steel schedule 10 pipes used in the engineer scale testbed have an inner diameter of 3.26 inches. The increase in the distance between the crawler and the pipeline wall requires the front and back cavities to expand further in order to anchor the crawler unit to the pipeline. During the navigational tests, this additional expansion caused the cavities to rupture after the crawler reached approximately 10 feet from the inlet point of the pipeline. Based on the projected displacement of the crawler per cycle, it is estimated that total of 3,600 cycles would be required for the unit to navigate a 500-ft pipeline. For all tests, the rupture occurred at the stress risers created at the clamps edges. In an attempt to eliminate the failure of the cavities, different materials, configurations, and clamping pressures were tested. The largest number of cycles recorded without failure was 1,260 using a Kevlar gasket placed between the clamps and the flexible cavity. Figure 29 (a) shows the crawler unit’s back rim using a 0.125 in thick material and Figure 29 (b) show a failure due to insufficient clamping force.



Figure 29. (a)Back rim with 0.125 in thick flexible gum, (b)failure due to insufficient clamping force.

An alternative approach to reach the 3,600 cycles without failure is to increase the distance between the clamps. This increment provides additional available material that can expand to reach the inner pipeline wall. An experimental fixture having the same outer diameter as the back rim was assembled and test were performed in a schedule 10 pipeline section. Two tests were performed: one having a distance between clamps of 1.75 in and the other a distance of 1 inch. The test for the 1.75 in provided a total of over 14,000 cycles with no evidence of failure. The test for the 1 inch clamp distance held for a total of over 15,000 cycles prior to failure. Figure 30 (a) shows the experimental fixture and Figure 30 (b) shows the test conducted using the 1 inch clamp distance.



Figure 30. Experimental fixture and experimental fixture inside pipeline section.

Using the experimental result from the fixture, prototypes of the rims having an additional length of 0.5 in to provide 1.25 in of distance between clamps were prototyped and the crawler was then assembled. A pull test in a 90° elbow was conducted by routing a wire through the crawler and attaching it to the rear rim. Tests showed that the increased length of the rims caused the crawler to wedge against the inner surfaces of the elbow preventing it from turning. A second set of rims having 0.25 in of additional length (1 in of distance between clamps) were prototyped and tested using the same procedure. With these rims, the crawler was able to successfully navigate through a 90° elbow with a pull force of 12 lb. Figure 31(a) show the crawler unit with prototype rims and Figure 31(b) shows the crawler unit assembled with a clamp distance of 1.25 inches.



Figure 31. Crawler unit with prototype rims, assemble front end of crawler unit.

CONCLUSIONS AND FUTURE WORK

Two technologies, APS and a peristaltic crawler, have continued to be developed and evaluated during FIU Year 3.

The APS was used on an engineering scale testbed consisting of a symmetric configuration using approximately 135-ft of 3-inch sch-40 threaded pipe with one 90° elbow on each side of a plug. A 3-ft the kaolin-plaster plug was placed in the middle of the testbed and different pulse frequencies were applied to plug. After analyzing the data, several observations were made. As in the previous testing done on smaller loops, air in the system has a major effect on the system's performance. Just as in the tests conducted on smaller loops, the consequence of entrained air can be mitigated by either applying a vacuum to the pipeline or increasing the system's static pressure. The data obtained from this testing indicates that increasing the static pressure is a more effective method. During previous testing, the pulsing frequency played a major role in the system's unplugging capability. However, during these tests on the larger loop this was not observed which is believed to be due to the longer pipeline length damping the frequency.

During this phase of testing, several electronic failures were encountered, including pressure transducers and linear position transducers. This is believed to be due to locating the system outside in the elements. Future work will begin with a full analysis of the failures and determining what changes to the system design need to be done to prevent such failures from happening in the future. In addition, due to the size of the pipeline purging the all the air from the pipes was difficult. This too will be investigated to develop a better air purging method.

The tests conducted showed that the improvements in the design of the PCS have a significant effect on its navigational performance using the bench scale testbed. The relocation of the valves from the control box to the trailing capsule increased the speed of the crawler by 350%. An additional improvement in speed was observed using the thinner wall outer bellow which provided a speed of 38 ft/hr. The addition of the sleeve using the thicker wall outer bellow showed a maximum speed of 19 ft/hr. However, the use of a sleeve has the potential of hampering the crawler's maneuverability through a 90° elbow since it increases the radial stiffness of the crawler as well as creates a higher friction surface with the inner pipeline.

The force tests conducted showed that the thick wall and thin wall bellow have a similar response to force generated due to the supplied air pressure. The tests using the thin wall bellow was not taken above 50 psi to prevent potential permanent deformation. Future tests will include pressurizing the thin wall bellow to failure and fatigue tests to determine the acceptable life limit of the bellow. The use of the 0.051 in stainless steel wire wounded around the tether significantly reduced the pull force required to route the tether through the pipeline. It was observed that an increment of 8 lb was required per every 21-ft section of straight pipe. Routing the tether through an elbow required an increase of 33% in the force. Future test will include routing the tether and measuring the pull force through the remaining three elbows of the engineering scale testbed.

The difference in inner pipe wall diameter between the PVC pipe sections and schedule 10 pipe sections (2.9 in and 3.26 in, respectively) had a significant effect on the life of the rim cavities. Tests showed that a distance between clamps of at least 1 in is required to provide 14,000 cycles with no failure. The use of gaskets to decrease the stress rises at the clamps locations was not sufficient to prevent failure above acceptable limits (3,600 cycles). The prototype having a 0.25

in longer rims provided a clamps distance of 1 in and does not increase the force required for the crawler to navigate through an elbow. Future efforts will include manufacturing a stainless steel crawler with the dimensions of the prototype.

During the next period, final adjustments to the PCS will be implemented and the system will be tested using the total length of the engineering testbed (430 ft). Future tests will include performing unplugging and inspection operations.

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TASK 12: MULTIPLE-RELAXATION-TIME LATTICE BOLTZMANN MODEL FOR MULTIPHASE FLOWS

EXECUTIVE SUMMARY

Many engineering processes at various U.S. Department of Energy sites include the fluid flow of more than one phase such as air and water. Slurry mixing methods such as pulsed-air mixers, air sparging, and pulsed-jet mixing are a few examples where more than one fluid phase can exist in contact with another phase. Lattice Boltzmann method (LBM) is a computational fluid dynamics (CFD) method that can provide insight into the behavior of multiphase flows by capturing the interface dynamics accurately during the process and the effects of structures on multiple fluid phases.

Florida International University (FIU) aims to develop LBM-based computer codes that can be used by the U.S. DOE scientists and engineers as a prediction tool for understanding the physics of fluid flow in nuclear waste tanks during regular operations and retrieval tasks. In FY2009, a new task was initiated within FIU's research project on high-level-waste (HLW) in order to develop a computational code, which is based on the LBM in order to simulate multiphase flow problems related to HLW operations. A thorough literature review was conducted to identify the most suitable multiphase fluid modeling technique in LBM and a single-phase multi-relaxation-time (MRT) code was developed. In FY2010, FIU identified and evaluated a multiphase LBM using a single-relaxation-time (SRT) collision operator and updated the collision process in the computer code with an MRT collision model. In FY2011, the MRT LBM code was extended into three dimensions and the serial computer code was converted into a parallel code.

LBM requires attention when the fluid interface between multiple phases comes in contact with solid surfaces in order to yield the correct molecular forcing exerted on the fluids by the surface. During the current performance period, a proper contact angle model is presented within the LBM framework that allows changing the wetting characteristics of droplets and bubbles on solid surfaces. A procedure to incorporate complex geometries into the LBM simulation is also presented. In addition, a literature review was conducted in order to investigate the applicability of LBM for the simulation of non-Newtonian fluids as it is applicable to the waste mixing using pulse-jet mixers in WTP tanks since the slurry found at Hanford is categorized as Bingham plastic and this feature should be incorporated in the engineering calculations for accurate estimations of the performance of various waste removal and handling operations. Various turbulence models such as the Large Eddy Simulation model were also investigated for the current multiphase LBM code developed at FIU. Most of the engineering problems involving fluid flow in the DOE operations at Hanford, such as the pulse-jet mixing, involve high fluid velocities that include turbulent effects that need to be taken into consideration in any computer model that would be used to understand the behavior of such systems.

INTRODUCTION

The storage and mixing of the liquid waste requires understanding the behavior of the fluid flow inside the waste tanks. Multiple phases of fluids can exist inside the waste tanks where bubble dynamics can play an important role. Gas bubbles can exist in the waste entrapped in the liquid phase, inside cracks within the solid waste or on the surface of the tank walls. They can also be generated inside the waste naturally caused by chemical reactions such as hydrogen production or can be externally supplied via mechanical mechanisms such as air-purging, pulsed-air mixing etc.

An understanding of the physical nature of bubble dynamics inside the waste and the effects of the air release process to the tank environment need to be gained by considering various waste conditions. Such an analysis can be made possible by developing a numerical method that can simulate the process of air bubble generation inside tanks and that can accurately track the interactions of the gas phase with the surrounding fluid and solid phases. The final computational program would serve as a tool for Site engineers and scientists to predict waste behavior and improve operational procedures during the storage, handling and transfer of liquid waste.

In this report, a numerical method based on LBM is presented, which can model multiphase flows accurately and efficiently. LBM is advantageous over traditional Navier-Stokes based computational models since surface forces are handled more easily in LBM. Another capability of LBM is the straightforward integration of complex geometries which eliminates the necessity to generate a complicated computational mesh which usually takes more time than the execution of the simulations in classical CFD. The performance of MRT LBM for multiphase flow simulations in complex three dimensional geometries is presented in this report.

The outline of the report is given as follows: first, the literature review conducted for modeling turbulent flows and non-Newtonian fluid flows using LBM is covered. Then, three-dimensional LBM method is introduced for the simulation of multiphase flows in complex geometries. The results obtained follow and finally, conclusions are drawn and discussions for future work are presented.

Turbulence modeling with LBM

The categorization of viscous flows falls into two types based on the structure of the flow; either laminar flow, which is the case where the slow structure is made of a smooth variation of layers, or turbulent flows where the flow has a random and chaotic structure [1]. The random behavior of turbulent flows presents itself in the occurrence of swirling regions of flow, called eddies, which can vary in size over a wide range. Kolmogorov predicts the size of the smallest eddies as proportional to the Reynolds number, $Re^{-3/4}$ [2]. Turbulent flow simulations are conducted using a fine grid that resolves all aspects of the flow including the smallest eddies are called direct numerical simulation (DNS). DNS is considered to be the most accurate numerical simulation tool; however, the computational load for such simulations becomes tremendous for problems that are characterized with large Reynolds numbers. In the case of a three-dimensional model of a turbulent pulsed-jet flow in a waste tank, the jet Reynolds number can be of the order of 10^4 [3] which requires the grid points to be in the order of 10^9 [4].

As a remedy, the mean values of the macroscopic values of the flow properties are calculated in computational fluid dynamics (CFD) simulations of turbulent flows with the addition of a

subgrid model for small scale details that cannot be captured with the coarse grid. The turbulence models are needed to provide an expression for the Reynolds stress tensor that appears in the Navier-Stokes equations as a result of the averaging process, which is also referred as the “filtering”. Large eddy simulations use this approach to calculate an effective viscosity to predict the effect of small eddies using coarse grid values. One popular approach is provided by Smagorinsky [5], where the Reynolds stress tensor is calculated using the local strain rate without the requirement of solving a separated partial differential equation.

An additional approach is to use the Reynolds Averaged Navier-Stokes (RANS) equations in conjunction with a turbulence model such as the k - ϵ model to model the turbulent scales. The Reynolds stress is represented in the k - ϵ model in terms of the strain rate, the turbulent kinetic energy k and the dissipation rate ϵ . The addition of the two unknown variables, k and ϵ , requires the solution of two additional transport equations. This approach yields time-averaged velocities and is found to be less accurate than the LES; however, fine mesh requirement of LES requires more computer power compared to RANS-based turbulence modeling. In summary, LES can be considered as a compromise between DNS and RANS in terms of computational resource requirements and accuracy.

The above discussion for various approaches to model turbulence using the Navier-Stokes equation is also applicable to the lattice Boltzmann equation. Martinez et al. compared LBM against pseudospectral method for decaying turbulence of a shear layer for an initial Reynolds number of 10,000 [6]. Other two-dimensional applications include isotropic forced turbulence [7; 8] and bifurcation of Poiseuille flow [9]. Three dimensional DNS of isotropic turbulence using LBM was shown in [10; 11] as well as in [12] for anisotropic shear flow and in [13] for channel flow. However, these applications were confined to simple turbulent flows since the DNS using LBM for high Re becomes impractical [14]. The Smagorinsky’s subgrid approach has been applied in various works. Hou et al. [25] used it to simulate two-dimensional cavity flow at $Re=106$ where three-dimensional turbulent pipe flow at $Re=50,000$ was simulated also in [16]. Eggels used LES-LBM approach to model the turbulent flow in a baffled stirred tank reactor with a $Re=1.07 \times 10^5$ [17]. The same problem has been studied by Derksen and Van den Akker in which the fine grid resolution requirement in the work of Eggels were removed using an interpolation scheme [18; 19]. Krafczyk showed that the LBM using a BGK collision model was able to obtain accurate solution of $Re = 40,000$ flow around a cube in a channel flow when combined with the Smagorinsky model [20] and later extended it to MRT-LBE model to improve the numerical stability [21]. Teixeira has incorporated algebraic and two-equation turbulence models in the LBE framework and compared the performance for turbulent pipe flow and a backward facing step problem [22]. Solving the k - ϵ equations within the LBE was also proposed by Succi et al. [23]; however, the capability to obtain the strain-rate tensor directly from the LBE computations in order to determine the subgrid stress in the Smagorinsky model and the higher accuracy of LES compared to other turbulence models make the LES-LBE approach more common. The LES-LBE can still be considered to be at an early stage of maturity and the literature is lacking any applications of LBM to multiphase flows at turbulence flow conditions.

Non-Newtonian fluid flow modeling with LBM

The rheology of the slurries found at Hanford Site show Non-Newtonian behavior which is classified as the Bingham plastic fluids [24]. The Bingham plastic acts as a solid until a threshold or yield stress is achieved, after which a linear relation between the stress and strain is observed.

Bingham yield stress has been reported to affect the performance of jet mixing [25] and gas bubble retention [26]. Therefore, the simulation capability using LBM should include the effects of Bingham plastic fluids in the modeling of fluid flow phenomena in DST and WTP tanks.

The simulation of power-law fluids using the LBM was shown in the work of Aharonov and Rothman for non-Newtonian flows in porous media by modifying the relaxation time parameter in order to relate the viscosity to the stress tensor [27]. Gabbanelli et al. also improved this approach to shear thickening and shear thinning fluids by proposing a truncated power-law model [28]. Boyd et al. further improved this approach by proposing a second-order accurate calculation of the shear-rate [29], which was later used by Tang et al. [30] to simulate electric field effects on non-Newtonian fluids using LBM. Power-fluid simulations were also conducted in the study of Yoshino et al. [31] with their lattice kinetics approach where the relaxation parameter is fixed to one for stability purposes and the effect of the shear-rate on the viscosity is achieved via a constant parameter added to the equilibrium distribution function.

There have been a few models proposed in regards to the LB applications to Bingham plastics. Wang and Ho used the expression for the shear stress proposed by Papanastasiou [32] which avoids the discontinuity in the definition of the shear stress, which occurs for Bingham plastics depending on the local stress value [33]. By using a large exponential growth factor in the definition of the shear stress, Wang and Ho was able to approximate the shear stress relation to the shear rate for Bingham plastics. They also derived a new equilibrium distribution function that incorporates the local shear-rate effect. Recently Tang et al. applied the same theory to the pressure-based He-Luo incompressible LBM to remove the compressibility effects in the simulations [34]. Chai et al. has also developed the application of LBM for modeling generalized Newtonian fluid flows including the Bingham plastics, using the multiple-relaxation-time (MRT) collision model to improve the stability of the method for low viscosities [35].

Numerical method

The lattice Boltzmann method developed for this task is based on the continuous Boltzmann equation given by

$$\frac{\partial f}{\partial t} + \boldsymbol{\xi} \cdot \nabla f + \mathbf{F} \cdot \nabla_{\boldsymbol{\xi}} f = \Omega. \quad (1)$$

Here f is the single particle density distribution function, $\boldsymbol{\xi}$ is the particle velocity, \mathbf{F} is the combination of interfacial force and external body forces such as gravity and wall adhesion force and Ω is the collision term. The term $\nabla_{\boldsymbol{\xi}} f$ can be approximated as,

$$\nabla_{\boldsymbol{\xi}} f \approx \nabla_{\boldsymbol{\xi}} f^{eq} = -\frac{\boldsymbol{\xi} \cdot \mathbf{u}}{\rho c_s^2} f^{eq}, \quad (2)$$

where f^{eq} is the equilibrium density distribution function, \mathbf{u} is the macroscopic velocity, ρ is the density and c_s is the speed of sound. The continuous Boltzmann equation given in Eq. (1) can be discretized in the velocity space by expressing as

$$\frac{\partial f_{\alpha}}{\partial t} + \boldsymbol{\xi}_{\alpha} \cdot \nabla f_{\alpha} = \Omega + S_{f_{\alpha}}, \quad (3)$$

where

$$e_\alpha \equiv \xi_\alpha = \begin{cases} (0,0,0) & , \quad \alpha = 0, \\ (\pm 1, 0, 0), (0, \pm 1, 0), (0, 0, \pm 1), & , \quad \alpha = 1, 2, \dots, 6 \\ (\pm 1, \pm 1, 0), (\pm 1, 0, \pm 1), (0, \pm 1, \pm 1) & , \quad \alpha = 7, 8, \dots, 18 \end{cases} \quad (4)$$

and

$$S_{f\alpha} = \frac{(e_{\alpha i} - u_i)(F_i + F_i^W + G_i)}{\rho c_s^2} f_\alpha^{eq}. \quad (5)$$

In Eq. (4) α is the discrete particle velocity distribution using the D3Q19 lattice structure shown in Figure 32, e is the particle velocity between lattice points.

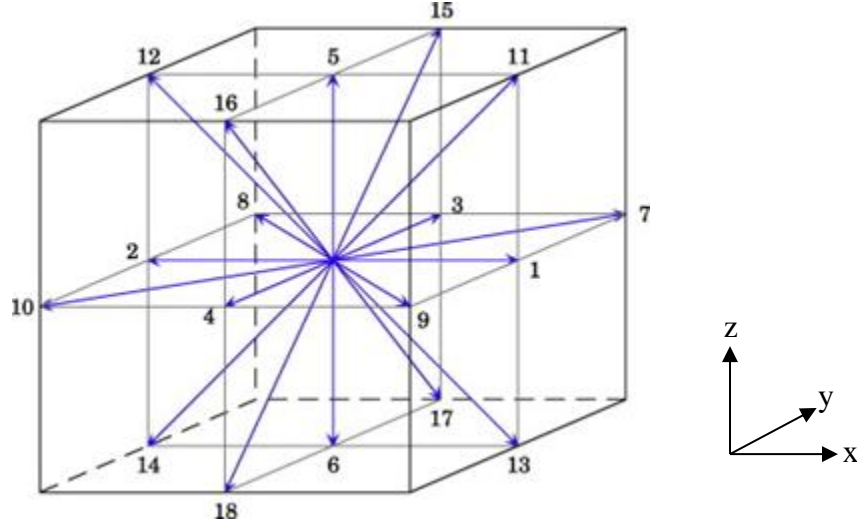


Figure 32. D3Q19 lattice structure.

In the single-relaxation-time LBM, the collision term Ω is represented using the BGK model that uses a single relaxation time parameter (λ), $\Omega_c = -\left(\frac{f_\alpha - f_\alpha^{eq}}{\lambda}\right)$. In the MRT LBM, using a collision matrix A , the collision term on the right hand side of Eq. (3) is represented by

$$\Omega = -A_{\alpha\beta} (f_\beta - f_\beta^{eq}). \quad (6)$$

The equilibrium distribution function, f_α^{eq} , is written as

$$f_\beta^{eq} = w_\alpha \rho \left[1 + \frac{e_{\alpha i} u_i}{c_s^2} + \frac{(e_{\alpha i} e_{\alpha j} - c_s^2 \delta_{ij}) u_i u_j}{2c_s^4} \right], \quad (7)$$

where w_α is the weight function given by

$$w_\alpha = \begin{cases} 1/3, & \alpha = 0, \\ 1/18, & \alpha = 1, 2, \dots, 6, \\ 1/36, & \alpha = 7, 8, \dots, 18. \end{cases} \quad (8)$$

The force F_i in Eq. (5) is responsible for phase separation and is given by

$$F_i = \frac{\partial}{\partial x_j} (\rho c_s^2 - P) \delta_{ij} + \kappa \rho \frac{\partial}{\partial x_i} \frac{\partial^2 \rho}{\partial x_j \partial x_j}. \quad (9)$$

Here, P is the pressure and κ is the surface tension parameter which is related to the surface tension σ through the relation

$$\sigma = \kappa \int \left(\frac{\partial \rho}{\partial r} \right)^2 dr, \quad (10)$$

where r is the direction of integration normal to the interface. The force F_i is expressed by Lee and Lin [36] as

$$F_i = \frac{\partial \rho c_s^2}{\partial x_j} \delta_{ij} - \rho \frac{\partial}{\partial x_i} \left(\frac{\partial E_f}{\partial \rho} - \kappa \frac{\partial^2 \rho}{\partial x_j \partial x_j} \right). \quad (11)$$

E_f is the excess free energy at the interface over the bulk free energies and is obtained from an equation of state (EOS) expressed as follows:

$$E_f(\rho) \approx \beta (\rho - \rho_g^{sat})^2 (\rho - \rho_l^{sat})^2, \quad (12)$$

where β is a constant and ρ_g^{sat} and ρ_l^{sat} are densities of gas and liquid phases at saturation, respectively. This EOS results in a density profile given by

$$\rho(z) = \frac{\rho_l^{sat} + \rho_g^{sat}}{2} - \frac{\rho_l^{sat} - \rho_g^{sat}}{2} \tanh\left(\frac{2z}{D}\right), \quad (13)$$

where z is the spatial location normal to the interface and D is the interface thickness.

The constant β along with κ can control D and σ through the relation

$$D = \frac{4}{(\rho_l^{sat} - \rho_g^{sat})} \sqrt{\frac{\kappa}{2\beta}}, \quad (14)$$

and

$$\sigma = \frac{(\rho_l^{sat} - \rho_g^{sat})^3}{6} \sqrt{2\kappa\beta}. \quad (15)$$

The forcing term $F_i^w = -\rho K_w \mathbf{n}$ that also appears in Eq. (5) is introduced to represent the effect of surface molecules where K_w is the parameter that determines whether the solid surfaces act attractive or repulsive. In addition, the gravitational force is implemented in Eq. (5) via the $G_i = \rho \mathbf{g}$ force term where \mathbf{g} stands for the gravitational acceleration vector.

The evolution equations given above for the particle density distribution function is mapped into the moment space by multiplying the terms in Eq. (3) with the transformation matrix \mathbf{T}

$$\mathbf{T}^T = \begin{bmatrix} \langle \rho |, \langle e |, \langle e^2 |, \langle j_x |, \langle q_x |, \langle j_y |, \langle q_y |, \langle j_z |, \langle q_z | \\ \langle 3p_{xx} |, \langle 3\pi_{xx} |, \langle p_{ww} |, \langle \pi_{ww} |, \langle p_{xy} |, \langle p_{yz} |, \langle p_{xz} |, \langle m_x |, \langle m_y |, \langle m_z | \end{bmatrix}, \quad (16)$$

where

$$|\rho\rangle = |e_\alpha|^0, \quad (17)$$

$$|e\rangle_\alpha = 19|e_\alpha|^0 - 30, \quad (18)$$

$$|e^2\rangle_\alpha = (421|e_\alpha|^{04} - 53|e_\alpha|^2 + 24)/2 \quad (19)$$

$$|j_x\rangle_\alpha = e_{\alpha,x}, \quad (20)$$

$$|q_x\rangle_\alpha = [5|e_\alpha|^2 - 9]e_{\alpha,x}, \quad (21)$$

$$|j_y\rangle_\alpha = e_{\alpha,y}, \quad (22)$$

$$|q_y\rangle_\alpha = [5|e_\alpha|^2 - 9]e_{\alpha,y}, \quad (23)$$

$$|j_z\rangle_\alpha = e_{\alpha,z}, \quad (24)$$

$$|q_z\rangle_\alpha = [5|e_\alpha|^2 - 9]e_{\alpha,z}, \quad (25)$$

$$|3p_{xx}\rangle_\alpha = 3e_{\alpha,x}^2 - |e_\alpha|^2, \quad (26)$$

$$|3\pi_{xx}\rangle_\alpha = (3|e_\alpha|^2 - 5)(3e_{\alpha,x}^2 - |e_\alpha|^2), \quad (27)$$

$$|p_{ww}\rangle_\alpha = e_{\alpha,y}^2 - e_{\alpha,z}^2, \quad (28)$$

$$|\pi_{ww}\rangle_\alpha = (3|e_\alpha|^2 - 5)(e_{\alpha,y}^2 - e_{\alpha,z}^2), \quad (29)$$

$$|p_{xy}\rangle_\alpha = e_{\alpha,x}e_{\alpha,y}, \quad (30)$$

$$|p_{yz}\rangle_\alpha = e_{\alpha,y}e_{\alpha,z}, \quad (31)$$

$$|p_{xz}\rangle_\alpha = e_{\alpha,x}e_{\alpha,z}, \quad (32)$$

$$|m_x\rangle_\alpha = (e_{\alpha,y}^2 - e_{\alpha,z}^2)e_{\alpha,x}, \quad (33)$$

$$|m_y\rangle_\alpha = (e_{\alpha,z}^2 - e_{\alpha,x}^2)e_{\alpha,y}, \quad (34)$$

$$|m_z\rangle_\alpha = (e_{\alpha,x}^2 - e_{\alpha,y}^2)e_{\alpha,z}. \quad (35)$$

The resulting evolution equation in moment space takes the form

$$\frac{\partial \hat{f}_\alpha}{\partial t} + \hat{\xi}_\alpha \cdot \nabla \hat{f}_\alpha = -\hat{\Lambda}_{\alpha\beta} (\hat{f}_\beta - \hat{f}_\beta^{eq}) + \hat{S}_{f\alpha}, \quad (36)$$

where

$$\hat{f}_\alpha = \mathbf{T}f_\alpha, \quad (37)$$

$$\hat{f}_\alpha^{eq} = \mathbf{T}f_\alpha^{eq}, \quad (38)$$

$$\hat{S}_{f\alpha} = \mathbf{T}S_{f\alpha}, \quad (39)$$

and

$$\hat{\Lambda} = \mathbf{T}\Lambda\mathbf{T}^{-1}. \quad (40)$$

The equilibrium distribution function in moment space is written as

$$(\hat{f}^{eq})^T = \left[\rho, e^{eq}, (e^2)^{eq}, j_x, q_x^{eq}, j_y, q_y^{eq}, j_z, q_z^{eq}, 3p_{xx}^{eq}, 3\pi_{xx}^{eq}, p_{ww}^{eq}, \pi_{ww}^{eq}, p_{xy}^{eq}, p_{yz}^{eq}, p_{xz}^{eq}, m_x^{eq}, m_y^{eq}, m_z^{eq} \right], \quad (41)$$

where the equilibrium distributions of the moments are given by

$$e^{eq} = -11\rho + 19(j_x^2 + j_y^2 + j_z^2)/\rho, \quad (42)$$

$$(e^2)^{eq} = w_\varepsilon\rho + w_{\varepsilon j}(j_x^2 + j_y^2 + j_z^2)/\rho, \quad (43)$$

$$q_x^{eq} = -\frac{2}{3}j_x, \quad (44)$$

$$q_y^{eq} = -\frac{2}{3}j_y, \quad (45)$$

$$q_z^{eq} = -\frac{2}{3}j_z, \quad (46)$$

$$p_{xx}^{eq} = \frac{1}{3} \left(2j_x^2 - (j_y^2 + j_z^2) \right) / \rho, \quad (47)$$

$$\pi_{xx}^{eq} = w_{xx} p_{xx}^{eq}, \quad (48)$$

$$p_{ww}^{eq} = (j_y^2 - j_z^2) / \rho, \quad (49)$$

$$p_{xy}^{eq} = (j_x j_y) / \rho. \quad (50)$$

$$p_{yz}^{eq} = (j_y j_z) / \rho. \quad (51)$$

$$p_{xz}^{eq} = (j_x j_z) / \rho. \quad (52)$$

$$\pi_{ww}^{eq} = w_{xx} p_{ww}^{eq}, \quad (53)$$

$$m_x^{eq} = 0, \quad (54)$$

$$m_y^{eq} = 0, \quad (55)$$

$$m_z^{eq} = 0. \quad (56)$$

In the works of d'Humieres et al. [37], the constants in Eqs. (42-56) are selected as $w_\varepsilon = 0$, $w_{ej} = -475/63$ and $w_{xx} = 0$ for single-phase flows; however, in this work we follow the selection of Premnath and Abraham [38] used in their MRT LBM for multiphase flows where $w_\varepsilon = 3$, $w_{ej} = -11/2$ and $w_{xx} = -1/2$.

The collision matrix in the moment space, $\hat{\Lambda}$, is given as

$$\hat{\Lambda} = \text{diag}[s_1, s_2, s_3, s_4, s_5, s_6, s_7, s_8, s_9, s_{10}, s_{11}, s_{12}, s_{13}, s_{14}, s_{15}, s_{16}, s_{17}, s_{18}]. \quad (57)$$

The diagonal elements are inverses of relaxation times for the distribution functions in the moment space, \hat{f}_α , and they are used to relax the equilibrium distribution functions in the moment space, \hat{f}_α^{eq} . In this work, the diagonal elements are selected as a combination of the values reported by d'Humieres [37] and Premnath and Abraham [38] as

$$\hat{\Lambda} = \text{diag} \left[1, 1.19, 1.4, 0, 1.2, 0, 1.2, 0, 1.2, \frac{1}{\tau}, 1.4, \frac{1}{\tau}, 1.4, \frac{1}{\tau}, \frac{1}{\tau}, \frac{1}{\tau}, 1.98, 1.98, 1.98 \right]. \quad (58)$$

The parameters s_{10} , s_{12} and s_{14-16} are related to the single relaxation time, τ , in the single-relaxation-time LBM and are used to determine the viscosity, $\nu = \frac{1}{3} \left(\frac{1}{s_{10}} \right)$ and the Reynolds number, $Re = UD/\nu$.

As discussed in Lee and Lin [36], it is possible to compute the hydrodynamic variables of interest such as local density ρ , velocity \mathbf{u} , and pressure P , from f_α . This approach, however, is prone to numerical instabilities due to the steep density gradients involved in the computation of the source term $S_{\alpha\beta}$. Therefore, a separate distribution function g is introduced to compute pressure and momentum. We denote g in the lattice velocity direction α as g_α . A pressure function p is also defined, which varies smoothly across the interface. It is related to the actual pressure P through

$$p = P - \kappa\rho \frac{\partial^2 \rho}{\partial x_k \partial x_k} + \frac{\kappa}{2} \frac{\partial \rho}{\partial x_k} \frac{\partial \rho}{\partial x_k}. \quad (59)$$

In the bulk phase, $p \cong P$ as density gradients are nearly zero. Use of p in the momentum equation increases the stability of the scheme at high density ratios. This definition of P and the choice of D in Eq. (13) are critical to the capability of the model to simulate high density ratios. Based on the definition of p in Eq. (39), Eq. (9) for F_i may be re-arranged as

$$F_i = \frac{\partial}{\partial x_j} (\rho c_s^2 - p) \delta_{ij} + \kappa \frac{\partial}{\partial x_j} \left(\frac{\partial \rho}{\partial x_k} \frac{\partial \rho}{\partial x_k} \delta_{ij} - \frac{\partial \rho}{\partial x_i} \frac{\partial \rho}{\partial x_j} \right). \quad (60)$$

However, within this framework of the MRT model, the pressure evolution equation must now be formulated to have a non-diagonal collision matrix.

To develop an evolution equation for pressure in moment space that is similar to Eq. (26), Lee and Lin proposed

$$\hat{g}_\alpha = \hat{f}_\alpha + \left(\frac{p}{c_s^2} - \rho \right) \hat{\Gamma}_\alpha(0), \quad (61)$$

where

$$\hat{g}_\alpha = \mathbb{T} g_\alpha, \quad (62)$$

$$\Gamma_\alpha(u) = \frac{f_\alpha^{eq}}{\rho}, \quad (63)$$

and

$$\hat{\Gamma}_\alpha(0) = \mathbb{T} \hat{\Gamma}_\alpha(0). \quad (64)$$

From Eq. (41), the total derivative of \hat{g}_α can be written as

$$\frac{D \hat{g}_\alpha}{Dt} = \frac{D \hat{f}_\alpha}{Dt} + \frac{1}{c_s^2} \frac{D p}{Dt} \hat{\Gamma}_\alpha(0) - \frac{D \rho}{Dt} \hat{\Gamma}_\alpha(0), \quad (65)$$

which can be simplified to

$$\frac{D \hat{g}_\alpha}{Dt} = \hat{\Omega}_c + \hat{S}_{g\alpha} \quad (66)$$

where

$$\hat{S}_{g\alpha} = \mathbb{T} S_{g\alpha}, \quad (67)$$

$$\begin{aligned} S_{g\alpha} &= \frac{(e_{\alpha i} - u_i) \partial_i (\rho c_s^2 - p)}{c_s^2} (\Gamma_\alpha(u) - \Gamma_\alpha(0)) \\ &+ \frac{(e_{\alpha i} - u_i) [\kappa \partial_i (\partial_k \rho \partial_k \rho) - \kappa \partial_j (\partial_i \rho \partial_j \rho) + \rho \mathbf{g} - \rho K_w \mathbf{n}]}{c_s^2} \Gamma_\alpha(u), \end{aligned} \quad (68)$$

and

$$\widehat{\Omega}_c = -\widehat{\Lambda}_{\alpha\beta} (\widehat{f}_\beta - \widehat{f}_\beta^{eq}). \quad (69)$$

To express Eq. (49) as a function of \widehat{g} we define

$$\widehat{g}_\alpha^{eq} = \widehat{f}_\alpha^{eq} + \left(\frac{\rho}{c_s^2} - \rho \right) \widehat{\Gamma}_\alpha(0). \quad (70)$$

From this we get the following pressure evolution equation in moment space as

$$\frac{D\widehat{g}_\alpha}{Dt} = -\widehat{\Lambda}_{\alpha\beta} (\widehat{g}_\beta - \widehat{g}_\beta^{eq}) + \widehat{S}_{g\alpha}, \quad (71)$$

The macroscopic properties of density, momentum and pressure are obtained from the following relations:

$$\rho = \sum_{\alpha} f_{\alpha}, \quad (72)$$

$$\begin{aligned} \rho u_i = \sum_{\alpha} e_{\alpha} g_{\alpha} + \kappa \frac{dt}{2} \left[\frac{\partial}{\partial x_i} \left(\frac{\partial \rho}{\partial x_k} \frac{\partial \rho}{\partial x_k} \right) - \frac{\partial}{\partial x_j} \left(\frac{\partial \rho}{\partial x_i} \frac{\partial \rho}{\partial x_j} \right) \right] \\ + \frac{dt}{2} \rho c_s^2 (\mathbf{g} - K_w \mathbf{n}) \end{aligned} \quad (73)$$

$$p = c_s^2 \sum_{\alpha} g_{\alpha} + u_i \frac{\partial \rho c_s^2}{\partial x_i} \frac{dt}{2} \quad (74)$$

RESULTS

In the first numerical test case presented here, the wall adhesion boundary condition implemented in the LBM code was evaluated for accuracy. The force on the solid surfaces (F_i^w) was made to repel or attract the liquid phases depending on the sign of the K_w parameter. In order to test the performance of the implementation of this forcing term, a numerical problem was generated in the computer where a semi-circular droplet was placed on a flat solid surface in a computational domain filled with a lighter medium. Under a zero gravitational field, the initial droplet was left to stabilize. Depending on the value and sign of the K_w parameter, the initial contact angle of the droplet (90°) was expected to converge to a value larger or smaller than 90° . Figure 33(a) shows that for a value of $K_w = 0.001$, the contact angle of the droplet is 77.5° while Figure 33(b) shows that for $K_w = -0.001$ the contact angle obtained becomes 102.5° . The initial results indicated that the forcing term implemented in the LBM code could simulate the effects of hydrophobic and hydrophilic surfaces.

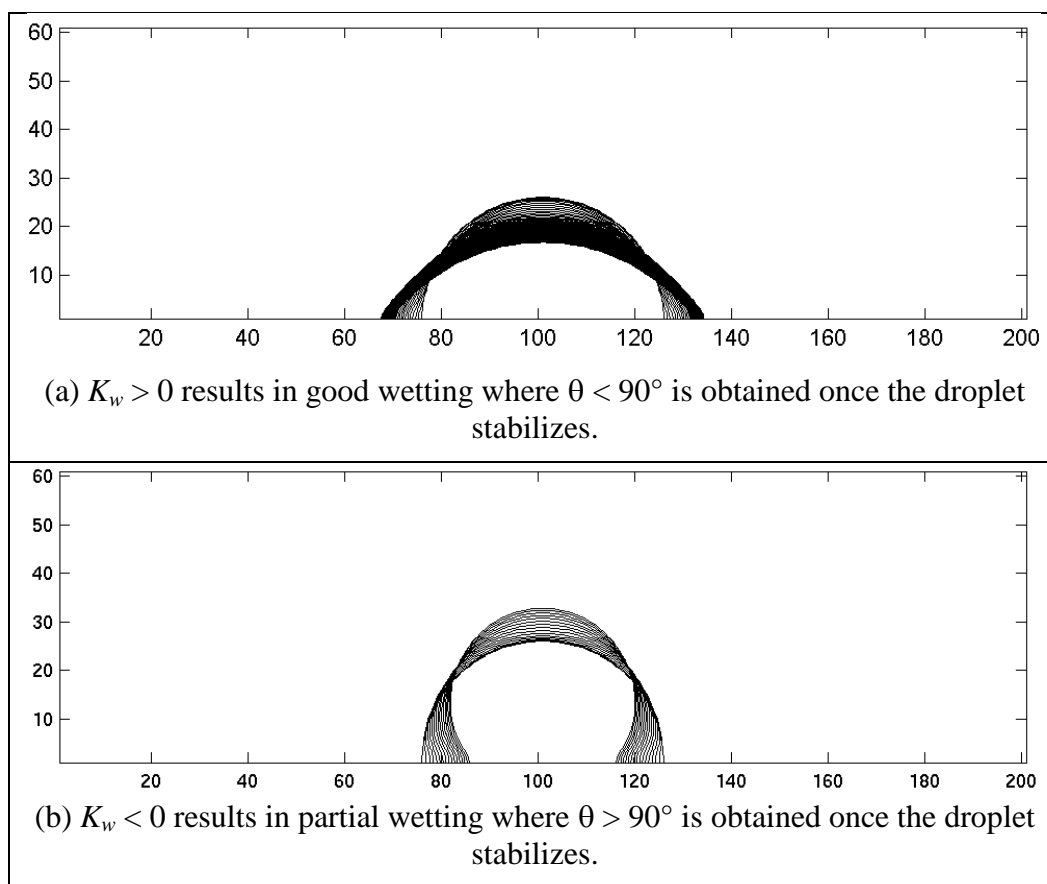


Figure 33. Evolution of droplet on a flat surface under hydrophilic or hydrophobic conditions.

Next, the 3D LBM code was utilized to investigate the motion of buoyant bubbles and droplets in various configurations where the geometry was simple. In these cases the fluids to be simulated was placed in a box where the sides were specified to be either a periodic boundary or a solid surface. No inlet and outlet flow boundaries were used. The cases studies included

collision of a droplet with a solid surface and on a wet surface. The droplet shown in Figure 34 was released at a certain height close to the bottom wall with zero initial velocity. The droplet was observed to change shape during the collision process and then attain an ellipsoidal shape once stabilized. In the case of a wet surface, the droplet was observed to mix into the liquid layer covering the bottom wall surface. We also tested the code in handling a large number of bubbles in motion in an open domain under the effect of buoyancy. The first image in Figure 36 shows the initial configuration of bubbles of equal size placed at zero initial velocity in a liquid box. The second and third images are obtained as the bubbles rise due to the buoyancy effect. It was observed that the bubbles at the middle region of the domain rose at a slower speed and were able to maintain their shapes as compared to the outer bubbles which had a faster rising speed and collided with surrounding bubbles and formed larger ones.

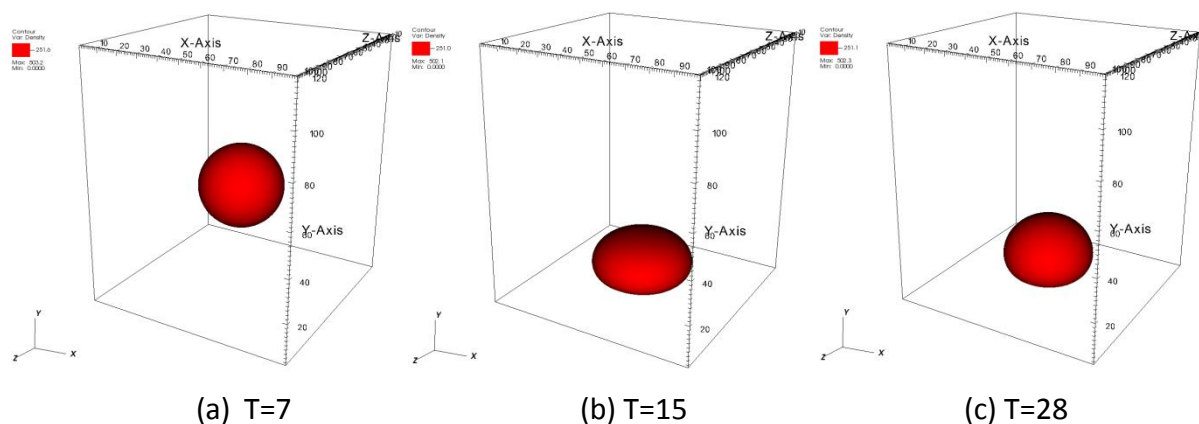


Figure 34. Collision of a droplet on a solid surface.

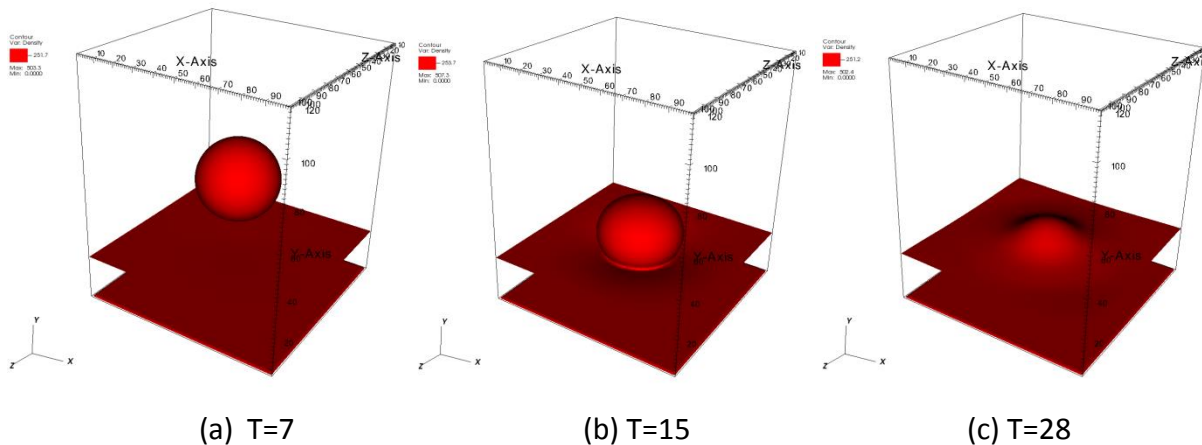


Figure 35. Collision of a droplet on a wet surface.

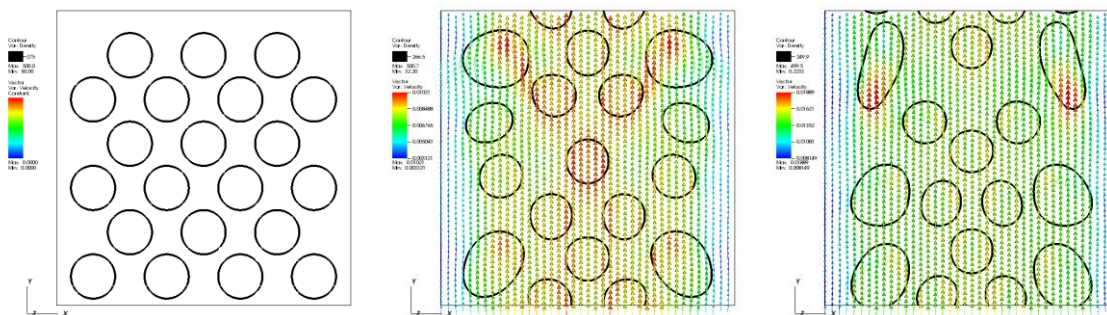


Figure 36. Evolution of an array of bubbles rising in an unconfined geometry.

Geometry importing to the LBM

The simulations so far have used simple geometries with rectangular plains in 2D and boxes in 3D. In real-life applications such as the bubble generation in double-shell tanks or mixing of HLW in tanks with cooling coils, the solid structures such as pipes and tank walls etc. play an important factor in the behavior of the multiphase flow in these environments. Therefore, during this fiscal year, FIU investigated options to represent complex geometries in the LBM framework accurately and efficiently. The ultimate goal was to import an output file from a computer-aided drawing (CAD) software such as Solidworks into the LBM code. This requires a process called ‘voxelation’ where the three-dimensional volume-elements, voxels, are used to represent the geometry on the structured grid of LBM similar to the bitmap representation of two-dimensional image data. The fluid and solid nodes take either the value of zero or one to be distinguished in the LBM code once the voxelation process is finished.

FIU worked with Oak Ridge National Laboratory (ORNL) related to this topic where two students have studied the CartGen procedure being developed at ORNL for the purpose of representing the complex nuclear reactor parts in the Pratham LBM code ORNL is developing for single-phase turbulent flow simulations. The CartGen method is currently under development and has given some issues when the geometry to be voxelated had too many corners. Therefore FIU has chosen to use the Binvox algorithm developed at Princeton University (<http://www.cs.princeton.edu/~min/binvox/>). Using this simple code, FIU were able to convert STL files obtained from Solidworks software into a RAW format which the LBM code were able to integrate. A separate subroutine in the 3D LBM code has been written for this purpose. A few examples of this procedure is shown in the following figure.

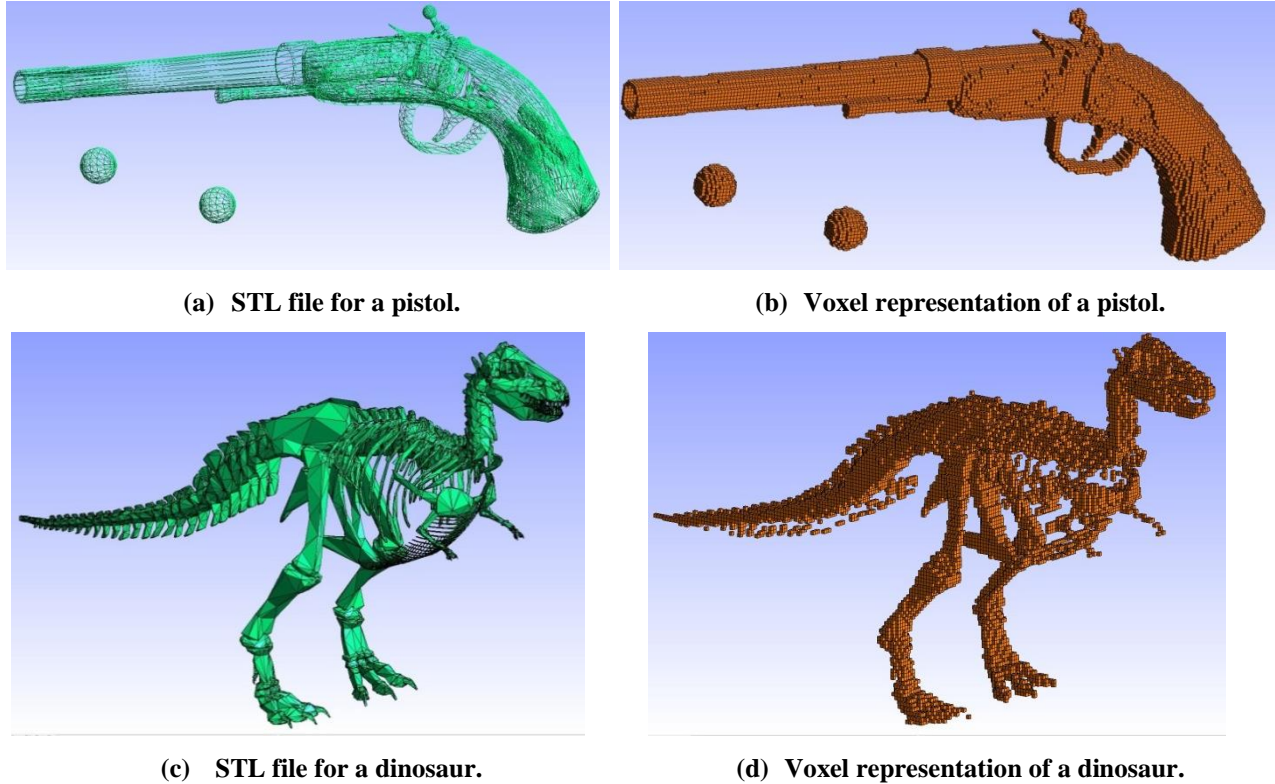


Figure 37. Results of Binvox voxelation procedure for converting STL geometry files into RAW format.

Boundary normals

The implementation of the wall adhesion force on solid surfaces with complex geometries requires the calculation of the surface normal vector, \mathbf{n} at the solid lattice boundary nodes. LBM simulations typically work with a discrete, uniform grid of cubic cells. There exist boundary facets between solid and fluid nodes. Each facet is perpendicular to the lines between node centers. For each facet of a 3-D model, there are six possible boundary normals:

$$\pm e_x = (\pm 1, 0, 0), \pm e_y = (0, \pm 1, 0), \pm e_z = (0, 0, \pm 1)$$

One method of calculating \mathbf{n} for each node consists of averaging the individual boundary facet normal vectors within a specified spherical radius from a given node center. The procedure is expressed as an arithmetic mean:

$$\mathbf{n} = n_b^{-1} \sum_{i=0}^{n_b-1} \mathbf{b}_i, \quad (36)$$

Here \mathbf{b}_i is the boundary normal of the i -th facet and n_b is the number of facets that exist within the specified radius r .

In certain instances, if r is either too large or too small, the averaging of normal may be unacceptable or not representative. If r is too large, it may encompass nodes from separate parts of the object's surface or from a separate object. If r is too small, the staircase approximation of the object surface may influence \mathbf{n} wildly from node to node.

In order to alleviate these issues, a weighted mean may be employed. Eq. (33) then becomes:

$$\mathbf{n} = (\sum_i w_i)^{-1} \sum_{i=0}^{n_b-1} w_i \mathbf{b}_i, \quad (37)$$

The weighting term is defined as: $w_i = 1/(1 + d_i)^\gamma$, where d_i is the distance between nodes (i.e. the center node and the node with normal \mathbf{b}_i). γ determines the strength of the weighting. For $\gamma = 0$ there is equal weighting for all nodes, thus reducing back to eq. (33). This algorithm was used to calculate surface normals over a cylinder. The results can be seen below.

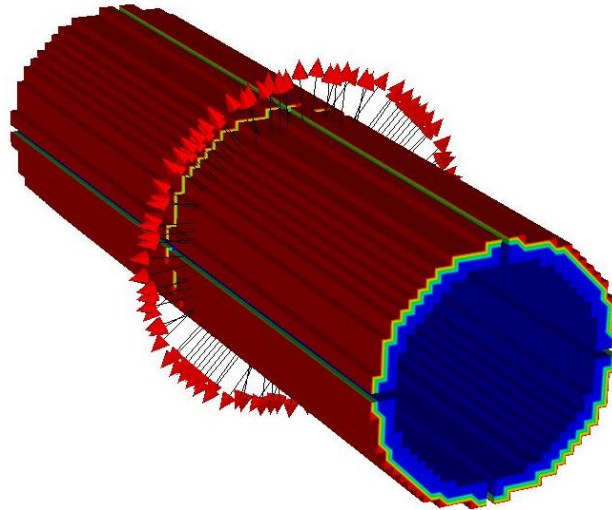


Figure 38. 3D Normal vectors for a cylinder.

Simulations with complex geometries

In the following figures, the LBM simulations that included the voxelation process for geometry import is shown. This includes the motion of a droplet in a constricted channel under inertial flow and buoyant gas release event in a simplified waste tank. The cases studied allowed us to observe the effect of structures on the two phases. In the case of the constricted channel problem presented in Figure 39, the circular droplet retained an elongated shape after it passed through the constriction. The droplet slows down at the constriction and the time when the droplet leaves the domain at the exit was affected by the wall surface adhesion coefficient K_w .

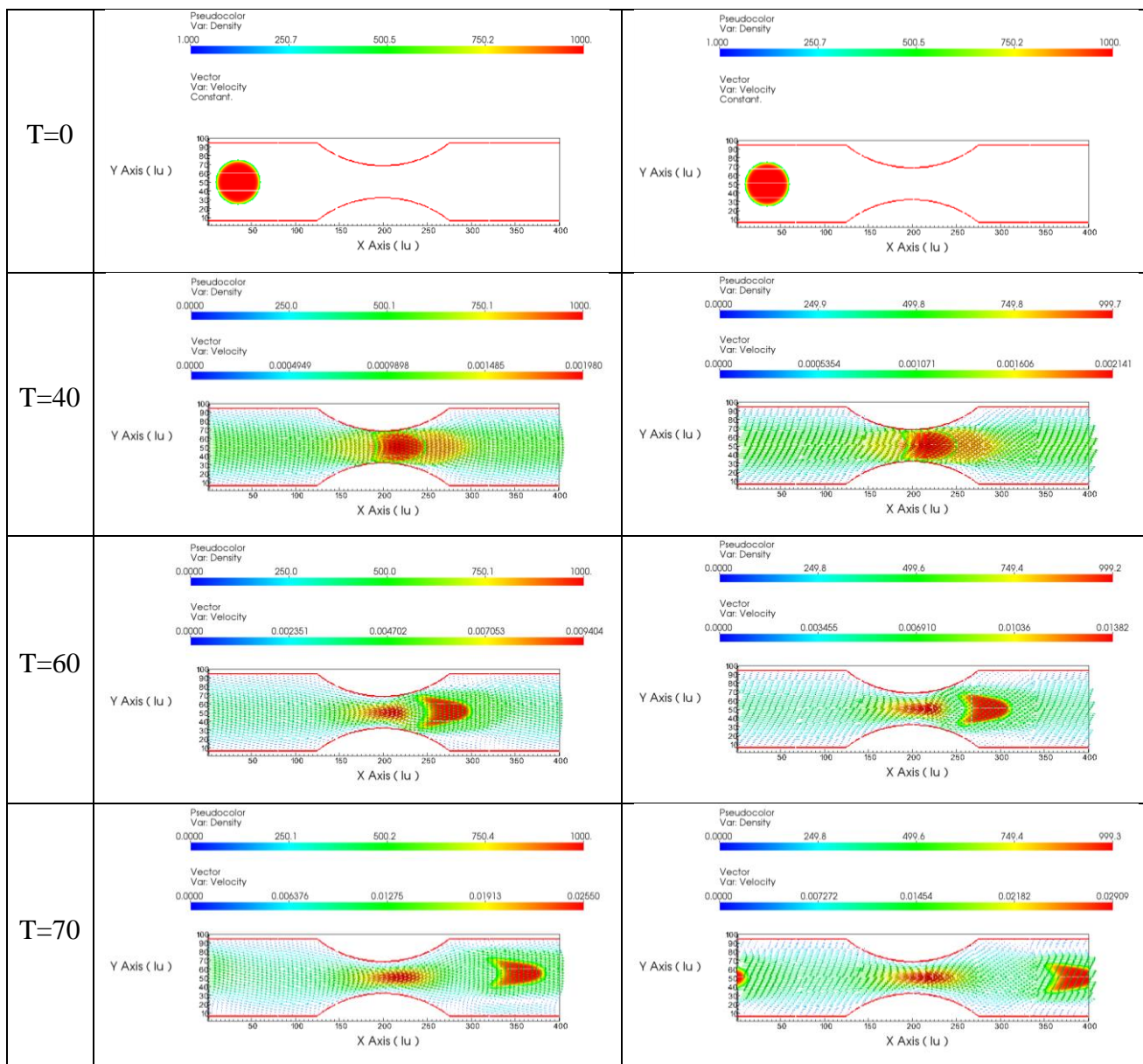


Figure 39. Droplet flow in a constricted channel with attractive walls (left) and repulsive walls (right).

A simulation has also been conducted involving buoyant bubble rise in a complex geometry environment in order to evaluate the performance of the LBM code with the geometry import and contact line dynamics features for a case that may be relevant to the DOE waste tank operations. The simulations aimed to test the applicability of the LBM simulations for the estimation of buoyant displacement gas release event that may occur in the DSTs. To mimic the complex geometry of a waste tank with the cooling structures as shown in Figure 40 for a tank in Savannah River Technology Laboratory, FIU generated a Solidworks drawing for the cross-section of a tank that has similar features.

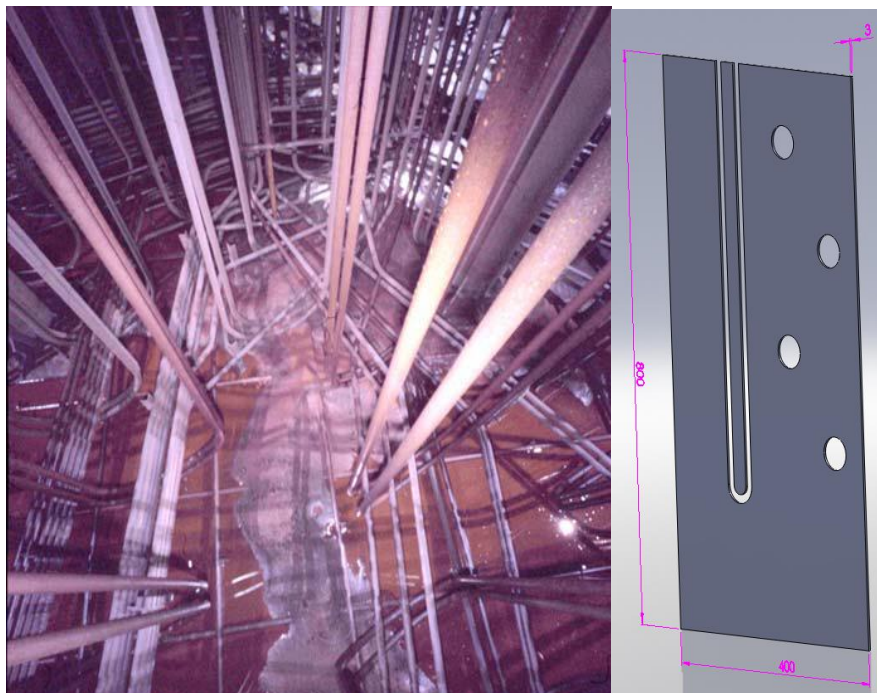


Figure 40. Cooling coils in Tank 1 in SRNL on the left and the cross section of a similar geometry in Solidworks on the right.

Since the LBM code doesn't have the feature of a bubble generation mechanism built-in, we placed 7 bubbles of equal size at the bottom of the tank at a density ratio of 1 to 10 compared to the surrounding liquid. Once the gravity is turned on and the bubbles were let to rise, it was observed that some bubbles were restricted to rise by the horizontal pipe structures represented in the domain as circular objects which delayed the increase of the gas phase in the head-space. Others have come together and formed larger bubbles. The influence of the vertical pipe was found to be less compared to the horizontal pipes represented by the circular solid structures in the domain. The figure below shows the change of the interface of the bubbles as they rise as well as the fluid velocity vectors colored with velocity magnitude. Although the simulation domain was a simple representation of the real complex environment in a waste tank, the potential of the LBM code was proven for the understanding of bubble entrapment and release in the multiphase flow environment in nuclear waste tanks. FIU will try to get the CAD file of such a tank geometry and conduct larger scale simulations including the full tank.

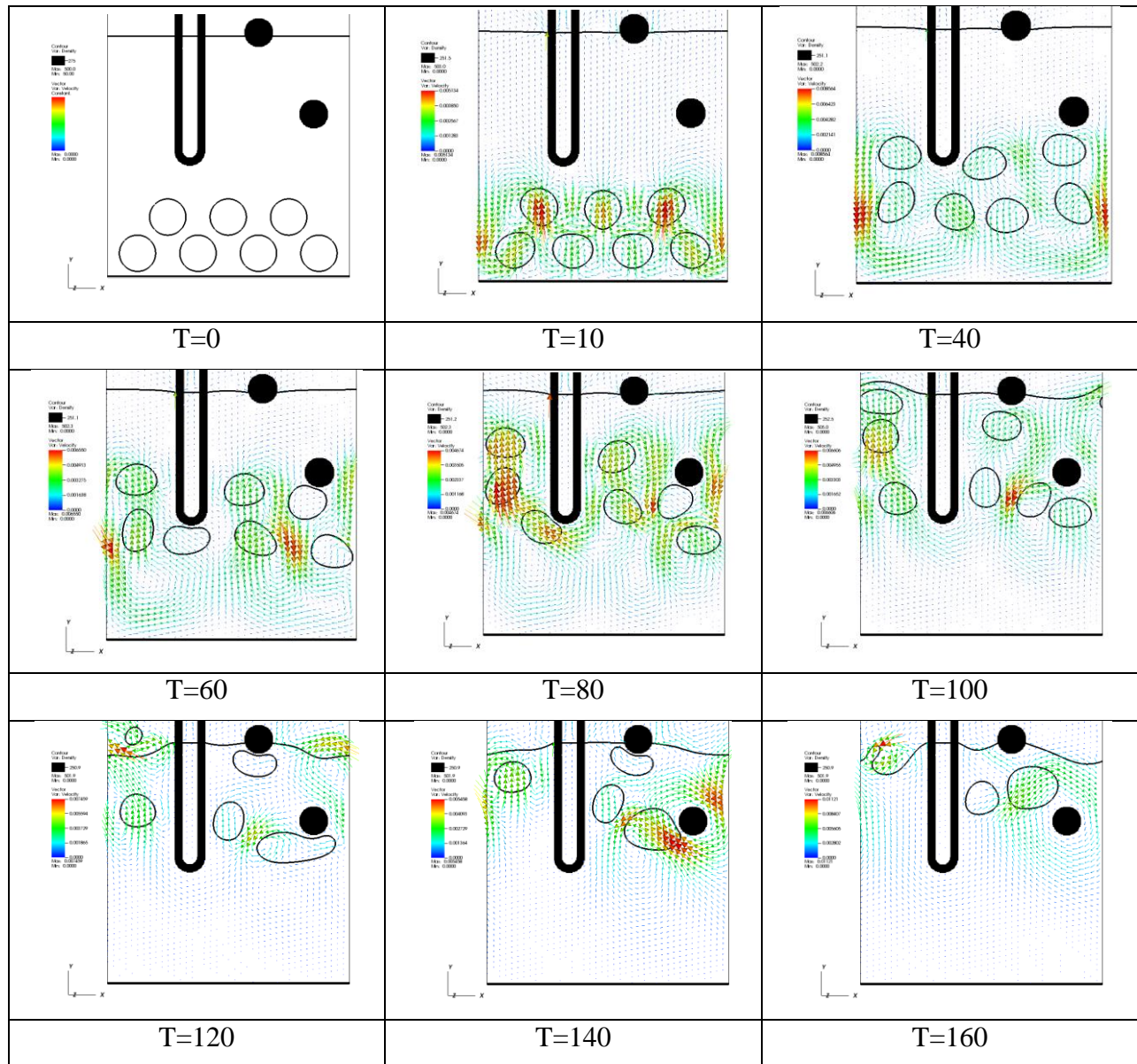


Figure 41. Droplet flow in a constricted channel with attractive walls (left) and repulsive walls (right).

CONCLUSIONS

In this work, we presented a concise literature review on modeling turbulent flows and non-Newtonian flows with LBM. Our study revealed that the extension of the lattice Boltzmann method for the simulation of complex fluid flows including turbulent flows and non-Newtonian fluids is possible via various approaches, although a well-established method has not yet been achieved. In comparison to the turbulence models proposed for LBM, the incorporation of non-Newtonian fluid properties in the LBM simulations have been presented more consistently. The LES-LBE applications show more promise in terms of turbulence modeling with LBM due to its simpler implementation and higher accuracy over two-equation models. It should be noted that the authors were not able to find an application of LBM for turbulent flows in a multiphase flow configuration; therefore, the future LBM task at FIU that aims to simulate turbulent flows in nuclear tanks with multiple phases of fluids would be a challenging and scientifically important effort, especially if both of the subgrid models and Bingham plastic effects could be incorporated simultaneously in the viscosity definition of the lattice Boltzmann equation.

The formulation of a Multiple Relaxation Time LBM based on the Lee and Lin multiphase model was presented for dynamic bubbles and droplets in three dimensional domains with complex geometries and under the effect of wall adhesion and gravitational forces. Numerical cases of moving bubbles and droplets in confined and open computational domains have been presented and the capability of the method to simulate dynamic interface tracking of the fluid interface in contact with the solid surfaces has been shown. The numerical method based on a multiphase LBM established with this research effort was able to provide promising preliminary results for a case that represents similar piping features of a DOE nuclear waste tank. However, further improvement of the method needs to be performed with appropriate inlet and outlet boundary conditions to simulate flows with injections and mixing of fluids.

Future work will include implementation of a turbulence model for injection type flows as well as modifying the viscosity definition in LBM to simulate non-Newtonian fluids. The LES-LBM will be the approach to follow for the turbulent flow simulations with LBM; however, the cascade-LBM will also be investigated if time permits which claims to not necessitate a separate turbulence model within the LBM framework. The extension of the LBM code to non-Newtonian flow modeling will also be conducted in the next fiscal year especially related to Bingham plastic fluids that are found in the DOE waste tanks.

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TASK 15: EVALUATION OF ADVANCED INSTRUMENTATION NEEDS FOR HLW RETRIEVAL

EXECUTIVE SUMMARY

As the DOE's Hanford Site begins preparations for the transfer of high-level radioactive waste (HLW) from the double-shell tanks (DST) to the Waste Treatment and Immobilization Plant (WTP), the influence of waste feed consistency on the waste stabilization process – and final stabilized waste form – is currently under analysis. In order to characterize feed consistency prior to transfer, a suite of instrumentation will be required to monitor the waste preparation and mixing process in real time. FIU has focused its instrumentation efforts on identifying improvements to the in-situ, near-real time monitoring of the mixing process. Specifically, this project has identified innovative technologies applicable for in-tank monitoring during the mixing process.

After the current technology baseline plan and previous research efforts were reviewed, FIU performed an extensive literature and technology search for applicable systems that could provide waste parameters within the HLW tank environment. The literature search focused on available methodologies for in-situ analysis of slurries, emulsions and suspensions applied in all industries. In particular, the monitoring of bulk density and/or particle concentration/characteristics measurements was the focus of the search. For the technology search, vendors of applicable techniques were identified and contacted. The searches resulted in several academic, commercial and governmental reports/articles applicable to the monitoring needs of the HLW tanks. Instrumentation specifications were collected and reviewed to identify the technology capabilities and limitations. These capabilities were also used for a comparative analysis between technologies. Based on the information, all the technologies were down selected to five applicable systems/methods that could provide useful information if deployed in a HLW tank. The five applicable methods applicable to the in-situ monitoring of the waste feed consistency were focused beam reflectance measurement (FBRM), optical back-reflectance measurement (ORM), ultrasonic spectroscopy (USS), Lamb/Stoneley wave viscosity measurement, and vibration-based densitometers. Based on literature results and commercial options, the ultrasonic and vibration-based techniques showed the most promise for developing a technology that could be used for in-situ measurements within the aggressive environment of a HLW tank. Specifically, the vibration-based densitometers and USS systems could provide information on the density and concentrations of the mixed slurry. These techniques had the potential to be engineered for monitoring at various depths within the tank.

In order to evaluate a technology, the USS was tested on an array of slurries, and the density compared to a commercially available density meter that utilizes an accurate and repeatable technique for density measurement. A setup was conceived that allowed the testing of the USS and a commercially-available Coriolis mass flow meter side-by-side. The setup used a 10-gallon tank for agitation of the simulated slurry mixtures, with both systems sampling at the same location; the USS system probe was lowered into the tank, while a pick-up tube was used to transfer mixture to the Coriolis meter. In addition, the USS was subjected to several tests using a benchtop setup for more controlled evaluations of its concentration tracking capabilities. Both

these systems were subjected to solutions and suspensions that would simulate some of the physical and rheological properties of the HLW slurry found in AY-102. The simulated slurries consisted of one to three distinct solids suspended in a water or NaNO₃ solution as the supernatant.

FIU tested the USS between May and October of the previous year with delays associated with hardware and software issues that plagued the device under evaluation. In August, the system had to be returned to the vendor as a result of a hardware fault. A new system was sent to FIU in October, and was used to complete testing. The results indicated that the technology could provide a measurement of density, but the frequency at which it occurs varies with slurry characteristics. If one frequency is selected for analysis, the density values vary more than 10% when compared to the reference value. One major issue that occurred during testing was that the spectral response profile of the USS system changed between hardware changes, and this was something that the vendor attributed to an issue with the reference file used for the tests.

In order to address issues with system inconsistency, additional bench-scale tests were performed at ITS facilities during the first quarter of this performance period. The testing utilized similar materials, concentration profiles for NaNO₃ and solids loading profiles in an attempt to determine the root cause for the inconsistency of results, as well as to compare the results to the FIU data. The USS utilized consisted of a through-transmission configuration, which had a slightly larger transducer than those tested at FIU. The ITS test results indicated good agreement between the measured and calculated densities for solutions, and certain material and water/NaNO₃ suspension. The technology showed significant variance between the measured and calculated densities for complex mixtures (multiple solids). This variance was again attributed to the significant attenuation of the acoustic pulse after traversing the suspension. This attenuation caused issues in the detection of the correct return echo from the suspensions. With incorrect detection, the measured density value showed significant errors (> 20%).

Based on the results of the testing at ITS facilities, it is evident that the technology still requires the following: applied research in the design and selection of the necessary transducers to provide the needed sensitivity; additional engineering analysis to improve signal acquisition and detection at low signal levels; and basic research into the influence of complex mixtures that contain matter that displays an inverse relationship between speed of sound and concentration. With these limitations, the technology is not at the readiness level required for deployment at the WTP waste feed mixing tank AY-102. The USS has several technical limitations that must be addressed before system would be at a TRL readiness level commensurate with a potential deployment in 2-3 years.

INTRODUCTION

The US Department of Energy's (DOE) Hanford Site is completing development of the Waste Treatment and Immobilization Plant (WTP), and preparing for the processing of over 54 million gallons of high-level radioactive waste (HLW). This waste shall be mixed within storage tanks, transferred to staging tanks, and processed by the WTP for final disposal. As the WTP development is completed, the window to integrate additional instrumented systems into the process loop will close. One such area that could benefit from additional instrumentation is the HLW tanks where this waste is currently stored. These +1M gallons of waste must be mixed into slurry that can be retrieved via pipelines to other tanks, or to WTP. The ability to characterize slurry in-tank would greatly optimize the mixing and retrieval processes. After several years of experience in the development of remote monitors for HLW tanks, FIU has undertaken a task to evaluate additional characterization and monitoring technologies that could be scaled-up for deployment inside the HLW tanks at Hanford.

During the 2010 performance period, FIU conducted a literature search for promising technologies that could be utilized in-situ and provide process engineers with near real-time information on slurry parameters during the mixing process. In particular, technologies that could monitor the slurry bulk density were highlighted and reviewed via addition literature search and screening. The search identified two technologies/methodologies that could be deployed in a HLW tank, and could provide useful bulk density information at various depths in the tank. These included a Coriolis-principle based densitometer, and an Ultrasonic Spectroscopy-based technology (USS). The Coriolis meter utilized was a commercially available Promass 63F by Endress-Hauser, and the USS was a commercially available system from Industrial Tomography Systems plc (ITS). FIU evaluated both these technologies in a laboratory setting with slurry simulants. The technology results were inconclusive; the USS system(s) showed varying results for the same slurry characteristics. These variations were attributed to a system performance issue. In order to address these variations, an additional round of testing was performed at ITS on a bench-scale setup with more controlled suspensions. This report summarizes the tests performed as well as addresses a timeline for data analysis.

EXPERIMENTAL APPROACH

In order to mitigate system issues from affecting the test success, it was determined that the tests should be performed at ITS facilities. The vendor would be capable of addressing any system issues during testing, and would follow the test requirements and goals defined by FIU. In order to collect sufficient information to ensure that the technology could proceed into a prototype development stage, the tests were divided into two phases: a (1) NaNO₃ concentration ladder phase, and a (2) solids loading ladders tests with different materials. The first phase involved concentration ladder of aqueous NaNO₃ solutions. A mixture was prepared and static (no mixing) measurements were taken in the cell for the following concentrations: 0.1%, 0.5%, 1%, 2%, 3%, 5%, 7%, 10%, 15%, 20%, 25%, 30%, and 37.5%. For the second phase, distilled water was used as the continuous phase in aluminum hydroxide, zirconium oxide and stainless steel mixtures and further on the ultrasonic measurements of the mixtures. Measurements were taken for the following concentrations (by volume): 0.1%, 0.5%, 1%, 2%, 3%, 5%, 7%, 10%, 15%, 20%, and 25%.

In this test program, the USS was configured in a through-transmission mode. This would address the expected large attenuation caused by the solids utilized. It is expected that a field deployable system will have to utilize a through-transmission system in order to address the large attenuation expected in Hanford's HLW tanks. All measurements were carried out using a USS in-line measurement cell in the static and dynamic conditions. The experiments consisted of 10 repeated measurements for each sample. The sample was then removed from the cell for disposal and new material was added. In the case of sodium nitrate concentration ladder, the spacing between transducers was set to 22-mm. That spacing was reduced to 10-mm for suspensions phase. The system was calibrated using deionized water as a reference media. Additional parameters of the experiment: room temperature, the volume of the sample 350 ml.

RESULTS & DISCUSSION

The test report provided by ITS is presented in Appendix A. Analysis of the NaNO_3 solutions show good agreement with the expected wave speed, and agrees with the data collected during FIU testing. The attenuation data shows that these solutions experience less attenuation than a water mixture. Analysis of the density measurement for the aluminum hydroxide solid loads indicates good agreement with the calculated value, even into the 30% concentration. The measurements collected from the suspensions of zirconium oxide in water showed the inversely proportional velocity to the concentration of the sample. The complex mixture data showed significant variations in the density as compared to the calculated value, which indicates significant attenuation of the ultrasonic pulse.

CONCLUSIONS

The results from this final phase of testing indicates that several areas of the technology still require substantive applied research in order to overcome current limitations. The technology holds potential as a simplified bulk density and concentration method, but it requires additional work in order to overcome current commercial design limitations. There are three areas that require more applied research to overcome the limitation. First, the design and selection of the necessary transducers to provide the needed sensitivity must be reviewed; the benefits of advanced piezo-composite transducers can improve signal sensitivity and fidelity. Second, additional engineering analysis must be performed to improve signal acquisition and detection at low signal levels. The utilization of higher sensitivity acquisition system, as well as better control of thermal voltage levels, can improve minimum detectable signal level. Finally, basic information on the complex mixtures, especially those that contain matter that displays an inverse relationship between speed of sound and concentration, must be investigated. The attenuation characteristics at varying spectra of the expected waste stream must be properly considered when evaluating the response to a broadband ultrasonic pulse.

TASK 16: COMPUTATIONAL SIMULATION AND EVOLUTION OF HLW PLUGS

EXECUTIVE SUMMARY

Pipeline plug formation is caused by changes in the chemistry and flow patterns within the pipe transfer system at Hanford. A better understating of the interactions between the chemical species leading to precipitate formation is required to reduce the risk pipe plugging. A need exists for a computational tool that can predict plug formation by considering the chemistry dynamics coupled with fluid particle interactions. The use of Computational Fluid Dynamics (CFD) software has been explored in the past to predict plug formation [1, 2]. The plugging mechanism simulated was settling of solids. Even though the efforts were promising, the models lacked incorporation of chemical reactions kinetics. Hence, a new task was initiated as part of Florida International University's research efforts to develop a multi-physics model using CFD software that could simulate the coupled flow and chemistry kinetics and aid in understanding the plug formation process.

A literature review was conducted using major search engines to advance the plug formation knowledgebase, with specific applications to the Hanford Site. The review focused on four areas: (i) Overview of the waste transfer system at Hanford, (ii) Plugging mechanisms and waste transfer dynamics, (iii) Current analysis tools-capabilities and limitations, and (iv) Use of computational fluid dynamics (CFD) modeling to predict the plug formation process.

Several models were evaluated simulating single phase and multi-phase reacting flow kinetics using Comsol Multiphysics 4.3bTM for laminar flow. The study of reacting flow involved investigating how the flow affected the chemical reaction process. For the single phase simulations, the baseline model consisted of a reacting flow entering a horizontal pipe section. The influence of bends on the reacting flow field was analyzed by adding a 90° elbow and a straight pipe section to the baseline model. The evolution of species was investigated as they were transported through the pipe. The velocity and concentration profiles were plotted to analyze the effect of flow field on the chemical reaction process. For the multi-phase simulations, the model consisted of a mixture flow (10% volume fraction of clay dispersed in water) through a horizontal pipe. The settling of solids along the pipe length was studied and the volume fractions of the continuous phase (water) and the dispersed phase (clay) were plotted. The results from the single phase and multi-phase simulations demonstrated the ability of Comsol Multiphysics 4.3bTM to simulate coupled flow chemistry models. Future work will include modeling single phase and multi-phase reacting flow at turbulent velocity, coupling chemical reactions to the multi-phase flow and incorporating dynamics processes such as precipitation, particle agglomeration, particle breakup to the models to gain a better understanding of reaction kinetics that lead to plug formation.

INTRODUCTION

Hanford's waste is a chemically complicated system. The main goal of Hanford's waste transfer system is to transfer such waste without plugging the transfer pipelines. The current tools that are used currently to support this objective include the Environmental Simulation Program (ESP) and critical velocity correlations. ESP is used to estimate the initial waste compositions and solids volume fraction and critical velocity correlations are then used to calculate flow velocities and pressure drops for waste transfers. The tools are used in tandem to mitigate the risk of plugging by ensuring that the waste meets the waste acceptance criteria (WAC).

Despite such efforts, several lines have plugged during the waste transfer process at Hanford. The plugging has been attributed to two main factors: precipitation/crystallization and settling of solids. Both ESP and critical velocity correlations address only one plugging mechanism—settling of solids. The tools do not account for precipitation dynamics. Computation fluid dynamics (CFD) models developed by others (MSU DIAL and PNNL) also do not incorporate chemical interactions and precipitation kinetics.

Changes in the chemical environment and fluid dynamic forces can result in dramatic changes in slurry properties and flow behavior during a transfer. Agglomeration and particle breakup due to changes in the chemical environment, particle-particle interactions, and precipitation, can result in waste behaviors that cannot be predicted by considering equilibrium chemistry or fluid flow alone.

A multi-physics model that accounts for waste chemistry coupled with fluid dynamics was proposed by FIU to address the identified shortcomings. In this report, results of the single phase reacting flow models and multi-phase mixture model are presented. The CFD model couples the two components: (1) Flow Interface and (2) Chemical Reaction Interface. The flow interface is solved first and its calculated velocity field is used as a model input for the chemical reaction interface to study the evolution of chemical species. The CFD software used for the simulations is Comsol Multiphysics 4.3b™.

The outline of the report is given as follows: First, the findings of a comprehensive literature review are presented. The review focuses on four areas: (i) Overview of the waste transfer system at Hanford, (ii) Plugging mechanisms and waste transfer dynamics (iii) Current analysis tools—capabilities and limitations and (iv) CFD modeling to predict the plug formation process. Then, the single phase and multi-phase simulations models are discussed and the results obtained are reported. Finally, conclusions are drawn and discussions for future work are presented.

Background

History of Plugged Pipelines

Plugging was first observed during the cross-site transfer operations conducted in the late 1950s and early 1960s [3]. The reason for the plugging was attributed to insufficient heating of the waste during transfer. As the slurry cooled, it was believed that solids precipitated out of solution and began to accumulate on the pipe walls. Eventually, the solids would block the pipe. The plug was blown out of the pipe with a 200 psi flush fluid. A more thorough investigation of line plugging event was documented in 1978 by [4] during which formation of tri-sodium phosphate

crystals was reported. The crystals impeded the flow within the pipe and completely plugged the pipeline. The formation of these phosphate crystals was studied extensively in the 1980s.^{2,3}

Several pipeline plugs were observed during salt-well transfers and a majority of them were attributed to precipitation of sodium phosphate [5]. In each case, the transfer of the waste was suspended, and the pipeline was not flushed with water. The line from tank BY-103 was pumping waste without dilution. Due to a flammable gas concern, the pumping was shut off without flushing the line. Some month's later a plug was discovered in the pipeline. The plug could not be broken and the line was abandoned. The transfer line from tank SX-104 was pumping waste with ineffective dilution. The pumping was stopped temporarily. On restart of the transfer operations, a plug was discovered in the unheated jumper. The plug was eventually broken using a hot water flush. Pumping was stopped in line from tank U-103 and the line was not flushed for over 24 hours. On restart, the plug was observed in a non-heat traced flexible jumper. A jumper was pulled to break the U-103 plug. In an effort to avoid further plugs, the modifications such as larger flex jumpers and additional heat tracing were made to the transfer system.

Plugging Mechanisms

The primary mechanisms that contribute to transfer line plugging at Hanford were identified from various reports [6, 7, and 8] and are as follows:

- (a) Chemical Instability (precipitation, crystallization)
- (b) Settling of solids – including hydrodynamic instability

Depending on the mechanism, plugging or its onset may result in a change in the particle size distribution or the density of the slurry. In addition, the onset of pipeline plugging may also lead to an increase in pressure drop in the transfer lines.

Chemical instability: Chemical instability during waste transfers results in a phase change (from liquid to solid) initiated due to drop in temperature, changes in local concentration or mixing and pumping of wastes that are not in equilibrium. The solid phase precipitates or crystallizes depending on the solubility characteristics of the dissolved multiple species of the waste, their chemical interaction, and temperature. The solids may precipitate out of the solution and accumulate along the pipe walls. The pipe walls would then serve as a nucleation site where the solids would nucleate and grow rapidly and eventually form an interlocking needle-like crystal network. The needle-like crystal network then impedes the flow within the pipe and commences the formation of the plug. At Hanford, the crystallization or precipitation process has mainly been observed in phosphates-and aluminates-containing supernatant wastes.

² Internal Letter No. 65453-80-296 by D. L. Herting, "Evaporator Feeds High in Phosphate," Rockwell Hanford Operations, Richland, Washington (1980).

³ Internal Memo No. 65124-068-80, by C. H. Delegard, "Viscosity/Cooling Data for Tanks 107-S and 105-BX Waste Liquors," Rockwell Hanford Operations, Richland, Washington (1985).

Phosphate salts can form needle like crystals at low temperatures (40 to 60 °C) at typical concentrations found in Hanford waste types. The effect of temperature on phosphate crystal formation was investigated and reported in an internal letter [9]. The author found that at temperatures above 55°C, granular sodium phosphate crystals ($\text{Na}_3\text{PO}_4 \cdot 6\text{H}_2\text{O}$) were formed and that at temperatures below 55°C needle like sodium phosphate crystals ($\text{Na}_3\text{PO}_4 \cdot 12\text{H}_2\text{O}$) were formed. The main difference between the two forms was differing waters of hydration. The author made a key observation that “the viscous suspensions of $\text{Na}_3\text{PO}_4 \cdot 12\text{H}_2\text{O}$ needles could be converted by prolonged agitation to a well-settling mixture composed of 25% settled solids by volume”, indicating that the gel is shear-thinning, believed to be largely due to alignment of needle crystals. The author concluded that if a gel of the 12-hydrate crystals were to form a plug in a pipeline, it would be difficult to impart sufficient energy into the plug to cause shear thinning and remove the plug. In general, the formation of needle crystals is inhibited at high temperatures.

The solubility of sodium phosphate is also highly temperature dependent. Sodium phosphate solubility as a function of temperature and sodium ion concentrations was investigated [10]. He found that solubility of phosphate is strongly temperature dependent with a linear dependence on sodium ion concentration at 50°C and 60°C. For example at 50°C, at 8 M sodium, the phosphate solubility was reported as approximately 0.3 M increasing to 0.45 M at 5 M sodium. At 40°C, the data was much less dependent on sodium ion concentration, and indicated a phosphate solubility of approximately 0.1 M to 0.15 M above 5 M sodium. At 20°C, the solubility of phosphate was noted to be in the range of 0.03 M to 0.08 M. Hence, it can be concluded that phosphate is more soluble at higher temperatures.

The above discussed findings clearly illustrate the potential difficulties that can be encountered while handling phosphate waste. To avoid transfer problems with phosphate containing waste [11] recommended the following precautions: First, the tank waste should not be heated prior to the transfer. Second, the waste should not be allowed to cool during the transfer. Third, the waste should be kept moving during the transfer.

In addition to phosphate crystals, the precipitation of aluminates also has the potential to plug slurry transfer systems at Hanford. Aluminates can precipitate by at least two mechanisms: the formation of amorphous or crystalline solid phases such as gibbsite, and the formation of sodium aluminate. The solubility of aluminum was investigated as function of high ionic strengths by [12]. The author presented a phase map (sodium hydroxide vs. sodium aluminate) as shown in Figure 42 and found that the solubility of aluminum hydroxide was greatly increased at higher ionic strengths. In solutions with high sodium hydroxide concentrations (above 2 M), it was found that increasing aluminum concentration would eventually precipitate sodium aluminate salts. It was also observed that at lower sodium hydroxide levels, an increase in the molar aluminum concentration resulted in precipitation of aluminum hydroxides.

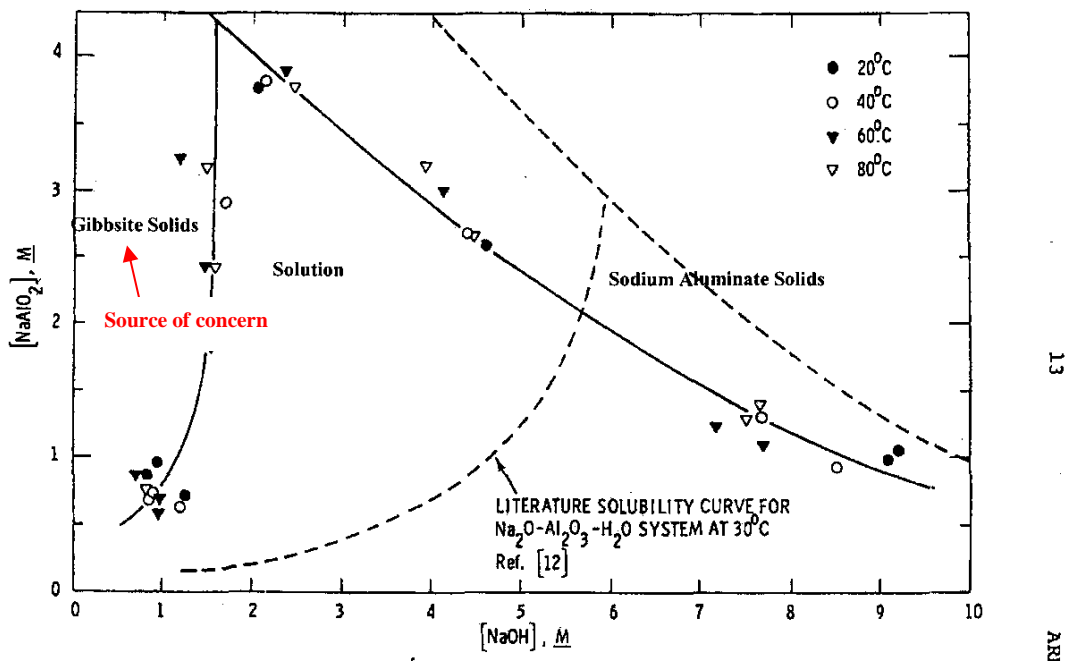


Figure 42. Aluminum solubility as a function of sodium hydroxide [12].

These aluminum hydroxides can take on several forms, including amorphous, gibbsite, and boehmite. The amorphous form is thought to precipitate out more quickly, sometimes forming gels that can be problematic during waste transport. The amorphous hydroxide is usually converted within hours or days into crystalline gibbsite phase. Gibbsite typically forms roughly spherical-shaped particles that are of colloidal size but have been observed in waste streams as stable agglomerates [13]. Gibbsite is thought to be relatively common among the waste aluminum phases and has been observed in samples from Tank 241-SY-103 [14]. Boehmite is thought to be less common in waste, and had been typically observed in tanks (Tank 241-T-111) where self-boiling of waste occurred due to high decay heat [15].

One of the control strategies recommended in [16] while handling aluminum rich waste included avoiding over dilution of aluminum rich supernatant with water as that could shift the waste composition towards the left hand phase boundary at which gibbsite solids could precipitate. The use of caustic solution of sodium hydroxide was recommended for dilution rather than water, to shift the waste composition to the right, away from the gibbsite phase boundary. In the event of a line plugging by gibbsite precipitation, caustic solution was again recommended to be used as a flushing fluid to re-dissolve the precipitates and clear the line. Gibbsite precipitation plugged a U-Farm transfer line during saltwell pumping in 2003. Caustic soak was used to unplug that line.

Settling of Solids: The presence of precipitates and/or agglomerates increases the solids concentration and increases viscosity of the slurry. The flow can transition from turbulent to laminar flow as a result of such changes during transit and the undissolved solids may settle when the flow velocity is not sufficient to keep them suspended. A moving bed of particles begins to accumulate during a slurry transport operation. Settling solids in a moving bed of particles form a stationary bed that eventually fills the pipe and blocks flow. Critical velocity correlations are used to estimate the minimum velocity to prevent settling of solids.

Influence of Waste Transfer Dynamics

In order to accurately evaluate the risk of plugging, all the plugging mechanisms that lead to plug formation need to be thoroughly understood and investigated. Much of the current waste transfer analysis focus has been on only one plugging mechanism - settling of solids which assume that the PSD is static. However, in actual waste transfers the PSD is dynamic due to particle-particle interactions, precipitation, agglomeration, chemical reactions, etc. A change in chemical environment influences the properties and quantity of solids in the waste stream, which in turn influences the slurry viscosity. The slurry viscosity and the solids properties both influence the flow properties. Chemistry affects flow and flow affects chemistry. The chemical-flow coupling must be taken into consideration while evaluating the risk of plugging and be incorporated to accurately predict plug formation [17]. Figure 43 shows the chemical-flow coupling and its influence on various slurry and flow properties.

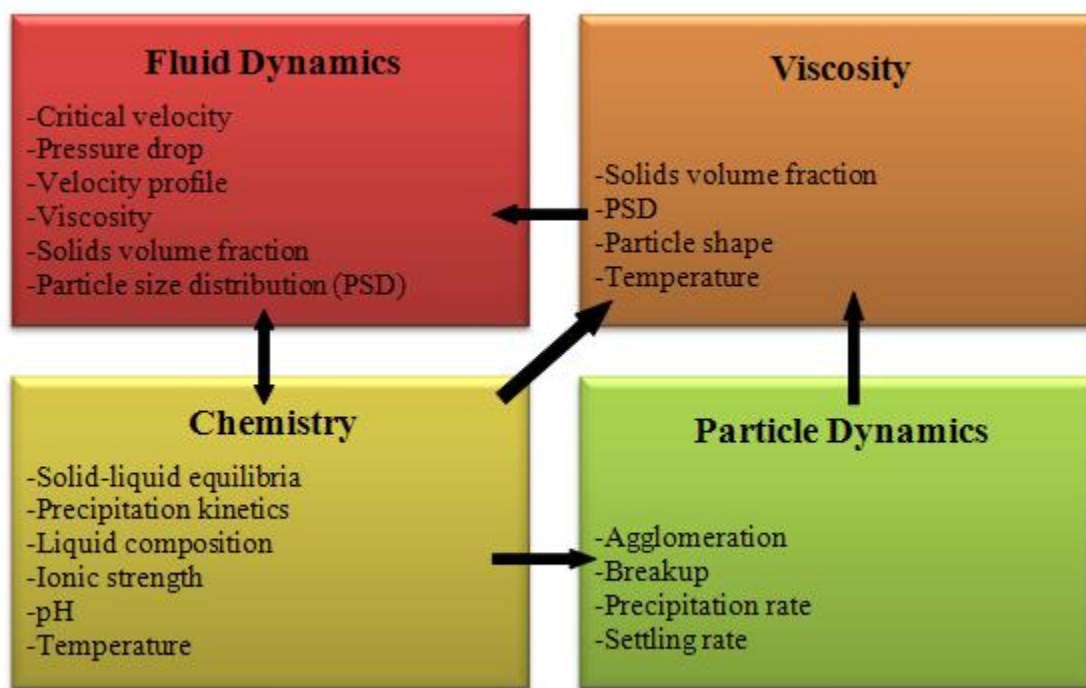


Figure 43. Chemical-flow coupling (adapted from [17]).

Waste Transfer System

The underground storage tanks from which the waste must be retrieved comprises of 149 older single-shell tanks (SSTs) and 28 newer double-shell tanks (DSTs). Together the tanks contain about 56 million gallons of high-level radioactive and hazardous waste. The waste was generated during nuclear weapons production from the 1940s through the late 1980s. The storage tanks are located in two physically separated areas called the 200 East Area and 200 West Area, about seven miles apart. A waste transfer system is used to facilitate the transfer between tanks and from tanks to processing facilities. A brief description of the key components is provided in the following sections. The description of the transfer systems and technical specifications were gathered from various documents [5, 6, 18, 19, and 20]

The primary components of waste transfer system include:

- Transfer lines
- Ancillary equipment
 - Pump Pits
 - Valve Pits
 - Diversion Boxes

A typical transfer line to Tank SX-104 farm is shown for illustration purposes in Figure 44. This transfer line included a valve pit between the tanks. A typical valve pit includes vertical and horizontal pipe sections as well as 90° bends with no radius elbows (miter bends). There is plugging potential at the different bends present in the valve pit as they create low flow zones. In addition, there is an elevation between the tanks. Inside the valve and pump pits, the transfer lines are connected remotely by short sections of pipe called jumpers that provide the flexibility and maintenance capability for the transfer line network. The jumpers are connected to nozzles inside the diversion boxes to route waste between any two tanks in the transfer system.

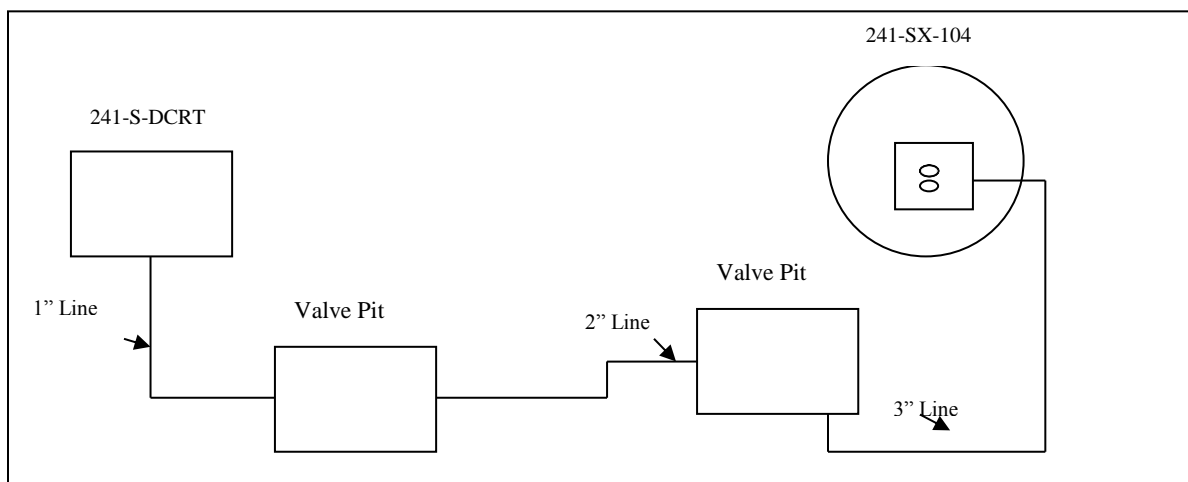


Figure 44. Transfer line for tank SX-104 (Adapted from [5]).

A brief description of the components is provided in the following sections.

Transfer Lines

The transfer lines at Hanford have evolved over decades from varying materials of construction. The early transfer lines were largely made from 300 series stainless steel and were single-walled, direct buried. These have evolved to either stainless steel or carbon steel encased in a concrete encasement (pipe-in-concrete) to the present design of carbon steel or stainless steel pipe encased in an outer pipe of carbon steel or stainless steel (pipe-in-pipe). A brief description of the types of transfer lines found at Hanford is provided below:

Underground Transfer Lines

- a) **Direct Buried Transfer Lines:** These transfer lines are single walled (no encasement) and are buried directly in the soil. Examples of such lines include drain pipes and ventilation pipes. These are made of carbon steel and are typically 1 to 3-in diameter welded pipe. They are buried under 3-ft of ground cover, either in shallow trenches or at ground surface level.

- b) **Encased Transfer Lines:** In these types of transfer lines, the primary pipe is completely enclosed within a secondary confinement barrier. The secondary confinement barrier usually is a larger pipe (pipe-in-pipe), or a concrete jacket (pipe-in-concrete). The primary and secondary pipes are made of carbon steel or stainless steel. The transfer lines are buried at nominal depths of 2.5-ft to 4-ft. The lines are sloped a minimum of 3 inches in 100-ft to facilitate complete draining of liquids and to minimize corrosion impacts over the life of the pipe.

Overground Transfer Lines

- a) **Hose-in-Hose Transfer Lines:** These transfer lines are above ground and are used for shorter, temporary waste transfer routes. The transfer lines consist of a 2-in primary hose encased in a 4-in secondary hose and hence are known as Hose-in-Hose Transfer Lines (HIHTL). The primary hose is fitted with hose end connections suitable for attaching to valve manifolds with remote connectors. The primary hose is wrapped with heat tracing and the secondary hose is insulated to aid in maintaining acceptable temperatures during a transfer, as a means to mitigate the potential for forming a plug in the line, and to prevent freezing due to low ambient temperatures (Figure 45). In the event of a plug in the hose-in-hose transfer lines, the above ground hoses are replaced on these routes. For transfer lines too long to be made out of one continuous length of hose, two or more HIHTL assemblies are joined at mid-point connections to establish the required route.

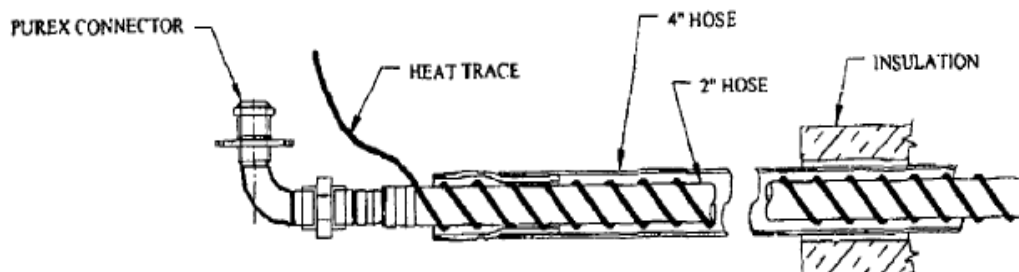


Figure 45. HIHTL configuration [21].

Ancillary Equipment

The transfer lines interface with associated ancillary equipment during the transfer operations. The transfer lines originate in a pump pit and end in a valve pit. Inside the valve and pump pits, the transfer lines are connected remotely by short sections of pipe called jumpers that provide the flexibility and maintenance capability for the transfer line network. The jumpers are connected to nozzles inside the diversion boxes to route waste between any two tanks in the transfer system.

Pump Pits

A pump pit is a concrete structure located above the tank dome near the center of the tank. It is used to collect spills or leaks from the equipment within the pit. The pump pit contains transfer and agitator pumps and jumper connections to the transfer lines and valves. Tank waste flow from the SST is routed through the pump pit to a DST. Each transfer pump is equipped with a leak detector, which will detect the leak in the pits. Each of the 28 DSTs has a pump pit and there are 117 pump pits associated with SSTs depending on the type of waste stored and the

function of the SST. The pits typically have a floor drain that drains directly back to the SST it services.

Valve Pits

Valve pits are below ground reinforced concrete structures that contain valve and jumper assemblies to route the liquid waste through the connected pipelines within a tank farm. Heavy, thick, grade-level blocks cover each of the valve pits. When several tanks are undergoing simultaneous pumping to a single receiver tank, the flow is routed to a valve pit. In the valve pit, the transfer lines of the sending tank are manifolded to the receiver tank line by means of a series of valves and jumper connections. There are 17 valve pits associated with SSTs and 8 interconnected valve pits in the six DST farms. Each valve pit is equipped with leak detection that is interlocked to shut down pumps. Each valve pit also has a flush line connected to a flush pit or a drain line connected to an underground catch tank.

Valves: Two- and three-way valves are built into each jumper assembly to divert the flow in the required direction. The two-way valves are typically 2-in and 3-in full port, ¼-turn ball valves and the three-way valves are 3-inch T-ported ball valves. The valves are manually operated via T-handles that extend through the pit cover block to allow shielding for remote operations.

Jumpers: Jumpers are short removable sections of piping that are used to connect transfer lines as well as to connect nozzles inside diversion boxes and pits. There are approximately 220 DST and SST jumpers in the waste transfer system. Jumpers are either rigid or flexible and vary in length from 3-20 ft. The jumpers include 2-in Schedule 40 pipe for slurry transfer and 3-in Schedule 40 pipe for supernatant transfers. A number of lines have formed plugs in flexible jumpers and non-heat traced jumpers [5].

Diversion Boxes

Diversion boxes are below-grade, reinforced concrete structures that provide a flexible method of directing liquid waste from a process facility or tank to another process facility or tank. The top of the diversion box is a concrete cover block that usually extends above grade. The cover blocks vary in thickness from box to box. Some diversion boxes are lined with steel. Transfer lines terminate at a nozzle in the wall or floor of the diversion box. Transfer lines are connected in the diversion box by installing a jumper between the connecting nozzles (Figure 46). Diversion boxes contain drain lines connected to catch tanks that together serve to collect any waste leakage from the jumper connections. Diversion boxes also provide containment for leaks in encased waste transfer lines, which drain back to the diversion boxes through the encasement.

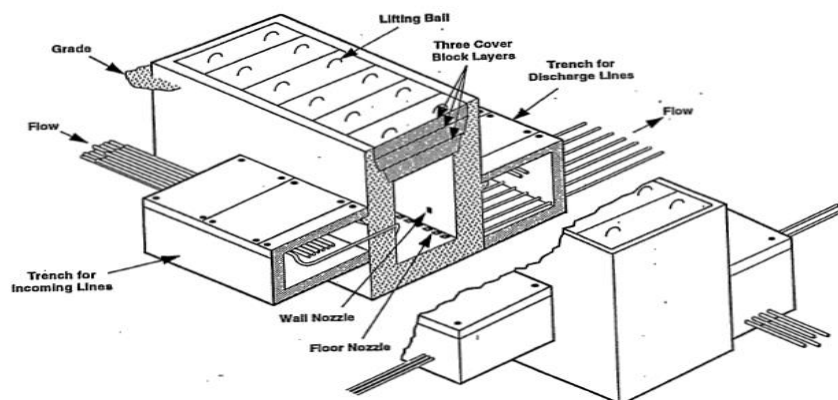


Figure 46. Diversion box details [22].

Existing Cross-Site Transfer System

The existing cross-site transfer system (ECTS) has been in service for 50 years and is used to transfer waste between 200 East and 200 West Area (Figure 47). The ECTS is approximately 3.5 miles long. It consists of six stainless steel pipelines, 3-in in diameter buried inside a steel reinforced concrete encasement and a vent station. The encasement is 5-ft wide x 2-ft high and has a 6-in high void space to accommodate the transfer lines. The encasement is buried from 5 to 15-ft below grade, depending on location. The encased void space is accessed via 58 swab risers spaced regularly along the pipeline. These risers are sealed with pipe caps and opened only to inspect the interior of the encasement and to check on the condition of the transfer pipes. The risers do not provide access to the interior of piping, but to the interior of encasement. There is no provision for sampling access to the cross-site transfer lines. Without drilling or cutting the lines, access to the line interiors is limited to the nozzles in the two diversion boxes. The diversion boxes used to re-route waste to other diversion boxes within the tank farms are constructed of reinforced concrete and measure approximately 45-ft long, 10-ft wide, and 17-ft deep. The vent station, 241-EW-151, is located roughly midway between the 200 East and West Areas and serves as an air exhaust intake point to vent the lines during waste transfer and flushing. The vent station is also made of reinforced concrete, measuring approximately 17-ft long, 10-ft wide, and 17-ft deep. From the vent station, the encasement slopes downwards in both directions and drains liquids back to the diversion boxes. The lines terminate at two diversion boxes, 241-ER-151 and 241-UX-154 in the 200 East and West Areas, respectively, where they interface with 200 East and West Area transfer piping.

The ECTS is only partially usable, given that 4 of the 6 lines are either plugged or failed, and the other 2 lines may or may not prove usable after pressure testing. Recently a single line was tested and approximately 435,000 gallons of supernatant was transferred from Tank 102-SY in the 200 West Area, to a DST in the 200 East Area [23]. However, the continuing viability of these lines is uncertain and cannot be relied upon. In addition, the ECTS was built prior to environmental protection requirements implemented by the Environmental Protection Agency under the Resource Conservation and Recovery Act (RCRA), 40 CFR 264.193, which requires double containment and leak detection capabilities for all hazardous waste tanks and ancillary equipment such as piping. For these reasons, the ECTS was removed from service in the late 1980s. A new replacement cross-site transfer system (RCST) was proposed to replace the existing pipelines.

Replacement Cross-Site Transfer System

The proposed replacement cross-site transfer system (RCSTS) will be approximately 6.2 mi long and consists of two new parallel encased pipelines (3-in diameter stainless steel pipes, each encased in a 6-in carbon steel outer pipe) to provide secondary containment as well as one diversion box, one booster pump, a vent station, and all associated instrumentation and electrical connections. The two pipes will connect the 241 SY-A and 241 SY-B valve boxes in the 200 West Area with the 244-A Lift Station in the 200 East Area (Figure 47). The line will be capable of transferring slurry waste from the SY Tank Farm in the 200 West Area to 200 East Area and supernatant waste in either direction. Non-slurry, low activity liquid waste can be transferred from 200 East Area to 200 West Area using the existing 200 East Area Tank Farm transfer pumps. The lines are sloped at least 0.25% to allow gravity draining and would be buried or shielded for radiation and freeze protection. The pipelines are designed to prevent corrosion from the metal pipes contacting the soil. Both pipelines are to be insulated with polyurethane foam and covered with a fiberglass jacket. Existing valve pits would connect the RCSTS to existing pipelines to facilitate supernatant waste transfer between the 200 West and East Areas. A booster pump will be located in the diversion box and provides the power to transfer slurries at the minimum required velocity to prevent the lines from plugging. A vent station will be located at the high point of the transfer system to introduce air into the lines after a transfer to facilitate draining the primary containment pipes. Both the diversion box and the vent station will be equipped with stainless steel liners and have provisions for washing down radioactive contamination, collecting accumulated liquids, and routing the liquids back to the tank farms via the RCSTS. The diversion box and the vent station will have connections for attaching portable ventilation systems during maintenance. A concrete cover with access blocks provides radiation shielding and weather protection from rainwater and snow melt.

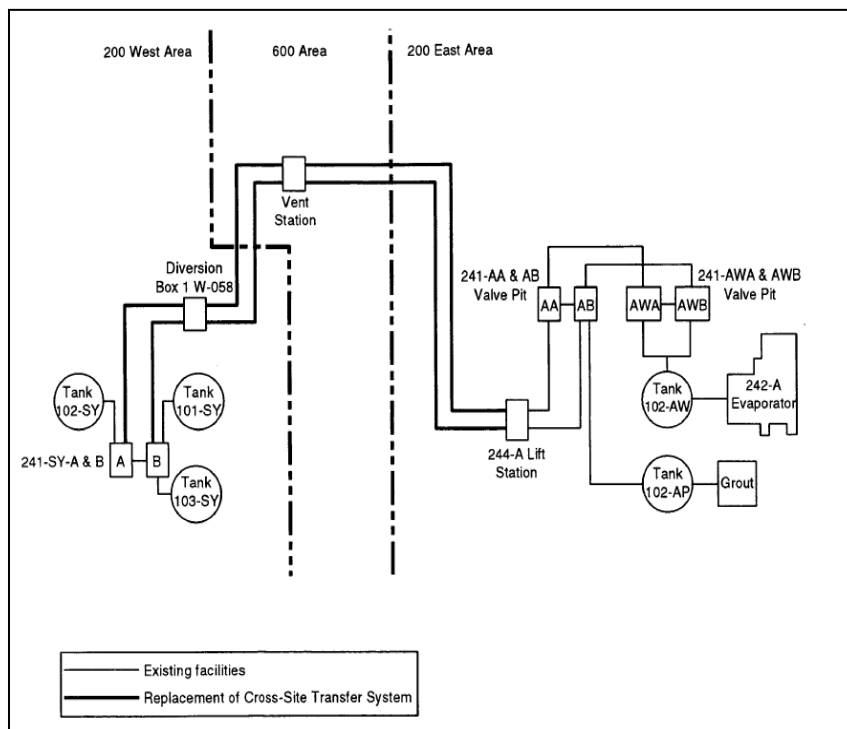


Figure 47. Trace of the replacement cross-site transfer system [24].

Key Technical Details of the Transfer System

Some of the key details related to waste transfer system were summarized as follows:

- The 200 Area DSTs uses buried encased transfer lines to transfer waste from tank to tank farms. The supernatant transfer lines are 3-in NPS with 6-in NPS encasements and the slurry transfer lines are 2-in NPS with 4-in NPS encasements.
- All of the DST transfer lines are of double-wall construction, with the innermost waste-carrying pipe encased within a larger pipe that provides the secondary containment barrier. The outer containment pipe is non-pressurized and includes leak detection.
- The lowest pressure rated lines (275 psig) exist within the inter-farm pipelines and the higher-pressure rated lines (400 psig) connect the farms together and to the WTP. The cross-site lines and the WTP feed routes are longer and hence consist of higher-pressure rated lines (1490 psig for cross site lines and 1000 psig for WTP feed lines). In the event of unplugging these lines, the maximum operating pressures are limited by PUREX connectors to 400 psig unless they are not part of the unplugging effort.
- The waste transfer lines have standard expansion joints every 10-ft over 10-ft in length with two long radius 90° elbows.
- The cross-site transfer lines have U-shaped expansion joints every 600-ft with 36-in bend radius elbows.
- The transfer lines between tanks are several hundred feet long, the transfer lines length to WTP are approximately 3000-5000- ft long and the cross-site transfer lines connecting the SY Farm in the 200 West Area with the AN Farm in 200 East Area are about 29000-ft long.

Current Analysis Tools-Capabilities and Limitations

There are two predictive tools that are used in tandem at Hanford to evaluate the waste transfer and plan for waste feed and delivery. These include the Environmental Simulation Program (ESP) and Critical Velocity Correlations. A brief overview of the capabilities and limitations of these tools is discussed below.

Environmental Simulation Program (ESP)

ESP is a user friendly chemical equilibrium model used extensively at Hanford since 1982. It comprises of a large number of computational modules and extensive databanks. The computational modules can simulate steady state processes (Process Module), perform single or multiple point equilibrium calculations for a single composition (Express Module), do water analysis (Water Analyzer Module), simulate batch process dynamics (DynaChem) as well as do sensitivity analyses (Sensitivity Module). The extensive databanks contain thermodynamic and physical properties of various chemical species as well as parameters for the computational modules. The databanks are provided with the program (known as 'public' databanks); however, the user can create and use their own generated 'private' databanks. At Hanford the program has been used to simulate many processes such as caustic leaching of sludges to dissolve aluminum, salt dissolution to separate salts from salt cakes, purification of salts by evaporation, fractional crystallization, and washing, precipitation and separation of plutonium from acid wastes, scrubbing of nitric oxides to produce reusable nitric acid, speciation from nuclear waste on elemental and anion analyses and many other processes [25].

The main goal in use of this program for waste transfer analysis is to determine if the proposed transfer will cause a problem based on the chemistries of the wastes. Prior to transfer, ESP is used to estimate the initial solids fraction of solids for salt-solution transfers and estimate initial waste composition. It is also used to identify precipitation concerns and estimate the dilution required to mitigate the precipitation risk.

The prediction of chemical reaction kinetics is currently not supported by ESP. For example, it can evaluate for precipitation concerns, but cannot predict the rate at which the precipitates might form. The reaction kinetics is crucial as it can provide better insights into waste transfer planning and routing. For instance, it is known that phosphate precipitation is very fast (within hours) whereas gibbsite precipitation is relatively slower (days to weeks). Hence gibbsite precipitation is not a concern for shorter transfers such as those between DST to DST. But it is a concern for longer transfer routes such as those associated with low flow rates salt well transfers.

Another limitation with the program is that the program provides no information about the concentration of any intermediates formed in transit; it just gives information on final products. Nor can the program predict the formation of gels that are formed due chemical-flow coupling [17].

Critical Velocity Correlations

Critical velocity is the minimum velocity required to keep solids in the pipeline suspended so that they do not settle and plug the lines. Numerous correlations for determining critical velocities of particulate slurries flowing in pipelines exist in the literature. An extensive review of the published critical velocity correlations was conducted by [26]. The authors defined critical velocity as “slurry transport at the lowest velocity that prevents deposition of either a stationary or a moving bed of solids”. Six models were examined and the correlations developed from most of these models were based on slurries with solids having relatively large but narrowly distributed particles size (150 μm mean particle size). However, evaluation of particle size distribution of Hanford tank wastes done by [27] revealed that Hanford tank wastes have a broad PSD with a mean particle size of 110 μm and a substantial fraction of the particles being less than 100 μm . The models from [28] and [29] included some particles in the 100 μm range and 150 μm range respectively and hence were shortlisted. An earlier analysis by [30] of several critical velocity correlations from 1953 to 1980 also had recommended model from [28] because it yielded the most conservative critical velocity results for particles less than 100 μm diameter.

The Oroskar and Turian model is currently the basis of the WTP design guide [31] and the correlation for critical velocity is given by

$$v_{cr} = \sqrt{gd_p \left(\frac{\rho_s}{\rho_L} - 1 \right)} 1.85C_S^{0.1536} C_L^{0.3564} \left(\frac{d_p}{D} \right)^{-0.378} \bar{N}_{Re}^{0.09} X^{0.30} \quad (75)$$

where,

v_{cr} = critical velocity

g = gravitational acceleration

d_p = median particle diameter

ρ_s = solids density

ρ_L = liquid density

C_S = solids volume fraction

C_L = liquid volume fraction

D = pipe inside diameter

\bar{N}_{Re} = modified Reynolds number

$$\bar{N}_{Re} = \frac{D \rho_L \sqrt{g d_p \left(\frac{\rho_S}{\rho_L} - 1 \right)}}{\mu_L}$$

μ_L = liquid viscosity

x = eddy velocity factor

The equation is based on single component density particles forming narrowband PSD. The use of the equation for multi-component density particles, broadband PSDs, and/or median particle sizes less than 100 μm (typical Hanford waste) requires extrapolation beyond the database used in the development of equation. Moreover, the PSD is assumed to be static while deriving these correlations.

The correlation is applicable for calculating the critical velocity of Newtonian fluids in straight, horizontal piping. When applied to non-Newtonian fluids in horizontal piping, these correlations under-predict the critical velocities [32]. The transfer lines consists of vertical segments, pipe bends, Tee's, reducers, jumpers, connector and various other pipe components which can affect the critical velocity and plug formation process. No such guide is available that can be used to calculate critical velocities of such complex piping components.

CFD Modeling-Capabilities and Limitation of Previous Works

The application of CFD has been explored by Mississippi State University Diagnostic and Analysis Instrumentation Laboratory (MSU DIAL) and Pacific Northwest National Laboratory (PNNL) as an approach to predict plugging of pipeline simulating settling of solids mechanism during waste transfers. A brief review of the methods used and the model limitations is discussed below.

MSU DIAL Phoenix Model: MSU DIAL used the CFD software Phoenix 3.3 to create a multiphase model simulating stationary bed profiles as a function of the flow velocity, PSD, solid and liquid densities, viscosity, and volume percent solids. The critical velocity at which the settling occurs was validated with empirical correlations for the critical velocity. The simulation results were also compared with experimental results. The multiphase flow was simulated using a drift flux method. This is an Eulerian model that postulates that there exists one continuous medium in which there are various dispersed phase components (CHAM n.d.). The mixture of the continuous and dispersed phases behaves as a single fluid with properties that are referred to as the mixture properties. In this formulation, only one set of differential equations is solved to give the mixture-mean velocity at each point and time. Then separate sets of equations are solved, one for each phase, which govern its relative velocities, i.e., their differences from the mean. The later equations are algebraic ones, which are derived from the original conservation equations by neglect of the second-order terms. This entails that the relative velocities are computed by reference only to the local pressure gradients, body forces, and inter-phase friction. The volume fractions occupied by each phase, at each point and time, are calculated at the same time. This method is referred to in Phoenix as the algebraic slip method (CHAM n.d.).

A parametric study was performed to investigate particle settling and sedimentation along the transfer pipe as a function of velocity [1]. The transfer pipe was 39-inches in length and 1-inch in diameter. Particle sizes that were evaluated include 45 μm and 220 μm respectively. Simulations with entrance velocities ranging from 0.5 to 2.5 m/s were obtained, and included the effects of turbulence.

The author observed that the larger, heavier particles settled at the bottom of the pipe much faster relative to the smaller particles, especially at low flow velocities. At flow velocities of lower than 1.0 m/s, a stationary bed flow was formed mostly composed of the 220 μm particles. For velocities of greater than 1.0 m/s, the fluid established a moving bed regime where the particles moved along the bottom of the transfer pipe as shown in Figure 48. At still higher flow velocities, the particles moved much faster and stratified through the fluid layers.

The simulation results provided very good agreement with the experimental results of [33]. The model predicted 1.0 m/s minimum flow velocity for the formation of stationary bed regimes versus the 0.85 m/s reported by the FIU experimental results. The authors concluded that velocities high enough to prevent settling must be used to prevent stationary bed flow and the possible formation of plugs in the transfer line.

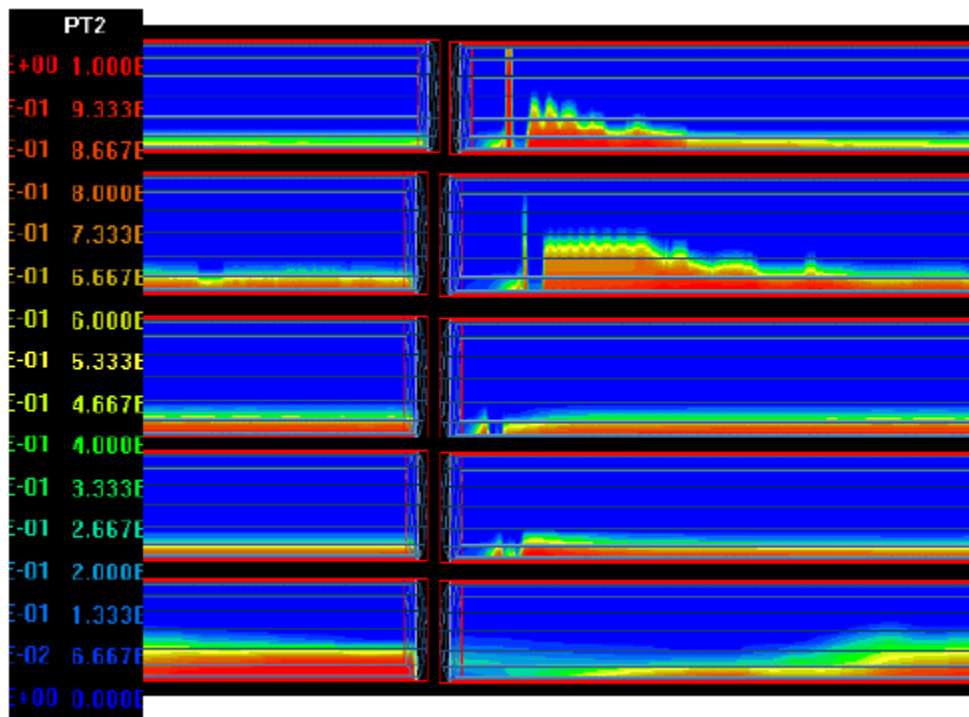


Figure 48. A 220 μm particle concentration along the transfer pipe center line as a function of flow velocity at the entrance, $U = 0.5, 0.8, 1.1, 1.2,$ and 2.5 m/s [1].

Some of the limitations of this model include lack of support to simulate chemical reactions kinetics. It was found to be challenging using Phoenics code to simulate a dynamic sediment bed model that could interact with the turbulent flow field. As particles erode and deposit, the bed surface topography changes and the turbulent flow boundary conditions must continually change to account for the moving boundary. Moreover, the simulations took very long time to

run and the use of parallel computing capabilities to mitigate this problem was not feasible during that time. For these reasons, further modeling efforts were put on hold by the Task Focus Area (TFA) in FY 2002.

PNNL's Paraflow Model: Some of the shortcomings and challenges identified during the previous work of MSU DIAL were addressed by PNNL who developed a multiphase model simulating dynamic sediment bed profiles and suspended solids concentration for multiple particle sizes in a horizontal transfer pipeline. The flow field was modeled using a lattice based algorithm (lattice kinetics). Distribution functions containing information about the pressure tensor streamed to adjacent lattice sites. The summation of these quantities results in equations that were solved for mass, momentum and scalar transport. Turbulent quantities were calculated using a k-epsilon RANS (Reynolds-averaged Navier-Stokes) model. Solids were represented using two different continuum fields. The suspended solids were treated as passive scalars in the flow field, including terms for hindered settling and Brownian diffusion. Normal stresses created by the irreversible collisions of particles during shearing were added to the pressure tensor. The sediment bed was modeled using a phase-field representation where the phase interface was represented by a continuous order parameter profile that moved smoothly through the computational grid. The sediment bed changed through deposition by hindered settling and erosion when the surface shear stress exceeded a specified yield stress [2].

A series of simulations were performed to determine the critical velocity of several suspensions and results were compared with experiments [32]. Simulation began by establishing a steady-state condition with a high flow rate. The flow was then reduced, and if the flow was sufficiently low, a sediment bed develops and grown until an equilibrium state was reached (rate of sedimentation = rate of erosion). Figure 49 shows the simulated sediment bed. The dark region represents sediment bed and the colored planes are the solids concentrations in the suspension.

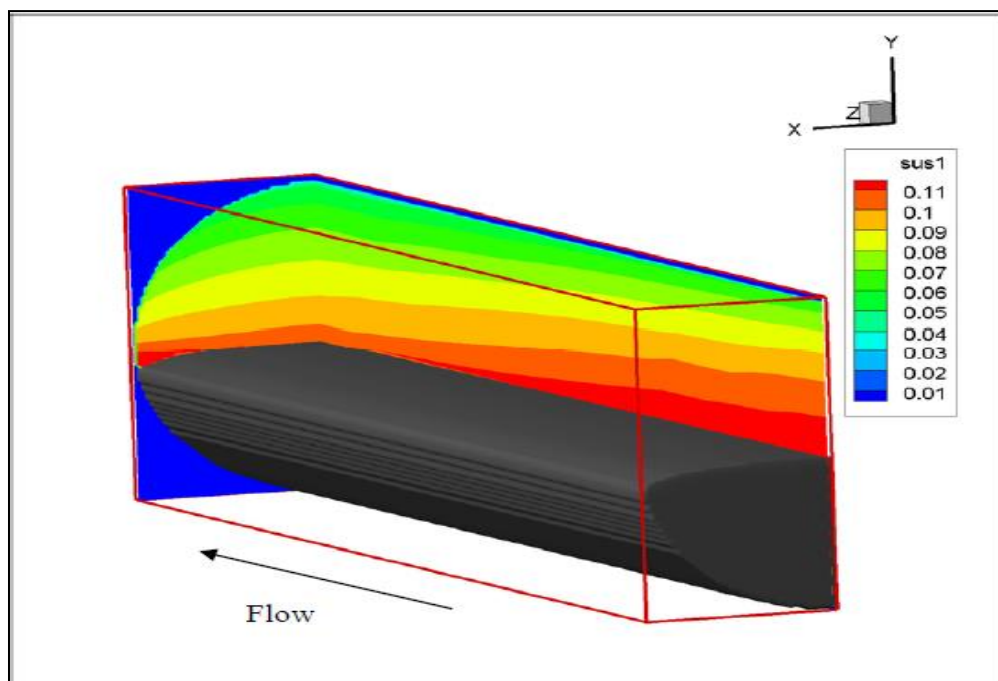


Figure 49. Simulation of sediment bed and solids concentration (vol %) in slurry pipe. Black represents the sediment bed [2].

After a number of simulations ran at different flow conditions, critical velocity was determined by identifying the inflection point in the pressure drop vs. flow rate curve as shown in Figure 50. The dotted line represents the turbulent pressure drop for a pure fluid; the colored circles are the electrical resistance tomograms at different flow rates. The dark blue region represents the sediment bed cross-section. The color in the upper region indicates the suspended solids concentration. At high flow rates, homogenous suspensions were observed. As the flow rate decreased, an increase in solids concentration was observed near the bottom of the tank and a decrease in concentration at the top of the tank. As flow rate decreased further, a moving bed was observed at the bottom of the pipe. At the flow rate just above critical velocity, a stationary bed was formed on the bottom of the pipe. The electrical resistance tomography images shown in Figure 50 that as the flow rate decreases, the sediment bed grows and reduces the flow area.

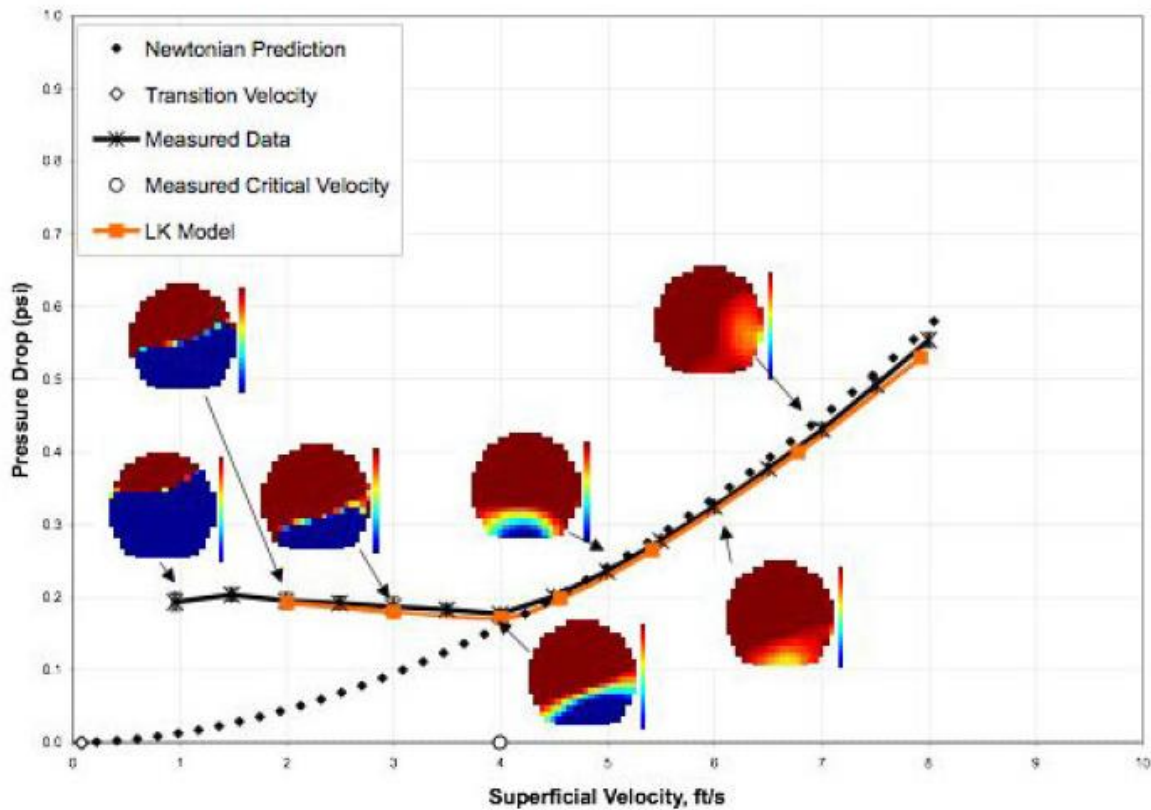


Figure 50. Pressure drop vs. superficial velocity and electrical resistivity tomograms for 100 μm glass beads in water [2].

NUMERICAL SIMULATIONS

Single Phase Simulations

The main goal behind the single phase simulations was to build a baseline model using Comsol Multiphysics that could accurately simulate single phase laminar flow chemistry in a pipe. The baseline model consisted of a reacting flow entering a straight pipe section at laminar velocity. The influence of bends on the reacting flow field was analyzed by adding a 90° elbow and a straight pipe section to the baseline model. The evolution of species was investigated as they were transported through the pipe. The velocity and concentration profiles were plotted to analyze the effect of flow field on the chemical reaction process.

The reacting flow modeling consisted of coupling of the two interfaces:

- Flow Interface-used to simulate laminar flow fields along the pipe length
- Chemical Reactions Interface-used to model chemical reaction between diluted species (A+B→C) transported at the velocities calculated by the flow interface.

Initially the studies were all modeled in two-dimensional (2D) space and numerical results were evaluated for the velocity and concentration profiles. Due to the trade-off between accuracy and longer computing times associated with three-dimensional (3D) modeling, only Case 3 was modeled in 3D space to get a more accurate representation of the reacting flow kinetics (Figure 51).

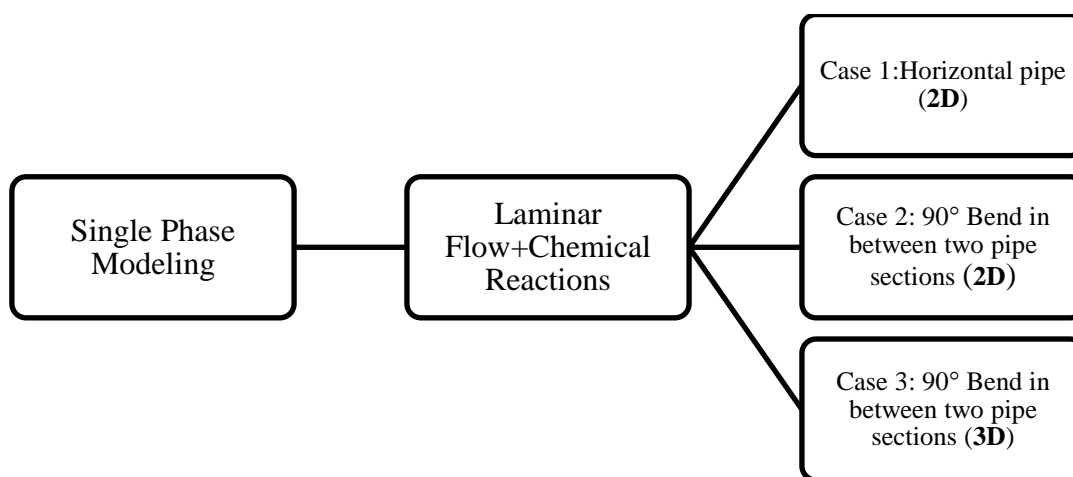


Figure 51. Brief overview of the cases simulated in single phase modeling environment.

Governing Equations

The laminar flow was simulated by using the Navier-Stokes equations of mass conservation (Eq.76) and momentum balance (Eq.77 and Eq.78). For a steady-state, isothermal and incompressible flow, they are:

$$\frac{\partial u}{\partial x} + \frac{\partial v}{\partial y} = 0 \tag{76}$$

$$\rho \left(u \frac{\partial u}{\partial x} + v \frac{\partial u}{\partial y} \right) = \frac{\partial p}{\partial x} + \mu \nabla^2 u + F_x \tag{77}$$

$$\rho \left(u \frac{\partial v}{\partial x} + v \frac{\partial v}{\partial y} \right) = \frac{\partial p}{\partial y} + \mu \nabla^2 v + F_y \quad (78)$$

where, ρ is fluid density (kg/m^3), p is fluid pressure (Pa), F_x and F_y (N/m^3) are force terms accounting for gravity or other body forces, and u , v are velocity components in the x and y directions, respectively.

The Laminar Flow Physics Interface first solves Eqs. (76), (77) and (78) for the velocity vector components u and v within the flow domain. Then, the calculated velocity field is used as a model input for the Transport of Diluted Species Interface to study the evolution of chemical species transported by diffusion and convection mechanisms. The interface assumes that all species present are diluted; that is, that their concentration is small compared to a solvent fluid (water in this case).

The chemical species transport via diffusion and convection is implemented using the transport equation of mass balance (Eq.79)

$$\frac{\partial c}{\partial t} + u \cdot \nabla c = \nabla \cdot (D \nabla c) + R \quad (79)$$

where, c is the concentration of the species (mol/m^3), D denotes the diffusion coefficient (m^2/s), R is a reaction rate expression for the species ($\text{mol}/(\text{m}^3 \cdot \text{s})$) and u is the velocity vector (m/s). The first term on the left-hand side of Eq.(79) corresponds to the consumption of the species. The second term accounts for the convective transport due to a velocity field u obtained from the solution of Laminar Interface. On the right-hand side of the mass balance equation (Eq.79), the first term describes the diffusion transport, accounting for interaction between the dilute species and the solvent. Finally, the second term on the right-hand side of Eq.(79) represents the reaction rate that determines the rate at which species are consumed or produced.

Boundary Conditions

For the fluid flow interface, at the inlet, a velocity field in x-direction was specified. At the outlet, the pressure was set to zero. At the wall, a no-slip boundary condition was selected assuming that the fluid at the wall was not moving. For the chemical reactions interface, the inlet concentration of species A and species B were specified using the inflow node and it was assumed that the species were transported by convection and diffusion. A no flux boundary condition for all species was applied at the walls representing no mass flowing in and out of these boundaries. A reaction node was added specifying the reaction rate (in $\text{mol}/\text{m}^3 \cdot \text{s}$) to account for the consumption or production of species. At the outlet, it was assumed that convection was the dominating transport mechanism by which the species were transported out of the model domain by a fluid flow, and therefore the diffusive transport was ignored.

Case 1: Reacting Flow through a straight pipe section (2D)

Liquid water was introduced into a straight pipe flowing at a velocity of 0.03 m/s. The liquid density was 1000 kg/m^3 and the viscosity was $0.001 \text{ Pa}\cdot\text{s}$. The inlet pipe width was 0.0762 m resulting in a Reynolds number of 2286. The pipe length was 1.828 m (Figure 52). A synthesis reaction ($A+B \rightarrow C$) was simulated within water. Two dilute species A and B with concentrations of 1 mol/m^3 and 0.5 mol/m^3 were introduced at the pipe inlet. The evolution of the reactant and product species for a laminar flow was tracked over the pipe length.

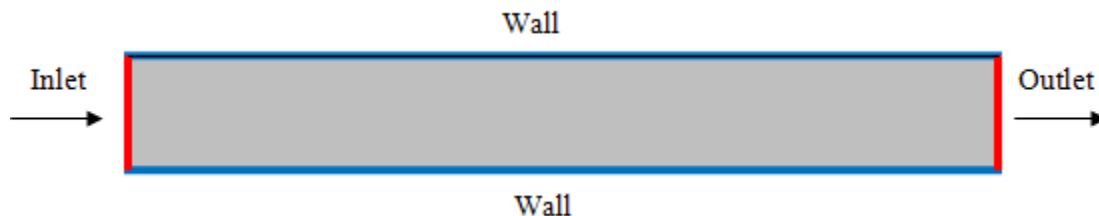


Figure 52. Depiction of the model geometry.

The model domain for the fluid flow interface was meshed using mapped elements (Figure 53(a)) whereas the model domain while using the chemical reactions interface was meshed using triangular elements (Figure 53(b)).

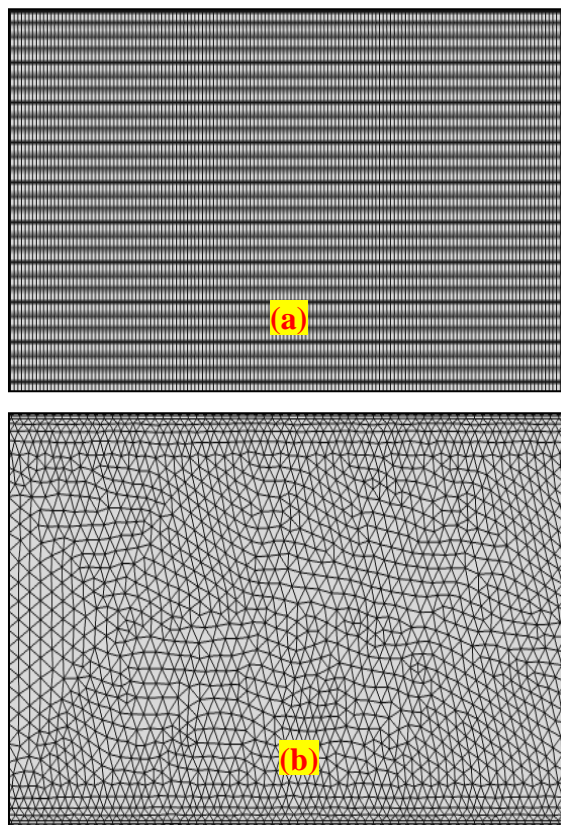


Figure 53. Type of mesh used for the model domain a) mapped and b) triangular.

The laminar flow interface was solved first and its calculated velocity field was used as a model input for the Transport of Diluted Species Interface to study the conversion of chemical species transported by diffusion and convection mechanisms.

Case 2: Reacting Flow through a 90° Elbow (2D)

The model geometry for Case 2 consisted of two straight pipe sections and a 90° elbow in between. The main objective of this study was to evaluate the effects of bends on the reacting flow kinetics. Liquid water was introduced at the inlet of a straight pipe flowing at a velocity of 0.01 m/s, passing through the elbow and then exiting at the outlet. The liquid density was 1000 kg/m³ and the viscosity was 0.001 Pa·s. The inlet pipe width was 0.0762 m resulting in a

Reynolds number of 762. The pipe length for the inlet and outlet pipe section was 1.828 m each (Figure 54). Two dilute species A and B with concentrations of 1 mol/m^3 and 0.5 mol/m^3 were introduced at the pipe inlet. The reaction rate of $0.01 \text{ mol/m}^3 \cdot \text{s}$ was specified. The evolution of the reactant and product species as they are transported over the model geometry consisting of two pipe sections and a 90° elbow was investigated.

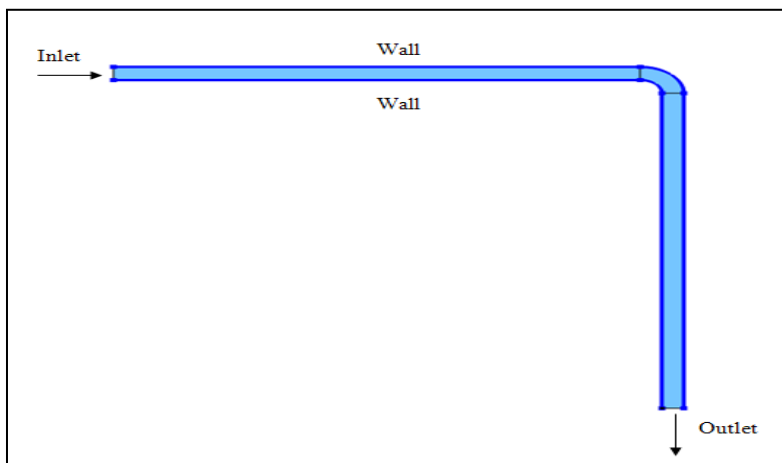


Figure 54. Model geometry for case 2.

The model domain for both the flow and chemical species interface was partitioned into triangular shape sub-domains using 5885 elements. Figure 55 shows the zoomed in view of the meshed elbow. A finer mesh was applied at the walls of the pipe than for the rest of the domain.

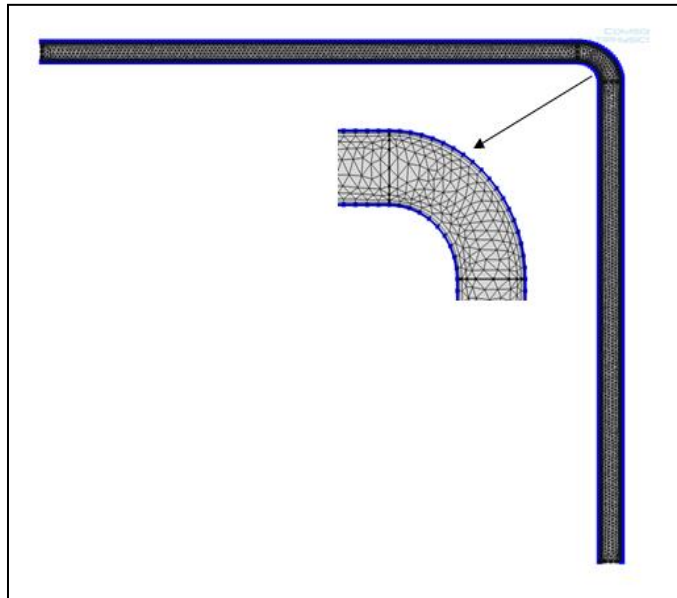


Figure 55. Meshed geometry for case 2.

Case 3: Reacting Flow through a 90° Elbow (3D)

The case 3 consisted of expanding the previous 2D model into a 3D analysis with the aim of getting a more accurate representation of the reacting flow kinetics. The model geometry consisted of two pipe sections and a 90° elbow in between. The pipe diameter was 0.0762 m and

the pipe length of the two sections was 1.828 m (Figure 56) each. The working fluid was water with a density of 998.2 kg/m^3 and a viscosity of $0.001 \text{ Pa}\cdot\text{s}$. The water entered the horizontal pipe section (inlet at the x-axis) at a velocity of 0.01 m/s , passed through the elbow, and exited the second pipe section (outlet at the y-axis). The fluid flow interface was coupled with the chemical reactions interface to simulate a simple synthesis ($A + B \rightarrow C$) reaction within the water flowing at the laminar velocity.

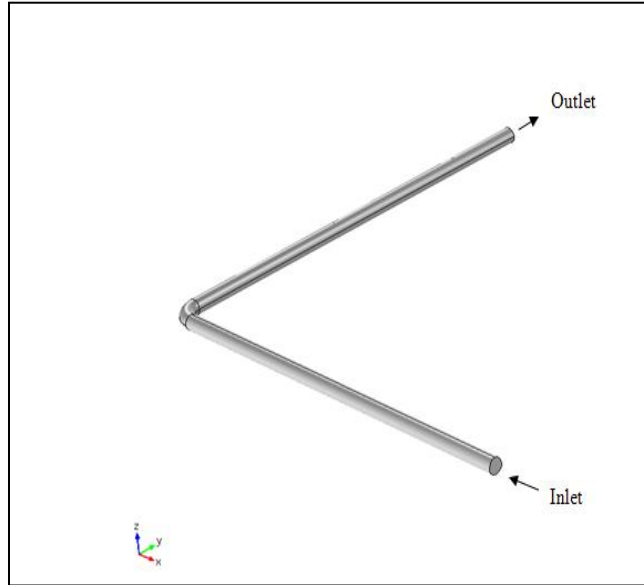


Figure 56. 3D model geometry for case 3.

The model domain was partitioned into sub-domains with the triangular mesh type and the element size selected was normal. The complete mesh consisted of 46233 domain elements and took about 2 hours to solve the coupled interfaces (Figure 57). A close up view of the elbow shows the structured triangular mesh across the entire elbow.

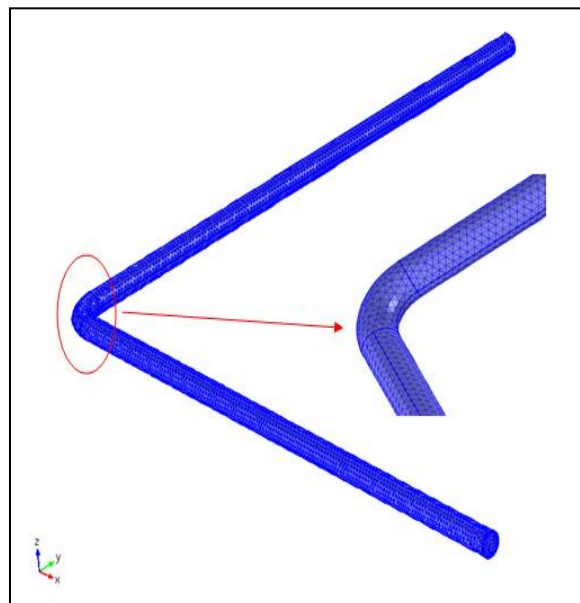


Figure 57. Mesh for the 3D model built for case 3.

Multi-Phase Simulations

The multi-phase simulations focused on modeling the two-phase flow of mixtures in a pipe with a laminar velocity. The two phases consisted of a solid phase (dispersed) and a liquid phase (continuous). A mixture interface based on the Euler-Euler approach was used to solve the multi-phase model. The dispersed phase volume fraction and the continuous phase volume fraction were computed at the end of the simulations.

The two phase flow at laminar velocity was modeled using the mixture model laminar interface. The mixture consisted of 10% volume fraction of clay dispersed in water (Figure 58). The mixture was introduced at the inlet of a horizontal pipe section flowing at a velocity of 0.01 m/s. The pipe diameter was 0.0762 m and the pipe length was 1.828 m. Water was selected as the continuous phase and the solids as the dispersed phase. The density of water used was 1000 kg/m³ and a viscosity of 0.001 Pa·s. The density of clay used was 1500 kg/m³ and the particle size of 0.001 m.

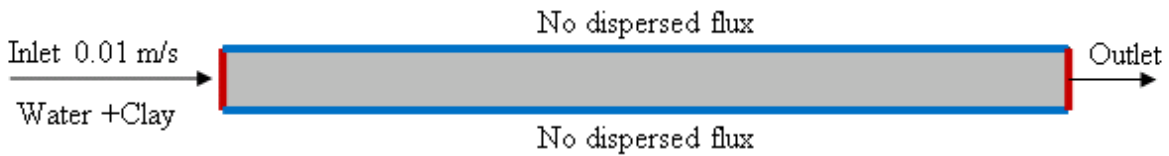


Figure 58. Model geometry of the multi-phase model.

Governing Equations

The equations governing the mixture model are the momentum and continuity equations and a volume fraction solution for each phase analyzed. Assumptions made in the simulations include constant density for each phase, constant pressure field and the velocity between the two phases can be ascertained from a balance of pressure, gravity, and viscous drag.

The momentum equation for the mixture is given by

$$\rho u_t + \rho(u \cdot \nabla)u = -\nabla p - \nabla \cdot \tau_{Gm} + \rho g + F - \nabla \cdot \left[\rho c_d (1 - c_d) \left(u_{slip} - \frac{D_{md}}{(1 - c_d)\phi_d} \nabla \phi_d \right) \left(u_{slip} - \frac{D_{md}}{(1 - c_d)\phi_d} \nabla \phi_d \right) \right] \quad (80)$$

where, u denotes mixture velocity (m/s), ρ is the mixture density (kg/m³), p is the pressure (Pa), c_d is the mass fraction of the dispersed phase (kg/kg), u_{slip} is the relative velocity between the two phases (m/s), τ_{Gm} is the sum of viscous and turbulent stress (kg/(m·s²)), g is the gravity vector (m/s²), and F is the additional volume forces (N/m³).

The mixture velocity u is defined as

$$u = \frac{\phi_c \rho_c u_c + \phi_d \rho_d u_d}{\rho} \quad (81)$$

Here, c and d denote the volume fractions of the continuous phase and the dispersed phase (m³/m³), respectively, u_c the continuous phase velocity (m/s), u_d the dispersed phase velocity

(m/s), ρ_c the continuous phase density (kg/m^3), ρ_d the dispersed phase density (kg/m^3), and ρ the mixture density (kg/m^3).

The mixture density ρ is given by

$$\rho = \phi_c \rho_c + \phi_d \rho_d \quad (82)$$

where ρ_c and ρ_d (kg/m^3) are the densities of each of the two phases.

The mass fraction of the dispersed phase c_d is given by

$$c_d = \frac{\phi_d \rho_d}{\rho} \quad (83)$$

The relative velocity between two phases is defined by

$$\mathbf{u}_d - \mathbf{u}_c = \mathbf{u}_{cd} = \mathbf{u}_{slip} - \frac{D_{md}}{(1-c_d)\phi_d} \nabla \phi_d \quad (84)$$

Here, u_{slip} (m/s) denotes the relative velocity between the two phases, D_{md} is a turbulent dispersion coefficient (m^2/s) accounting for extra diffusion due to turbulent eddies. In this case D_{md} is set to zero as no turbulent model was used.

The sum of viscous and turbulent stress is

$$\tau_{Gm} = (\mu + \mu_T)[\nabla \mathbf{u} + \nabla \mathbf{u}^T] \quad (85)$$

where, μ (Pa·s) is the mixture viscosity and μ_T (Pa·s) the turbulent viscosity.

The transport equation for Φ_d , the dispersed phase volume fraction, is

$$\frac{\partial}{\partial t}(\phi_d \rho_d) + \nabla \cdot (\phi_d \rho_d \mathbf{u}_d) = -m_{dc} \quad (86)$$

where, m_{dc} ($\text{kg}/(\text{m}^3 \cdot \text{s})$) is the mass transfer rate from dispersed to continuous phase and u_d (m/s) is the dispersed phase velocity.

With a constant density for the dispersed phase, the equation becomes

$$\frac{\partial}{\partial t}(\phi_d) + \nabla \cdot (\phi_d \mathbf{u}_d) = -\frac{m_{dc}}{\rho_d} \quad (87)$$

The continuous phase volume fraction ϕ_c is

$$\phi_c = 1 - \phi_d \quad (88)$$

The continuity equation for the mixture is

$$\rho_t + \nabla \cdot (\rho \mathbf{u}) = 0 \quad (89)$$

The Mixture Model interfaces assumes that the densities of each phase, ρ_c and ρ_d , are constant, and therefore uses the following alternative form of the continuity equation of the mixture

$$(\rho_c - \rho_d) \left[\nabla \cdot (1 - c_d) \mathbf{u}_{slip} - D_{md} \nabla \phi_d \right] + \rho_c (\nabla \cdot \mathbf{u}) = 0 \quad (90)$$

Mesh

The model domain for the mixture interface was partitioned into sub-domains using a triangular mesh. The element size used was a Finer size which generated 32676 elements over the entire domain (Figure 59).

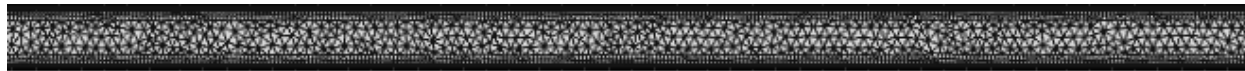


Figure 59. Meshed model geometry for the mixture interface.

Boundary Conditions

A mixture velocity of 0.01 m/s and a dispersed phase concentration of 0.1 (10%) was specified using the inlet node. The outlet was set to zero pressure, no viscous stress and the dispersed phase flow exited the pipe at mixture velocity. A no slip boundary condition was selected at the walls assuming that the fluid at the wall was stationary and no flux was entering the walls. A gravity node was added to account for gravity force in the negative y-direction over the entire domain.

RESULTS & DISCUSSION

Single Phase Modeling Simulations

Case 1: Reacting Flow through a Straight Pipe Section (2D)

Mesh Optimization

A mesh optimization study was carried out solely for the laminar flow interface to determine the optimal mesh that gave the best velocity profile with the least computing time. Four mesh sizes were simulated for this study: Extremely coarse, coarse, fine and extra fine. The mesh optimization study parameters are listed in Table 2.

Table 2. Mesh Optimization Study Parameters for Laminar Flow Interface Simulations

Mesh Size	No. of Domain Elements	Solution Time (s)
Extremely Coarse	3380	8
Coarse	17584	25
Fine	50812	67
Extra Fine	272390	854

The extremely coarse mesh took approximately 8 seconds to solve the model; however, the velocity profile computed was not smooth. The coarse mesh took 3 times longer to solve the flow physics but some unevenness in the velocity profile was still apparent as seen in Figure 60.

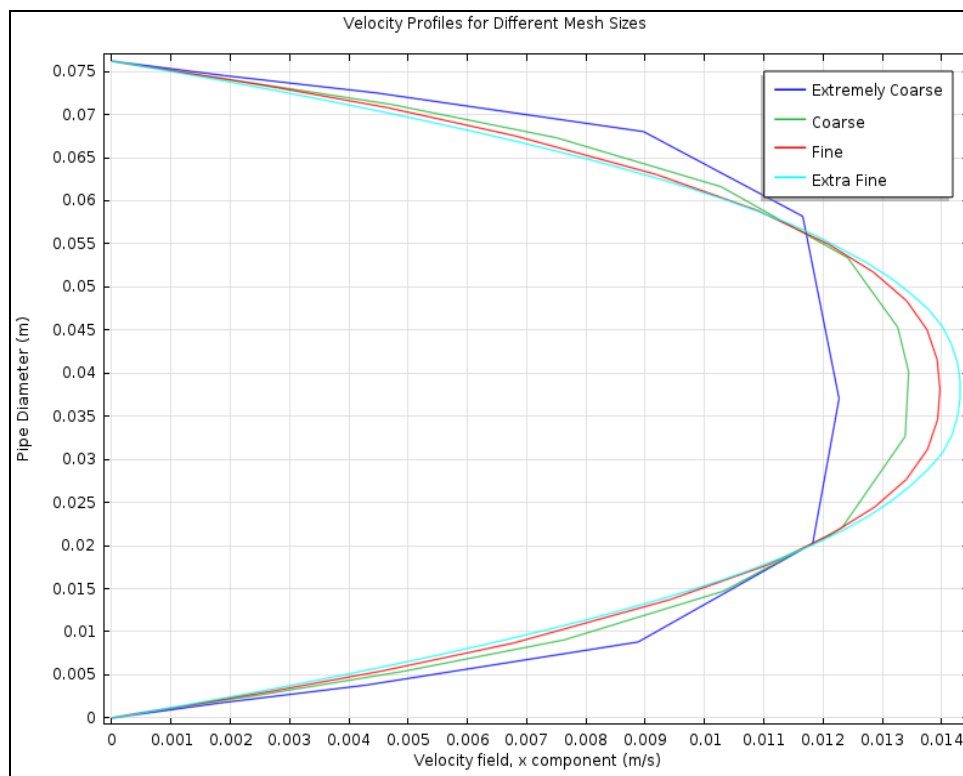


Figure 60. Comparison of velocity profiles for mesh optimization study.

The fine and extra fine mesh profiles generated a smooth parabola; the profiles generated were very similar, but the fine mesh took 67 seconds to solve the flow physics compared to the 854 seconds for the extra fine mesh. Hence, looking at the velocity profile and the solution times, it was determined that the fine mesh was the best choice balancing the solution accuracy and computing times for the subsequent cases 2 and 3.

Velocity profile

The velocity profile of the flow was tracked from the time the flow entered the pipe to when it became fully-developed. This was done by partitioning the model geometry into five sections along the pipe length and then plotting the velocity profile of those sections (Figure 61).



Figure 61. Domain partitioning to obtain section velocity profiles.

The velocity profiles at the various sections along the pipe length are shown in Figure 62. It was observed that the flow became fully developed once it passed the entrance region. The velocity profile of a fully developed flow resembles a parabola. The maximum velocity was observed near the center of the pipe and the minimum velocity was found near the walls of the pipe.

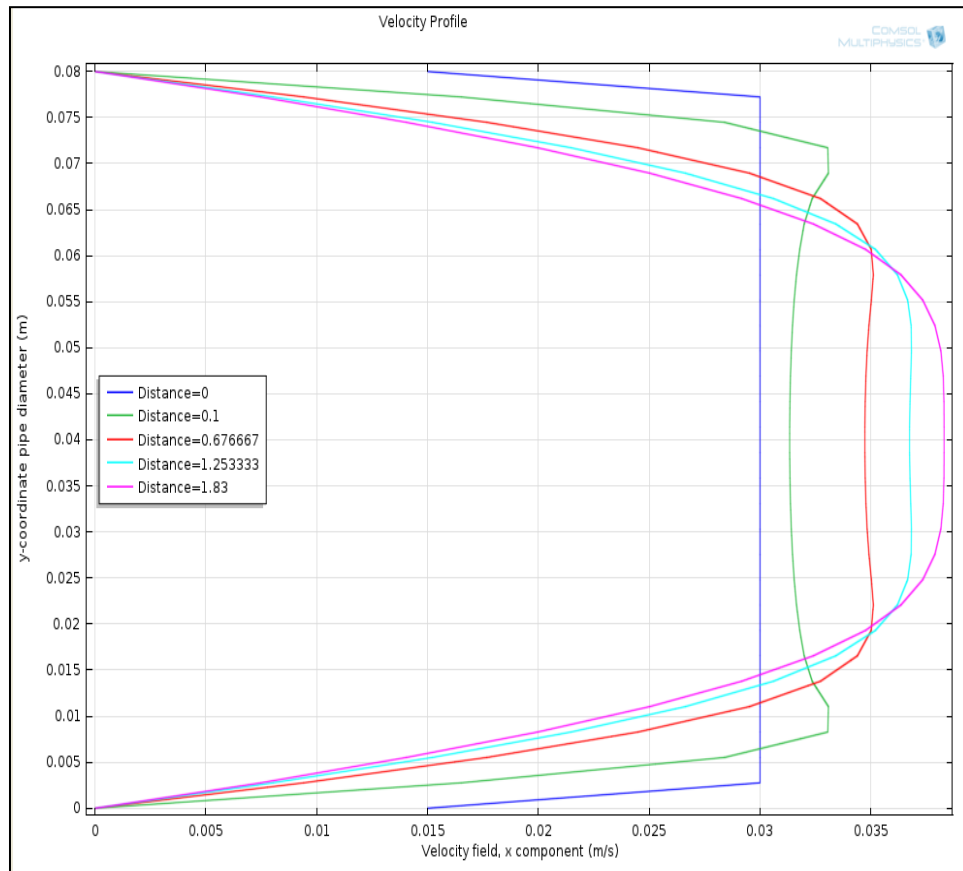


Figure 62. Velocity profiles at different sections of the pipe.

The red color in Figure 63 represents the zone where the maximum velocity was observed and the blue color represents the area of low velocity. The black arrows are the velocity vectors and reflected the changes in the velocity profile shape from the inlet to the outlet of the pipe. Initially the profile was rectangular (as seen near the inlet section of the pipe) and as the flow transitioned to fully developed, the profile became parabolic.

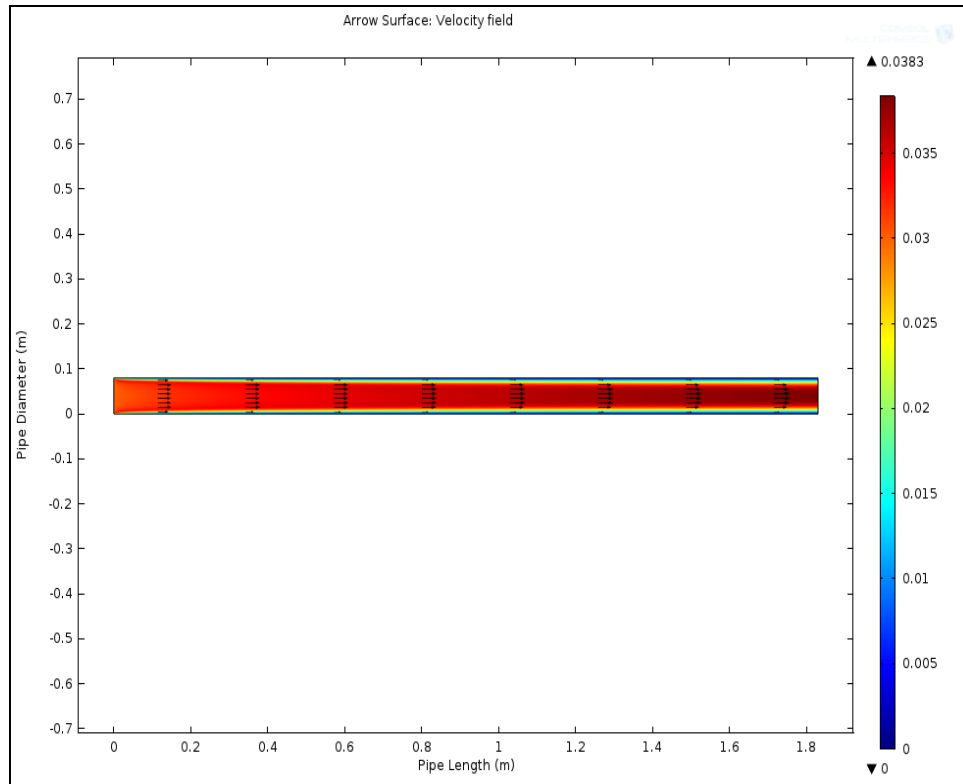


Figure 63. Velocity vectors profiling the shape of the flow along the pipe length.

The laminar flow was characterized by straight streamlines as seen in Figure 64. The fluid flowed very smoothly in the pipe in an organized fashion and no intermixing of the velocity streams was observed. The plotted streamlines were color coded with higher velocity represented by the color red and the lower velocity represented by color blue.

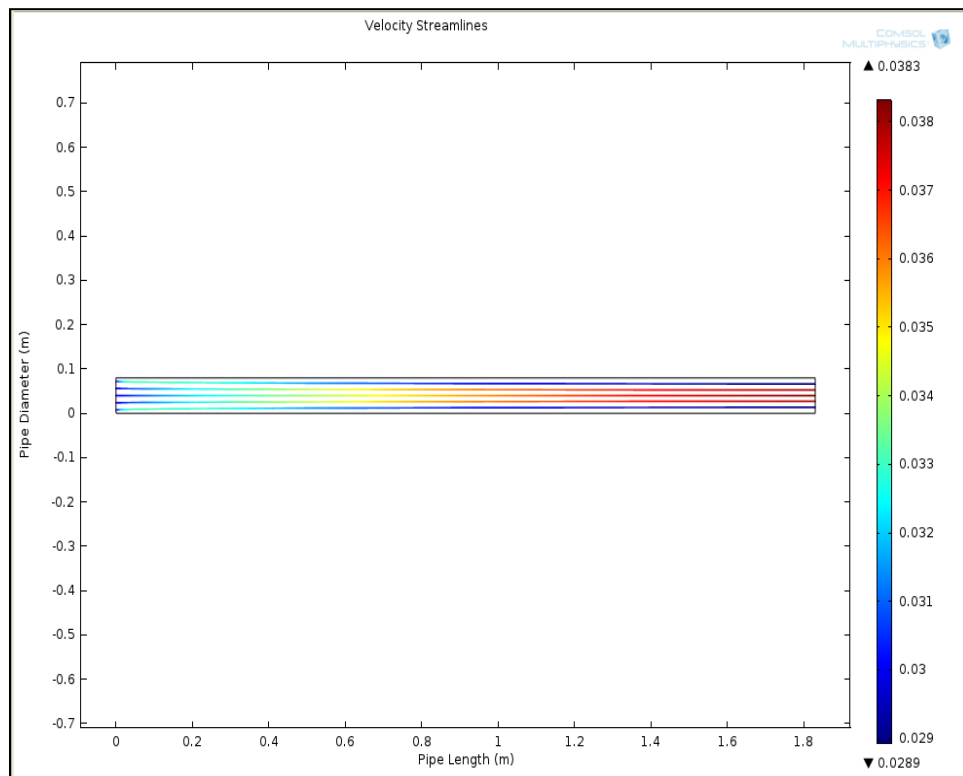


Figure 64. Velocity streamlines along the pipe length.

Concentration profile

The concentration snapshots showed that since in laminar flow, the highest velocity was observed near the center of the pipe, the residence time for the reactants was seen lower in that area. For example in the case of the reactant A, its concentration decreased more at the center as the reactant got consumed than at the wall area. Figure 65 shows the concentration snapshots for the reactant species A at time intervals of 0, 20, 40 and 60 seconds as it was converted.

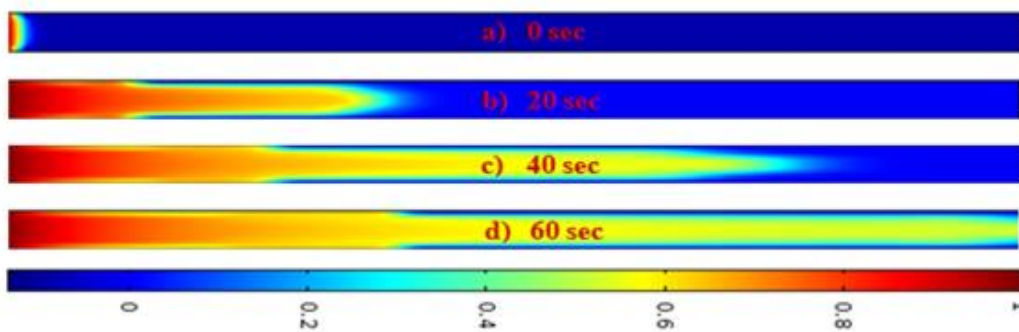


Figure 65. Conversion of reactant species A at a) 0 sec, b) 20 sec, c) 40 sec and d) 60 sec.

The reactants interacted with each other in the high velocity zone at the center area of the pipe and started the reaction process. Hence, the formed product species C had a higher concentration profile near the center of the pipe than at the walls. Figure 67 shows the concentration snapshots for the product species C at time intervals of 0, 20, 40 and 60 seconds as it formed.

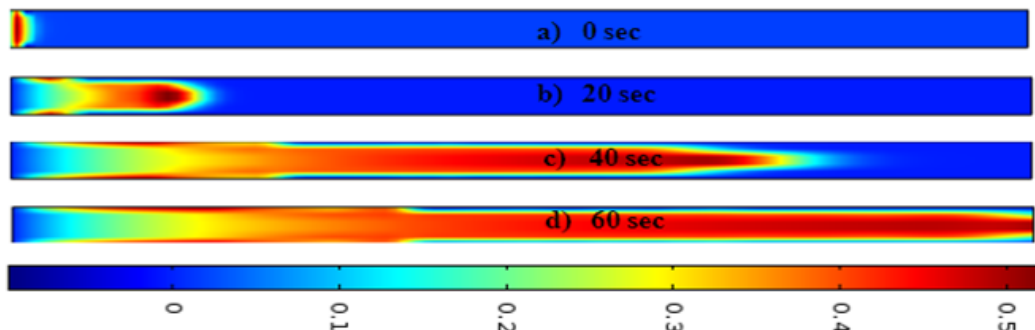


Figure 66. Production of species C at a) 0 sec, b) 20 sec, c) 40 sec and d) 60 sec.

The concentration profile for the three species as they undergo the chemical reaction is shown in Figure 67. The concentration of species A decreased from 1 mol/m³ to 0.5 mol/m³ and the concentration of species B decreased from 0.5 mol/m³ to 0 mol/m³ as they got consumed. At the same time, the product species concentration increased from 0 mol/m³ to around 0.48 mol/m³ as it was produced.

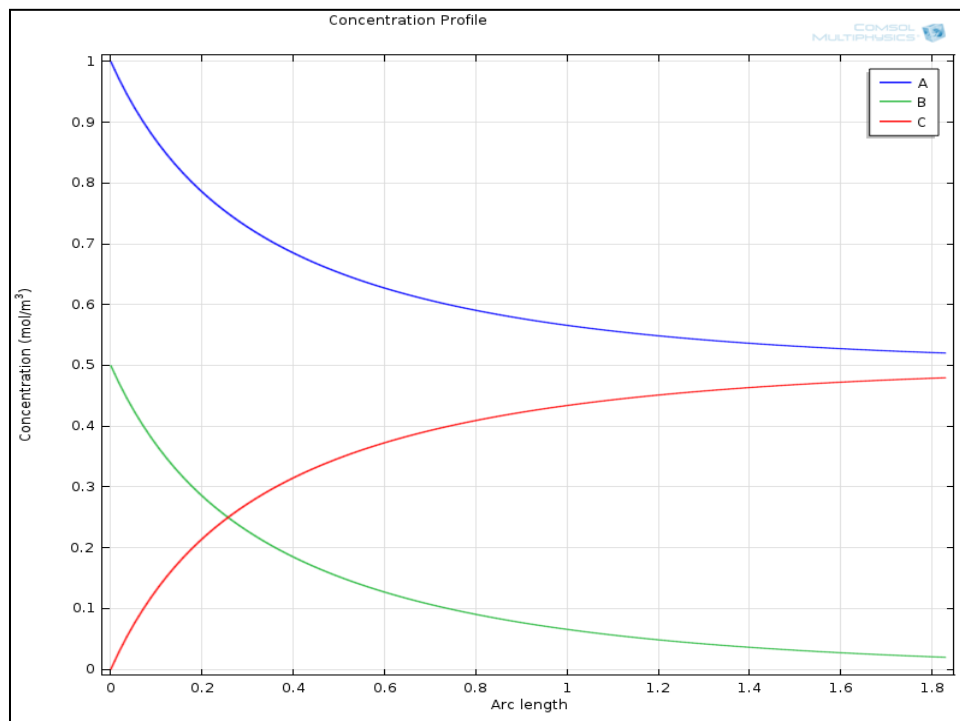


Figure 67. Concentration profile of the three species undergoing chemical reaction.

Case 2: Reacting Flow through a 90°Elbow (2D)

Velocity Profile

The velocity streamlines were plotted in Figure 68 to study the velocity distribution across the model geometry. A fully developed flow started forming right before the elbow as characterized by the high velocity areas forming near the center of the pipe (red color). As the flow entered the elbow, the direction of the flow was displaced from the center section to the inner wall of the elbow. The flow first impacts the outer wall of the elbow resulting in a decreased velocity (blue color) at that section. From the outer wall, the flow rebounded to the inner wall of the elbow at a higher velocity and then exited the elbow transitioning again to a fully developed flow at the downstream section of the pipe.

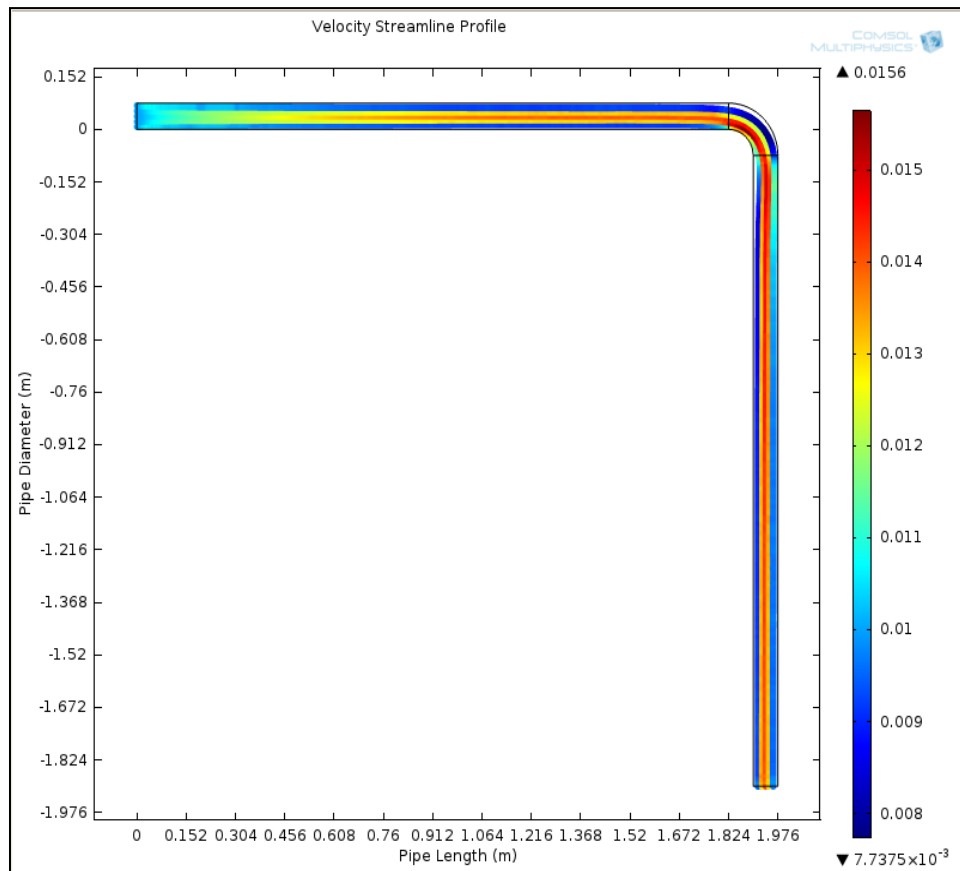


Figure 68. Velocity streamlines plotted for case 2 across the geometry.

Figure 69 shows the velocity profile as the flow travelled from the inlet, passed through the elbow section and exited at the outlet. The flow became fully developed right before the elbow. As the flow entered the elbow, a slight shift in the velocity from the center of the elbow to the inner wall of the elbow was observed. This was the result of the elbow induced disturbance. As the flow exited the elbow, it transitioned back to a fully developed flow.

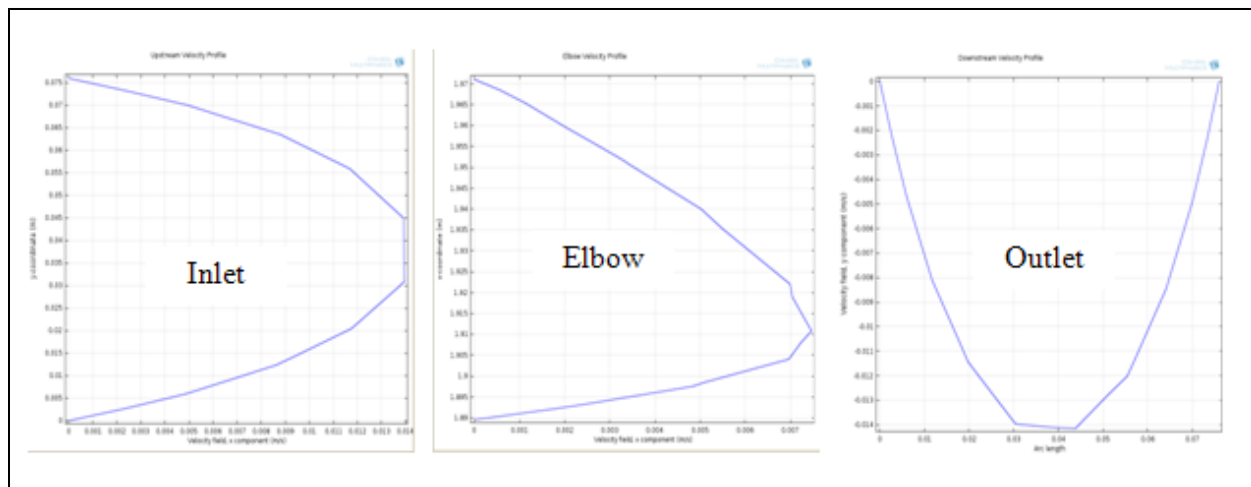


Figure 69. Velocity profiles plot as the flow travels through the inlet, elbow and outlet (x-axis is the velocity (m/s) and y-axis is the pipe diameter (m)).

Concentration Profile

The transition of the species from the reactants to a product as they underwent a chemical reaction is shown in Figure 70. The concentration of product species was found to be the highest at the inner wall of the elbow and near the center of the pipe as the flow moved towards the outlet. This was due to the higher velocity observed in those areas. The residence time of the reactants was lower at the higher velocity areas resulting in the production of species C in those sections. This was characterized by an increased concentration of product species C (red color) observed at the inner wall of the elbow and thereafter. Figure 70 shows the concentration snapshots of the product species C being formed at time intervals of 0, 60, 120, 200, 300 and 400 seconds, respectively.

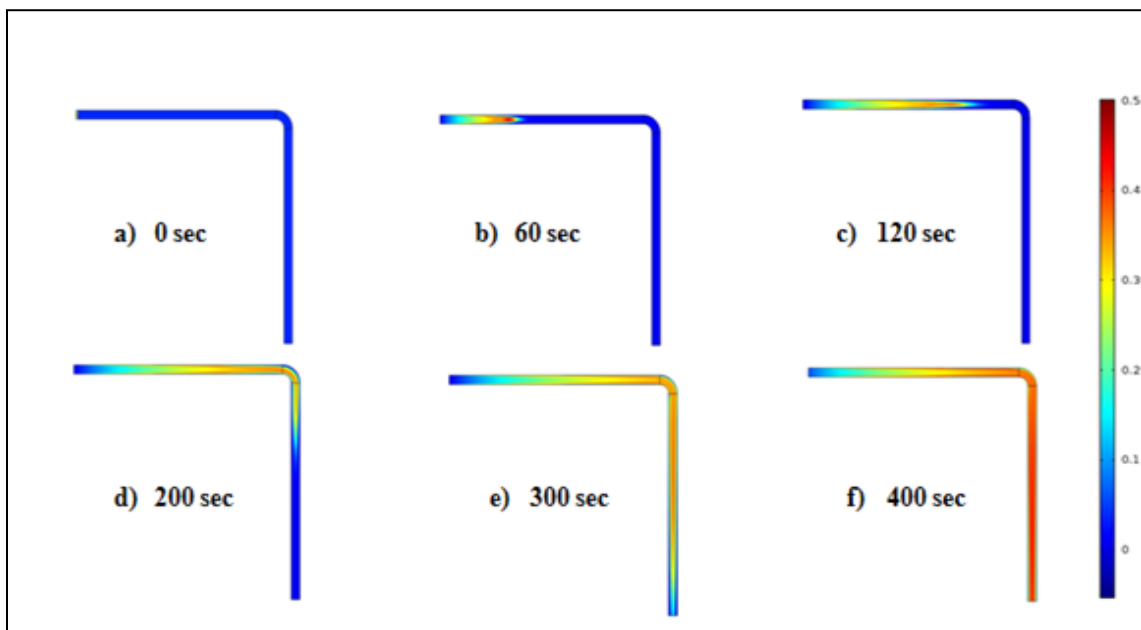


Figure 70. Concentration snapshots of species C at a) 0 sec, b) 60 sec, c) 120 sec, d) 200 sec, e) 300 sec and f) 400 sec.

Case 3: Reacting Flow through a 90° Elbow (3D)

Velocity Profile

As the fluid travelled through the pipe and entered the elbow entrance, the high velocity streamlines deviated from the center of the pipe to the inner wall due to the elbow induced flow disturbance. A maximum velocity of 0.016 m/s was observed at the inner elbow wall (red streamlines). The flow exited the elbow impacting the outer pipe wall at a higher velocity than at the center of the pipe and eventually the flow transitioned into a fully developed flow as it travelled towards the outlet (Figure 71).

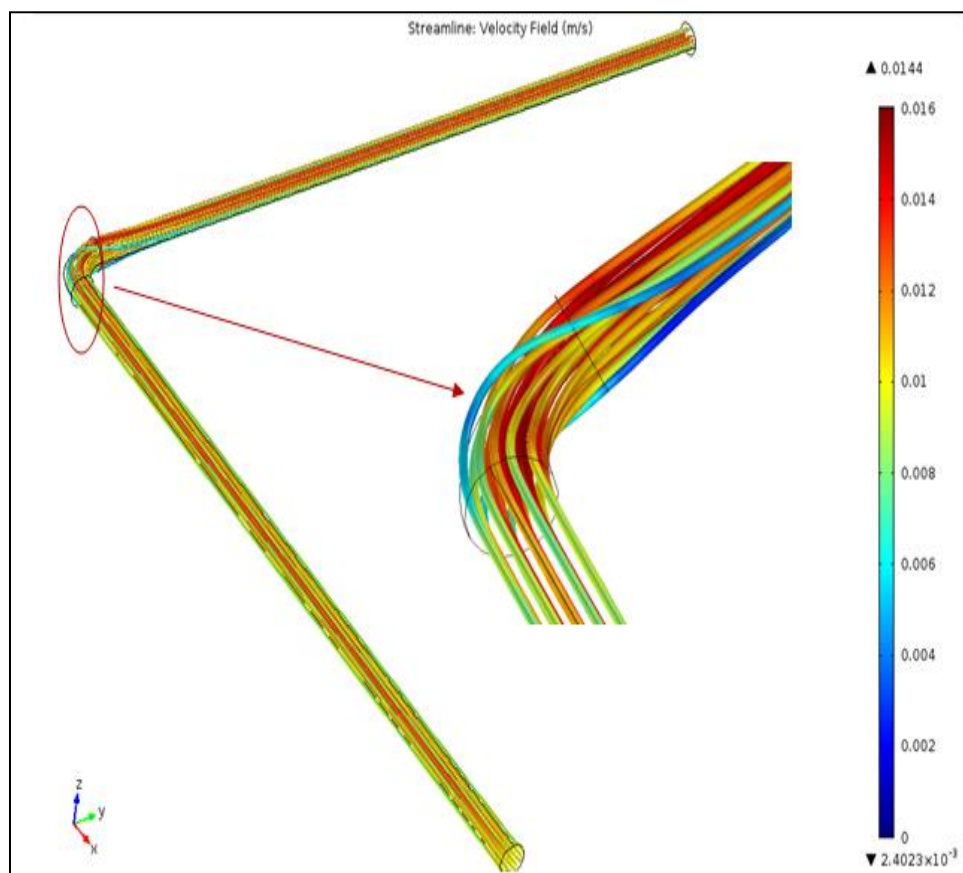


Figure 71. Velocity streamlines for the 3D laminar flow model.

The 2D model showed a maximum velocity of 0.0156 m/s (Figure 68) compared to the 0.016 m/s (Figure 71) computed by the 3D model. Hence it was concluded that the 2D model was sufficient to get an accurate representation of the laminar reacting flow kinetics and a 3D analysis was not needed.

Concentration Profile

The evolution of the species during the reaction process was studied from the concentration surface profile. The reactant species A and B were introduced at the pipe inlet and transported in the pipe via convection and diffusion mechanisms. Since the laminar flow is characterized by the maximum velocity occurring near the center of the pipe, the residence time of the reactants

was lower at the center of the pipe. The reaction spurred as they came in contact with one another resulting in the formation of the product species C. The concentration of species C was found to be the highest in the areas of higher velocity such as the elbow inner wall and the pipe outer wall after the reacting flow exited the elbow. The red color represents the maximum concentration of 0.5 mol/m^3 for species C. The concentration snapshots of species C at 0, 100, 200, 300, 400 and 500 sec is shown in Figure 72.

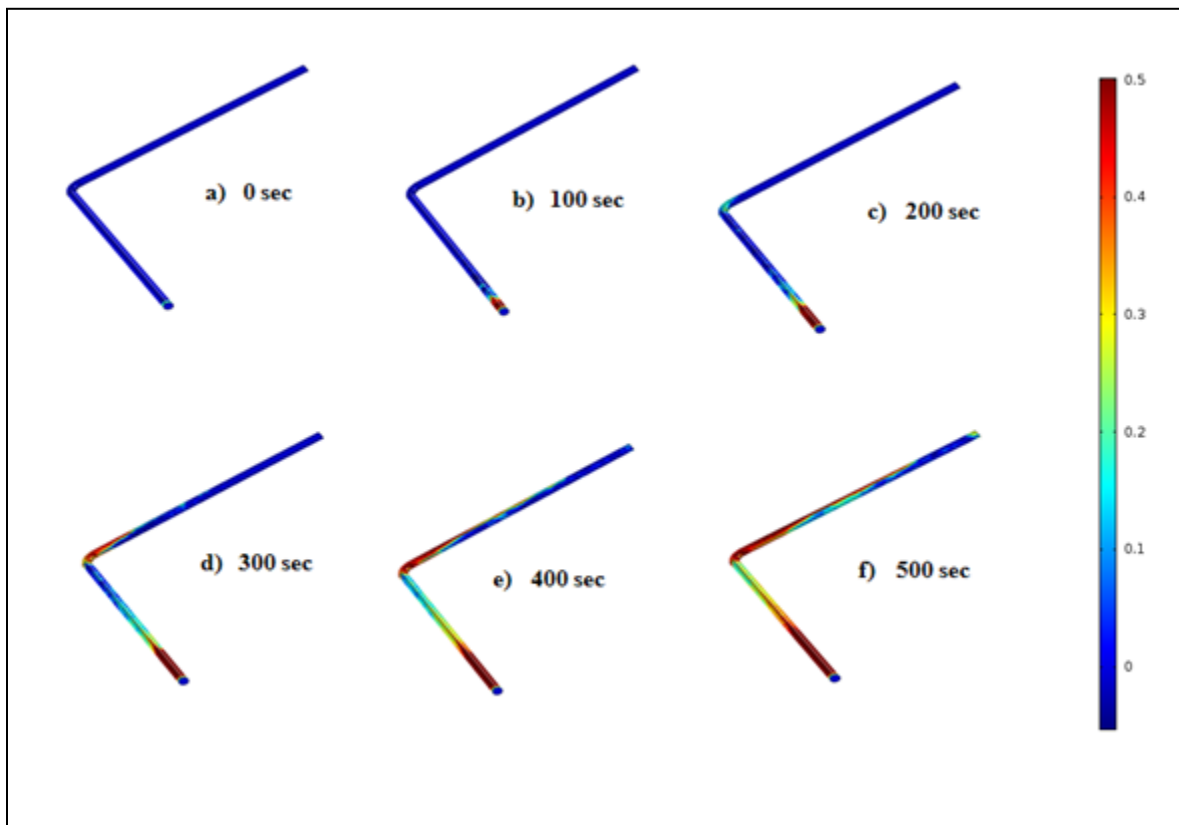


Figure 72. Concentration of species C at a) 0 sec, b) 100 sec, c) 200 sec, d) 300 sec, e) 400 sec and f) 500 sec.

Multi-Phase Simulations

Solids Settling

In Figure 73 the continuous phase (water) is represented by the blue color and the dispersed phase (clay) is represented by the pink color. As the mixture of solids and liquid was transported along the pipe length, the heavier solids started to settle at the bottom of the pipe and resulted into a gradual formation of a bed of solids. Figure 73 shows the surface dispersed volume fractions at 0, 10, 20, 30, 60, 80 and 90 seconds, respectively.

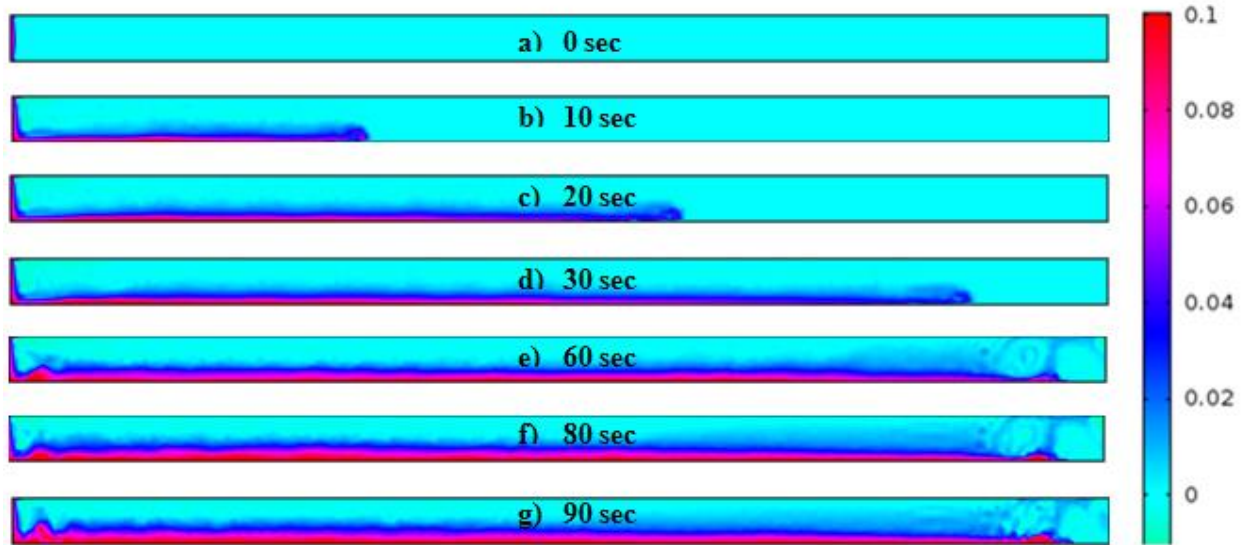


Figure 73. Solids settling at the bottom of the pipe at a) 0 sec, b) 10 sec, c) 20 sec, d) 30 sec, e) 60 sec, f) 80 sec and g) 90 sec.

The volume fraction for the dispersed phase (clay) increased from 0 to 0.15 as the bed of the solids started forming at the bottom section of the pipe. A gradual increase in the volume fraction along the pipe length was observed as plotted in Figure 74.

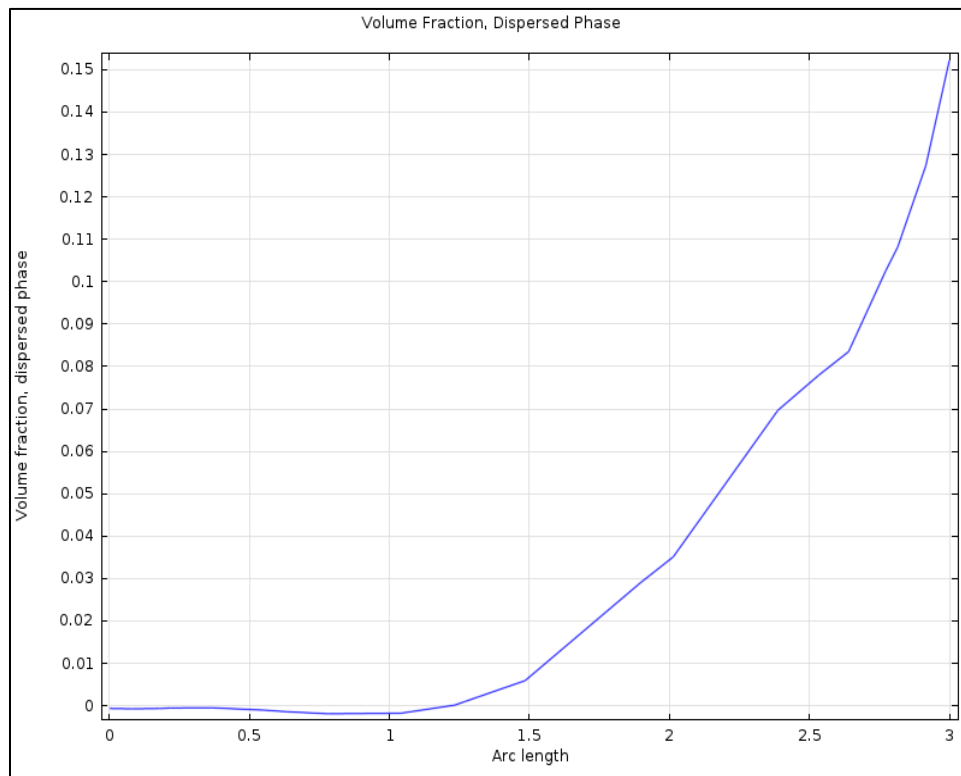


Figure 74. Volume fraction of dispersed phase as a function of arc length.

On the other hand, the volume fraction of continuous phase (water) decreased from 1 to 0.85. This was the result of the water being displaced by the solids as shown in Figure 75.

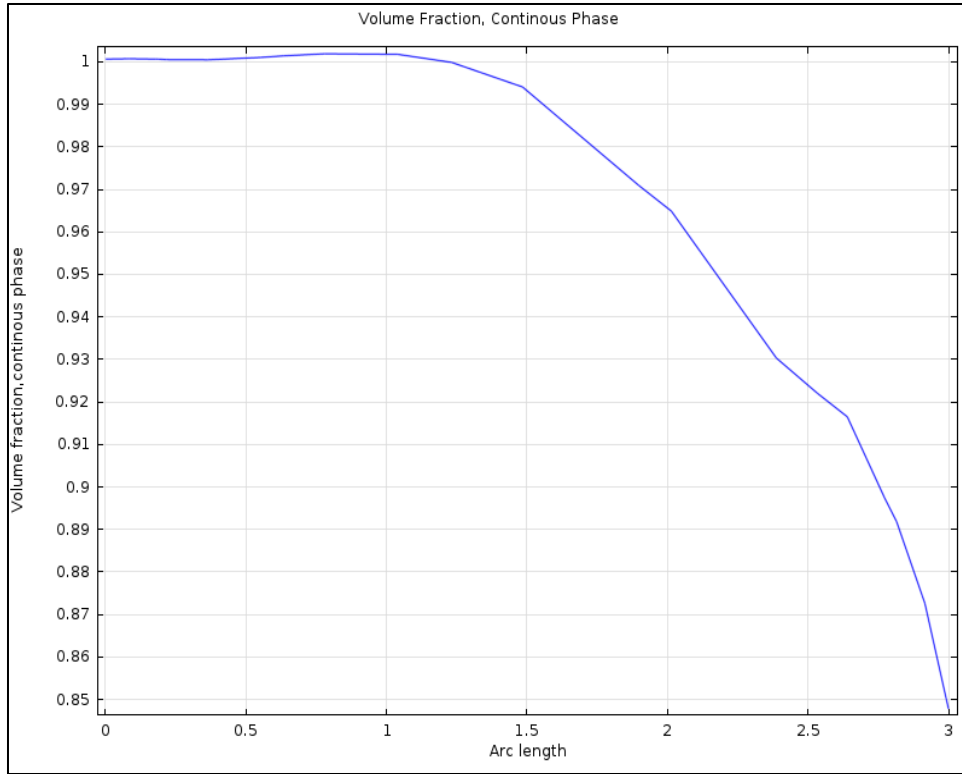


Figure 75. Volume fraction of continuous phase as a function of arc length.

CONCLUSIONS

The overall project conclusion was that Comsol Multiphysics 4.3bTM could efficiently model the reacting flow kinetics. The analysis of reacting flow kinetics included understanding how the flow field affected the chemical reaction process. For the single phase models, the maximum velocity was observed near the center of the pipe and the minimum velocity at the walls of the pipe as the fluid transitioned into a fully developed flow. The addition of an elbow displaced the higher velocity fluid from the center section to the inner wall of the elbow. Hence, the residence time of the reactants was found to be lower at the center of the pipe and at the inner elbow (high velocity areas) as they got converted into products. This was characterized by an increased concentration of product species observed in those sections. The comparison of the 2D and 3D reacting flow analysis for the case where flow was simulated across a 90° elbow showed similar results. The 2D model showed a maximum velocity of 0.0156 m/s (Figure 68) compared to the 0.016 m/s (Figure 71) computed by the 3D model. The concentration snapshots showed that the computed maximum concentration of product species was 0.45 mol/m³ for both the 2D and 3D analysis. Hence, the 2D model was sufficient to get an accurate representation of the laminar reacting flow kinetics and a 3D study was not needed for other cases.

The mixture model was only simulated for the flow interface. The coupling of the flow interface to the chemical reactions was not done during this performance period. The mixture simulations consisted of a two phase flow (10% vol. fraction of solids dispersed in water) across a horizontal pipe section. Settling of solids was observed near the bottom of the pipe as the mixture flowed from the inlet to the outlet. The volume fraction of the dispersed phase increased from 0 to 0.15 as the bed of solids started forming. As the water was displaced by the solids, the volume fraction of the continuous phase decreased from 1 to 0.85.

Future work will include coupling the chemical reaction to the two phase mixture model. The dynamic processes such as precipitation, particle agglomeration, particle breakup will also be incorporated. The modeling efforts will be expanded from the current laminar models to the turbulent models.

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APPENDIX A

ITS Final Test Report submitted to FIU

Feasibility: Ultrasonic Spectroscopy Characterisation of HLW Stimulant

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October 2012

Customer: Florida International University

Summary

Leading on from a feasibility study on ultrasound spectroscopy for the characterisation of a HLW stimulant at FIU in Miami, ITS agreed to repeat the experiments to further understand the practicality of the method and resolve measurement and analysis issues which were encountered.

The ITS experiments showed that ultrasound spectroscopy, in the through transmission mode, was superior to the pulse echo measurements for any given transducer configuration.

The reflection measurements were proposed as a means of monitoring the density of the solution. However, the data showed significant variability which was traced to two causes; the loss of the reflection signal in the stray echoes and the ill posed nature for the calculation of reflection co-efficient, which lead to a sensitivity to noise. Increasing the size of the transducers reduced the noise and improved the result but using 25mm transducers compared to 10mm devices did not eliminate the problem.

Closer examination showed that the density value results were largely dependent on the speed of sound rather than the reflection co-efficient, which was dominated by the mismatch between the delay line material and the sample. Hence the reflection method offers little advantage over the use of the speed of sound.

The effect of different materials at different concentrations, and different mixtures of materials, was examined to determine the effectiveness of the speed of sound measurement for this application. The modelling and the measurements both showed that the aluminium hydroxide acoustic velocity increased with concentration, while the zirconium dioxide and stainless steel acoustic velocities decreased with concentration. Measurements using mixed samples showed that this resulted in the acoustic velocities of the three samples effectively cancelling each other out with the combined acoustic velocity being less dependent on concentration.

Future work should concentrate on obtaining high quality attenuation data and velocity data using through transmission measurements and a specialist high concentration measurement system, to enable measurements at the concentrations encountered in the HLW tanks at Hanford.



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4. Introduction

FIU previously rented a U2s low concentration system with a 19mm probe from ITS. The objective was to study the possible use of the technique for the characterisation of High Level Waste (HLW) slurry from the holding tanks at Hanford, before or during transfer from the tanks.

The rental results were promising suggesting a possible application in measuring the density using the reflectance method but the data contained significant artefact. FIU and ITS agreed to collaborate to further study the application and determine the source of the artefact and verify the results from the original rental.

For further information on the background of the experiments the reader is referred to **FY2011 Project #1 Technical Report** ask 15: Advanced HLW Instrumentation **Ultrasonic Spectroscopy Phase I Test Summary Report [1]**

This report examined two aspects in detail; the attenuation and velocity characteristics were modelled using a first principles model based on work by Wang and Povey [2] and the measurement of the attenuation, acoustic velocity and density using the U2s system

The report also reviews the earlier work and determines the reasons for the artefacts but, due to the large amount of data and limited time, doesn't dwell on the detail.

Finally the report discusses the physical mechanisms of the interaction of the measurement with the samples and the resultant velocity and density data.

5. Ultrasound Spectroscopy

Ultrasonic spectroscopy (USS) is the characterisation of the ultrasonic response of a material to the propagation of a low level ultrasonic wave. It has a number of inherent and useful features:

- Non-destructive, non-intrusive and inline
- Can be used on opaque samples
- Can be used on both liquids and soft solids
- Simple sample handling with no requirement for manual intervention
- Able to operate over a wide range of frequencies

The ultrasonic spectroscopy data can be used to characterise a range of material or process properties and behaviours including:

1. Measurement of density and or concentration of discrete phases, mixed liquids or slurries
2. Rheological characterisation of liquids and soft solids through phase velocity
3. Indicative changes in attenuation spectra due to changes in material structure including particle size of colloids in both slurries and emulsions
4. The characterisation of liquids containing bubbles



5.1. Ultrasound particle sizing

Particles absorb and diffract ultrasound energy and hence the change in ultrasound energy of a wave in a slurry or emulsion is a function of the particle size. Particle size determination from the ultrasonic attenuation spectra is based on a first principles model.

The attenuation of ultrasound in a dispersed system is caused by a variety of mechanisms (the significance of which depends on material properties, particle size and sound). These mechanisms include

- Intrinsic losses
- Geometric scattering
- Thermal absorption
- Visco-inertial losses

The ultrasonic absorption is also dependant on the contribution of the continuous phase, which is known as the intrinsic attenuation. Therefore the addition or excess attenuation coefficient is given as

$$\alpha_{exc'} = \alpha - \alpha_{int}$$

Scattering: Ultrasonic scattering is the re-direction of acoustic energy away from the incident beam, so it is elastic (no energy is absorbed). The scattering is a function of frequency and particle size and is only significant at very large particle sizes

Thermal losses: Thermal losses are due to temperature gradients generated near the surface of the particle as it is compressed by the acoustic wave. The resulting thermal waves radiate a short distance into the liquid and into the particle. Dissipation of acoustic energy caused by thermal losses is the dominant attenuation effect for soft colloidal particles, including emulsion droplets and latex droplets.

Viscoinertial losses: Viscoinertial losses are due to relative motion between the particles and the surrounding fluid. The particles oscillate with the acoustic pressure wave, but their inertia retards the phase of this motion. This effect becomes more pronounced with increasing contrast in density between the particles and the medium. As the liquid flows around the particle, the hydrodynamic drag introduces a frictional loss. Viscoinertial losses dominate the total attenuation for small rigid particles, such as oxides, pigments, and ceramics.

$$\alpha_{exc} \approx \alpha_{vis} + \alpha_{th} + \alpha_{sc}$$

All of these mechanisms whatever their source are described by the general term scattering. The model seeks to account for each of these mechanism scattered sound fields around a single particle. Based on this, the propagation of sound through the dispersed system can be calculated. By assuming independent scattering it is possible to separately determine the total attenuation as the sum of the individual mechanisms.

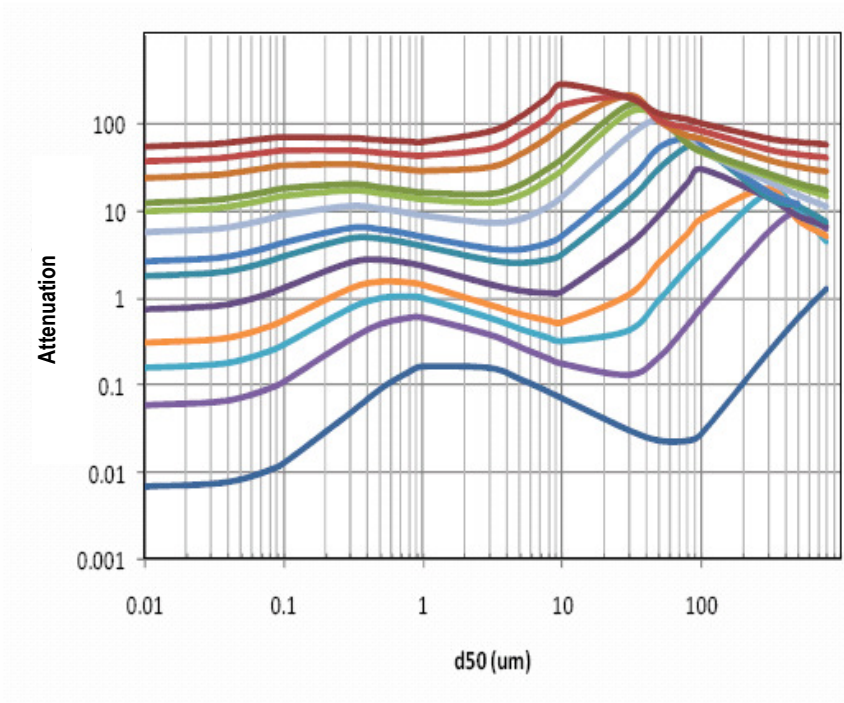


Figure 1: The relationship between particle size and attenuation loss for different frequencies. The figure shows the two major features of the attenuation versus size plot - at small particle sizes the spectra is dominated by differences in physical properties - the mid range is the visco-inertial region, which is a result of the movement of the particle in a viscous liquid - while at large particle sizes the attenuation increases rapidly due to geometric scattering.

5.2. Ultrasound rheology

Ultrasound can be used as a measure the rheological properties of a material. Specifically the attenuation and velocity information can be used to determine the longitudinal bulk modulus of a material.

$$\frac{\omega}{k^*} = \sqrt{\frac{M^*}{\rho}}$$

Where k is the ultrasound wave number ω is the angular velocity ρ and M the longitudinal modulus. * denotes complex values. The solution of the Navier Stokes equation allows the viscosity to be determined from the attenuation and velocity values

$$\eta = \frac{3\rho\alpha c^3}{2\omega^2}$$

The advantage in using the ultrasound spectra is the availability of frequency dependant data which provides an indication Newtonian or Non-Newtonian behaviour.

6. The U2s USS system

The U2s ultrasound spectroscopy system is based on the use of transient measurements and is specifically developed for in-situ or on-line applications. The use of fast transient methods allows data collection in real time allowing the use of the system on dynamic and rapidly evolving applications, such as crystallization. The Hi-C system is unique as it uses a novel measurement system which uses digital ultrasound signals to increase the dynamic range and resolution of the system

The sensor itself is custom designed to match the application. The sensor also does not require the adjustment of transducer spacing's during measurement and hence the sensors are robust and process compatible. The sensor can either be a flow through cell for pipes or a dip in probe for reactors and large vessels.

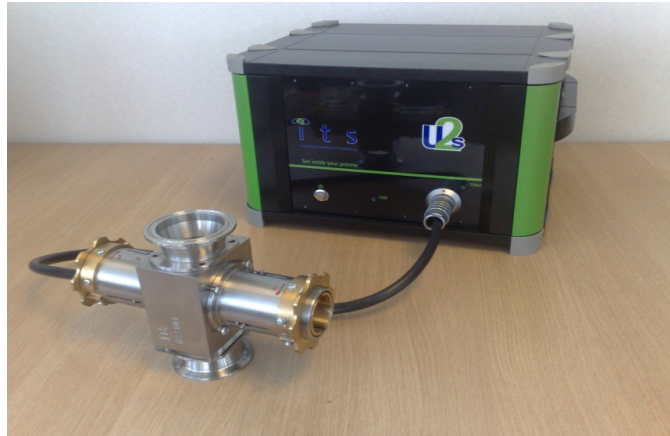


Figure 2: The U2s USS system with sensor used in the experiments

The U2s can be configured to operate in a number of different methods to best match the application.

1. The probe or pipe method: low concentration liquids and colloids and very fast measurements
2. The coded transmission method: low concentration and high concentration liquids and colloids, moderate update rate.
3. The reflectance method: developed for extremely highly attenuating materials close to solid.

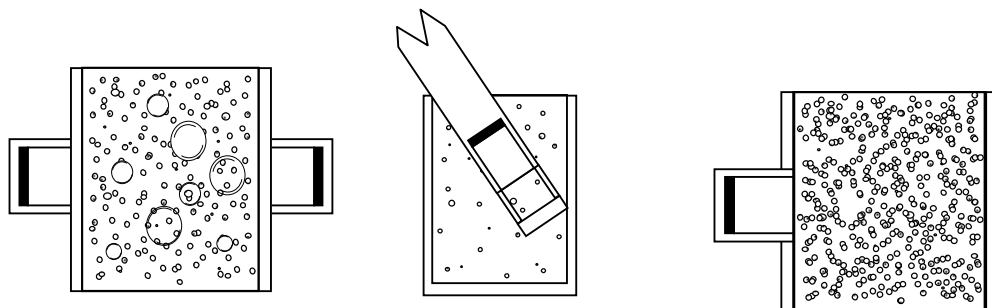


Figure 3 Sensor configurations: transmission, probe, and reflectance

6.1. Reflection measurements

The reflectance method is based on the contact impedance at the interface between the sensor and sample. Reflectance only measures the material characteristics at the interface but can operate in very highly attenuating materials.

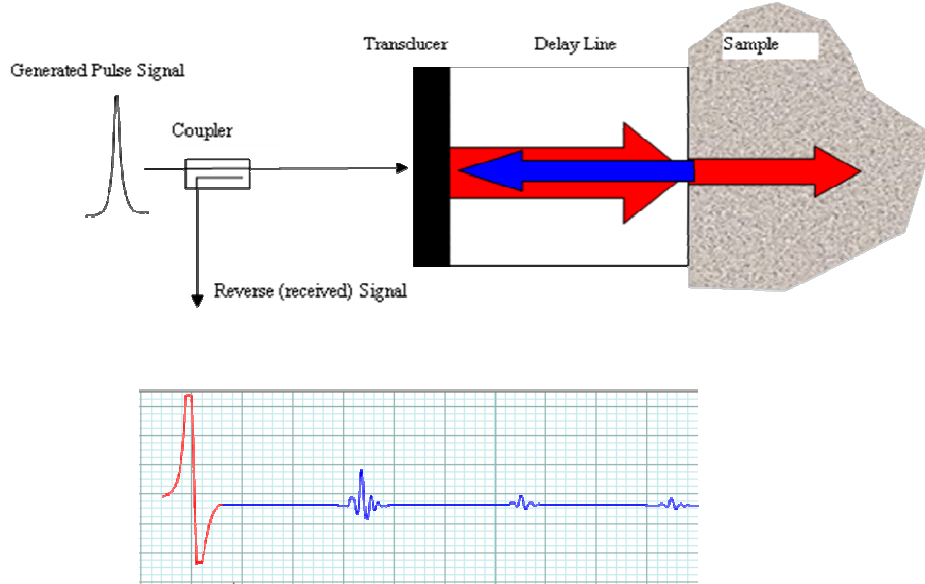


Figure 4 Reflection Mode Measurement Method

The reflectance parameter R can be used to determine the acoustic impedance of the material and hence to calculate the bulk modulus.

$$R = \frac{Z_1 - Z_2}{Z_1 + Z_2} \quad Z^* = R_s + iX_s = \sqrt{\rho \cdot M^*}$$



7. Modelling of Stimulant

The objective of this work was to provide some indication of the expected values for the attenuation and velocity of the samples and constituent materials, by using the theoretical models of acoustic propagation. The quality of these models needs to be qualified at the outset. The current state of the art of ultrasonic propagation modelling is limited to low concentration linear models, by which the attenuation has a linear dependence on concentration. The models do not account for high concentration crowded particle effects and the impact of non-linear rheology.

The source of information used for the physical properties to input to the models were provided by FIU and also by using information from papers and the internet. The initial modelling used information previously obtained from studies of Tank AY-102 to define the particle size and concentration of the samples - as shown in Table 1

Simulant Material	Specific Gravity (S.G.)	Particle size, d50 (um)	Volume %	AY-102 Tank Waste Represented
Gibbsite (Aluminium TriHydrate)	2.42	10	52	Represents 53% of waste by volume
Zirconium Oxide	5.7	12	37	Represents Fe2O3 and MnO2, which make up 40% of waste by volume
F120 Silicon Carbide	3.2	150	6	Ca5OH(PO4)3 makes up 5% solid by volume
Bismuth Oxide	8.9	38	2	Represents PuO2 (the bounding material density in the tank)
Stainless Steel (316)	8	128	3	Represent bounding density and particle size within tank

Table 1 Size and concentration information from Tank HY-102 used as the basis for the sample constituents for the modelling exercise

Material	Density (x1000 kg/m3)	Bulk Modulus (GPa)	Speed of sound (m/s)
Gibbsite (Aluminium TriHydrate)	2.42	85	5926
Zirconium Oxide	5.68	91.6	7380
F120 Silicon Carbide	3.2	178	11590
Bismuth Oxide	8.9	7.86	2526
Stainless Steel (316)	7.86	77.2	5790

Table 2 Physical properties for the materials used in the simulation



The previous literature also provided an example of a particle size distribution for the AY 102 tank as shown in Figure 5

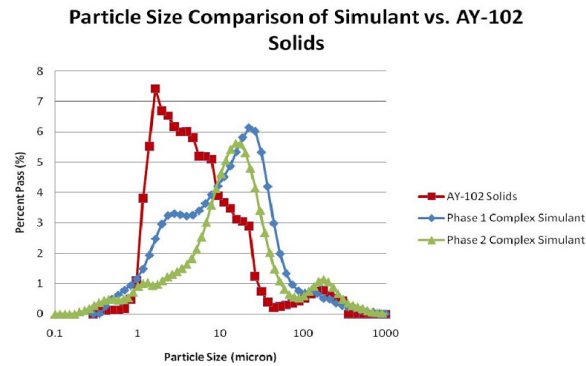


Figure 5 Previously published particle size distribution plots for the AY-102 and previous stimulant particle size plots

It should be noted that the particle size distribution for AY-102 in Figure 5 looks particularly non –physical and the sampling or PSD method should be questioned. Therefore the modelling of the stimulant was based on the D50 data given in Table 1 and the assumed distributions are shown for each component of the mixture in Figure 6.

7.1. Distribution of each discrete phase

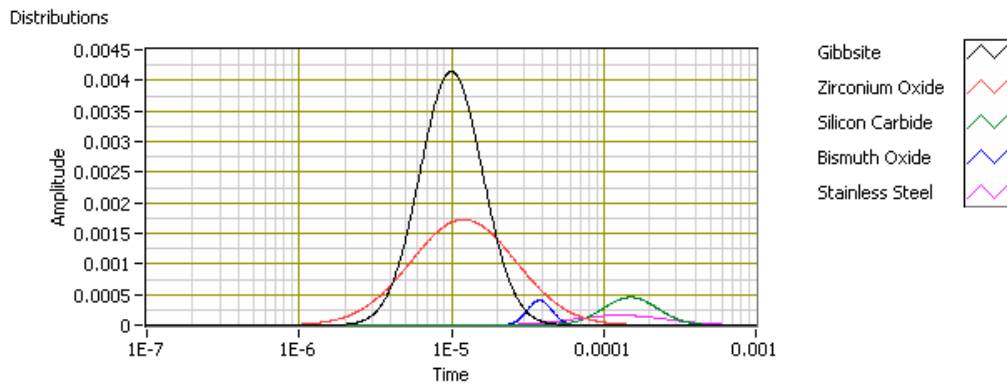


Figure 6 The particle size distributions for each of the constituents of the theoretical stimulant used in this report.

The total distribution for the sample is given in Figure 7, note these results are based on a volume distribution.

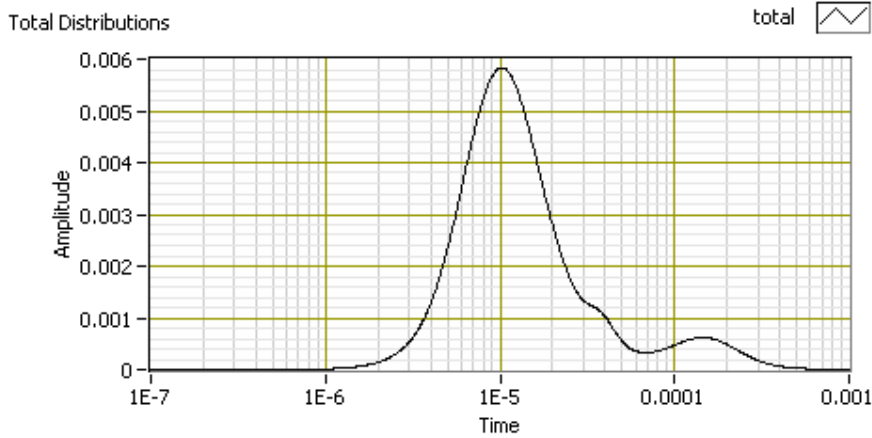


Figure 7 The overall volume distribution of the theoretical stimulant used in the model experiments for this report

The outstanding differences in the model distributions and the reference data from AY-102 are due to

1. The assumption of log normal distributions
2. The existence of a mode <10um in the actual which is not contained within the sample set given
3. The lack of information on the width of the distributions



7.2. Modelling results

The distributions shown above were used to generate spectra for both attenuation and velocity. To simplify the need for physical properties the continuous phase was assumed to be water. The model uses has been described by Wang and Povey and was selected as it included both visco-inertial effects and thermal effects, and is principally based on an explicit solution to Epstein Carhart, Allegra Hawley (ECAH) for the first two scattering terms A_0 and A_1 .

In the case of velocity, previously published work outlines two cases for the value of the velocity in mixtures of two components. Urick used a simple volume averaged form of the Wood equation to determine effective compressibility's and densities for the overall mixtures.

$$c = \sqrt{\frac{1}{\kappa_{eff} \cdot \rho_{eff}}}$$

However, in its original form it is unable to account for the additional effect of scattering due to interfaces within the sample and particles. Hence a modified version of Urick has been described which has already been discussed in the previous report [1]. However, this phenomenological model has limited application. By returning to a first principles scattering model, albeit with a reduced set of scattering terms, we should be able to get a closer agreement for suspensions.

7.2.1. Individual materials

As can be seen in Figure 8, at the concentrations specified in Table 1 the zirconium dioxide and aluminium hydroxide attenuations are substantially greater, but, even at very low concentrations, the stainless steel is contributing significant scattering and hence would be expected to have a substantial effect at higher concentrations.

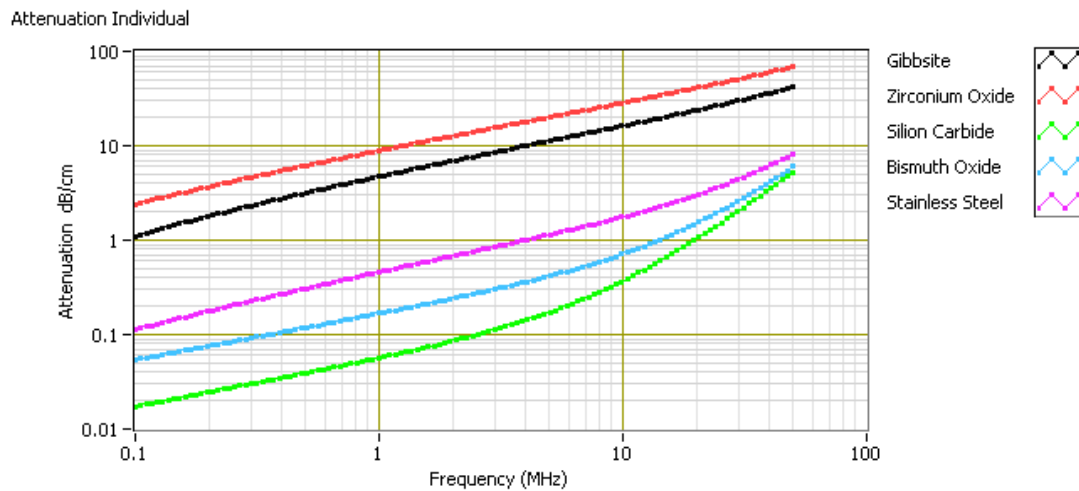


Figure 8 Simulated attenuation spectra for the individual constituents of the stimulant material

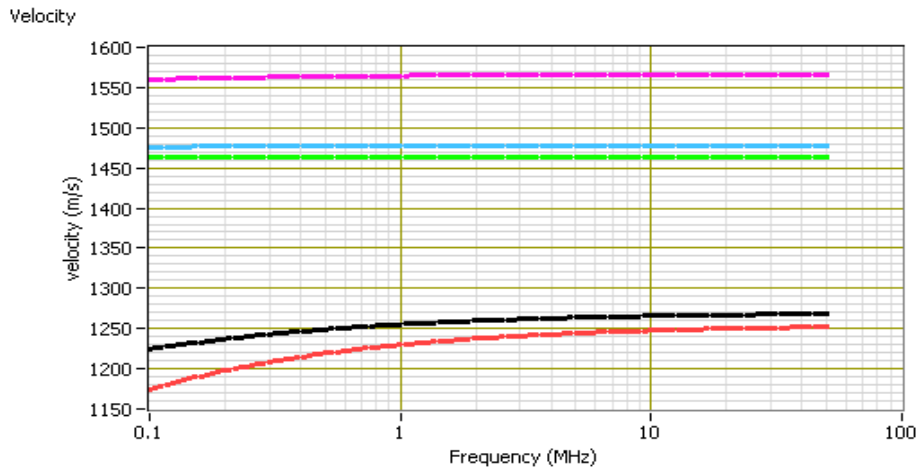


Figure 9 Simulated velocity spectra for the individual constituents of the stimulant material plot - labelling is identical to Figure 8

Due to the substantial density differences of stainless steel to water, the velocity spectra of the stainless steel shows the largest shift in magnitude.

7.2.2. Total sample attenuation

The attenuation spectra for the individual materials can be considered to be additive when determining the overall excess attenuation, while the scattering remains in the low concentration linear range. Hence the excess attenuation was calculated as shown in figure 10 and plotted on dB/m scales to allow comparison with previous data.

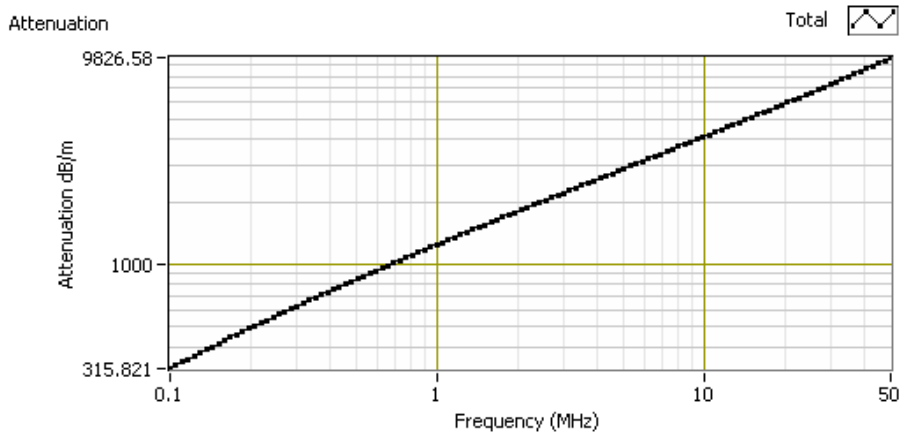


Figure 10 The spectra of the total attenuation for the stimulant material based on the values given in Table 1

The results show a highly attenuating sample even at low frequencies



7.2.3. Total sample velocity

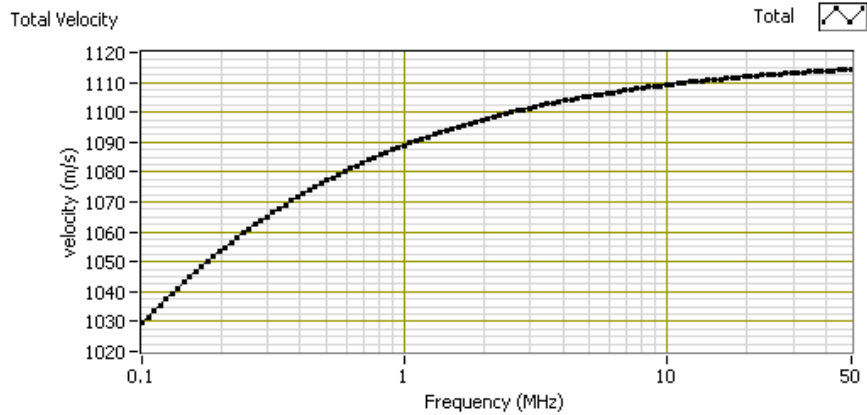


Figure 11 The spectra of the total velocity for the stimulant material based on the values given in Table 1

The predicted values of the velocity spectra are surprising low considering the density of the stainless steel and zirconium dioxide.

7.2.4. Comparison with actual

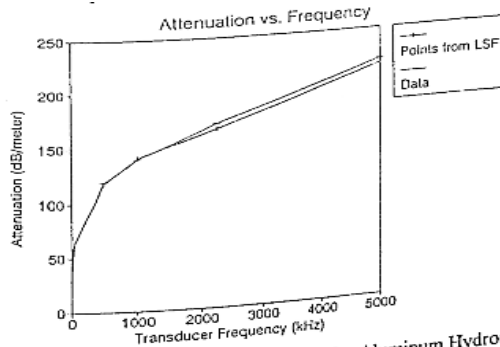


Figure 4. Attenuation vs. Transducer frequency for Aluminum Hydroxide Sludge.

*drawing is best quality available

Figure 12 Attenuation values from measurements of sample recovered from the tanks at Hanford

The model attenuation is lower than the actual attenuation in the tanks at Hanford this is probably due to the non-linear effects of the high concentrations of solids in the tanks as described below.

7.2.5. Effectiveness of high concentrations

At high concentrations the material's attenuation is reduced by particle to particle transmission. This is illustrated below for 100nm silica from 1-50v%. This is described as multiple scattering by some authors.

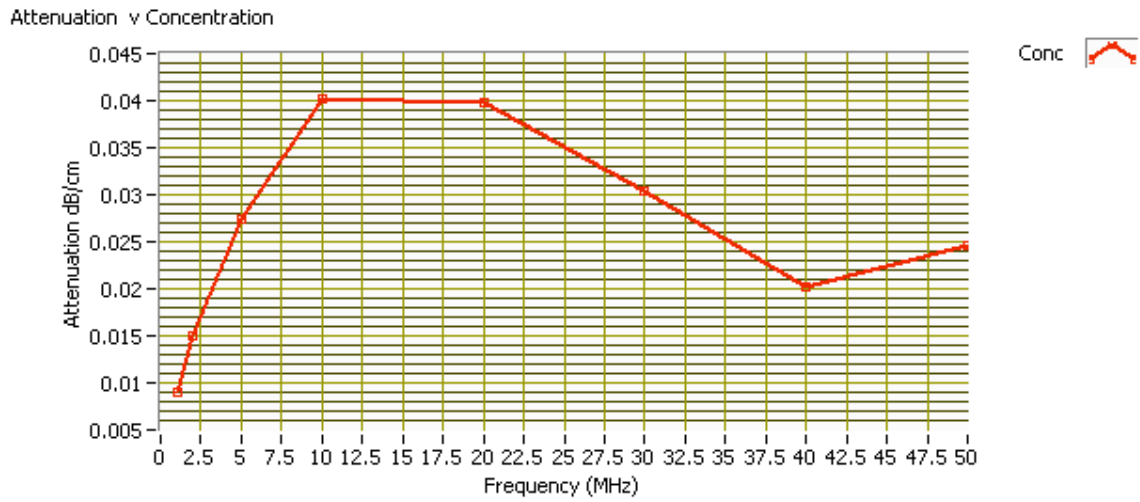


Figure 13 A plot of measured attenuation verses concentration for a 300nm silica sample showing the impact of the non-linear relationship between scattering and concentration at higher concentration levels.

Hence we can assume that the data from the model using “single scattering” is the maximum limiting case.

7.3. Simulated concentration ladders.

The experimental work described in this report is based on studying the effect of changes in concentration and hence density on the acoustic spectra. Knowledge about the likely impact of each constituent material on the attenuation and velocity spectra is useful in assessing the likely overall response and in the interpretation of the results.

7.3.1. Simulated aluminium hydroxide concentration ladder

Attenuation Spectra

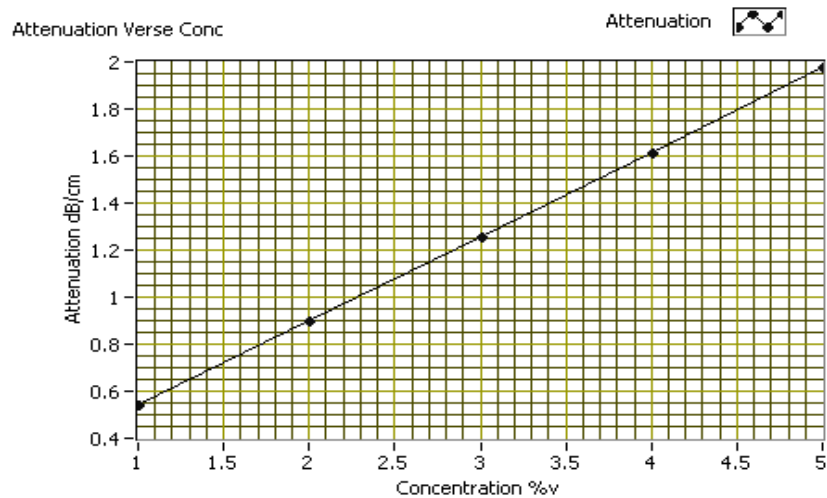


Figure 14 Simulated attenuation at 9.9MHz verses concentration (below 5%v) for aluminium hydroxide



Velocity Spectra

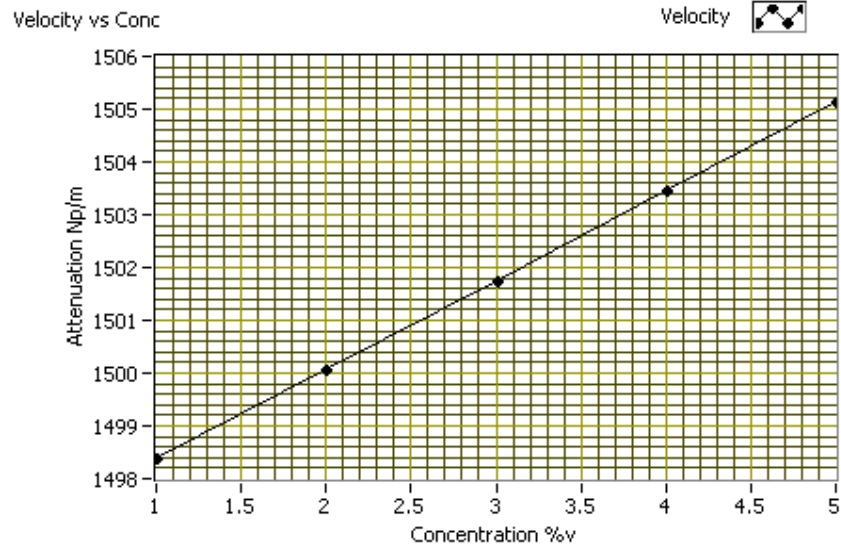


Figure 15 Simulated velocity at 9.9MHz verses concentration (below 5%v) for aluminium hydroxide

7.3.2. Simulated zirconium dioxide concentration ladder

Attenuation Spectra

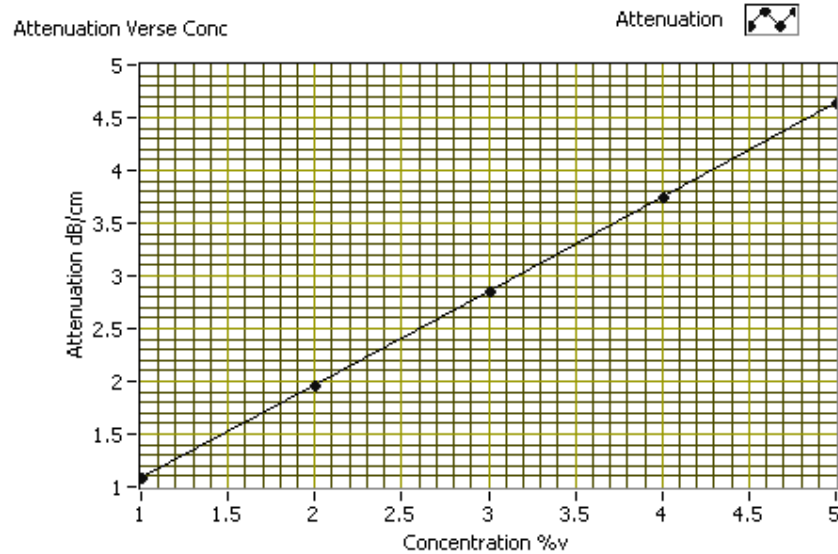


Figure16 Simulated attenuation at 9.9MHz verses concentration (below 5%v) for zirconium dioxide



Velocity Spectra

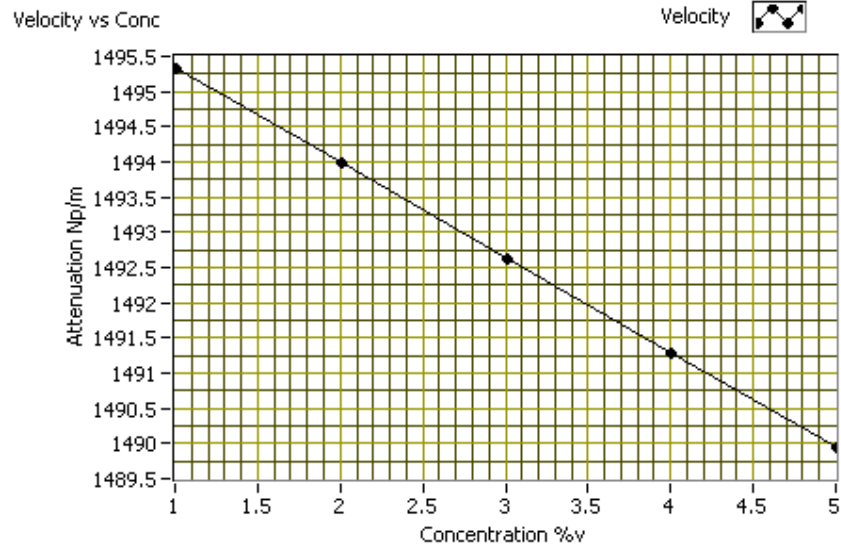


Figure 17 Simulated velocity at 9.9MHz verses concentration (below 5%v) for zirconium dioxide.

7.3.3. Simulated stainless steel concentration ladder

Attenuation Spectra

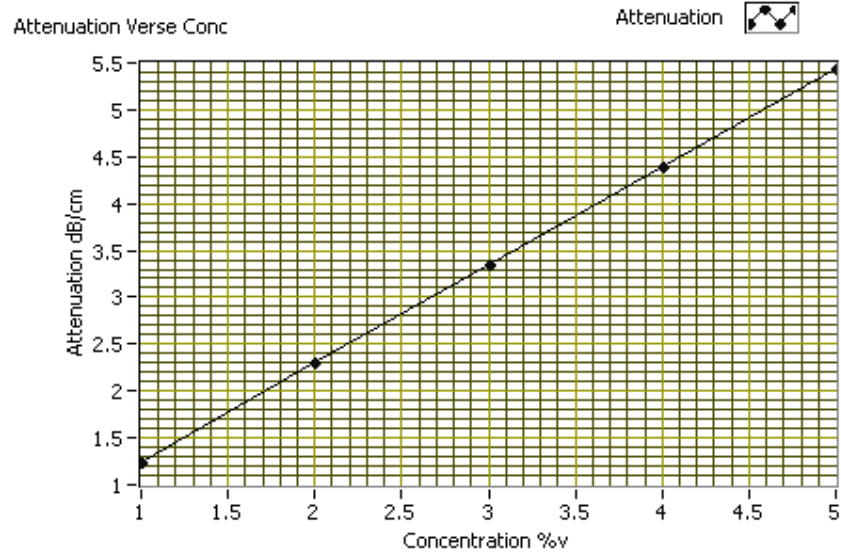


Figure 18 Simulated attenuation at 9.9MHz verses concentration (below 5%v) for stainless steel.



Velocity Spectra

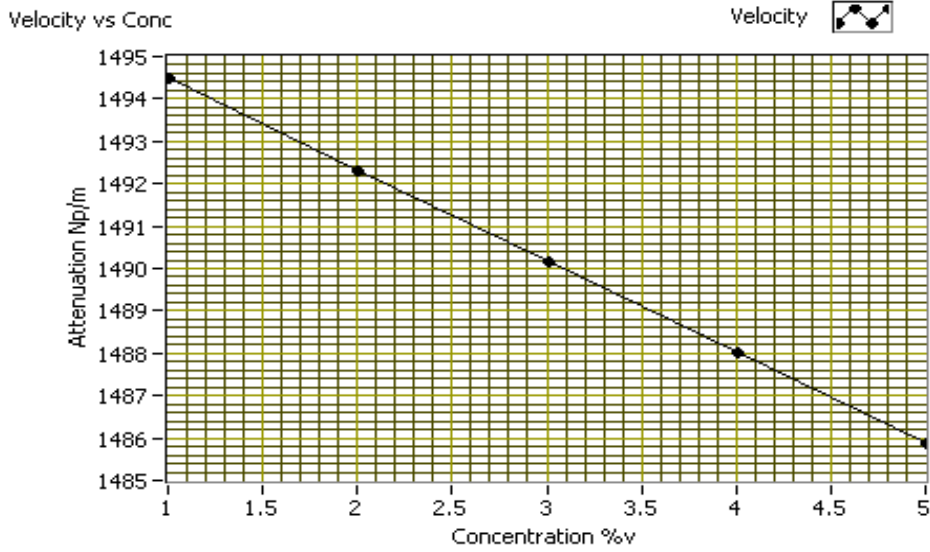


Figure 19 Simulated velocity at 9.9MHz verses concentration (below 5%v) for stainless steel



8. Experimental programme

8.1. Experimental plan

The experimental plan was to measure the ultrasonic characteristics of

- Sodium nitrate aqueous solution
- Aluminium hydroxide in water
- Aluminium hydroxide in sodium nitrate aqueous solution
- Zirconium dioxide in water
- Stainless steel in water
- Complex powder mixtures in sodium nitrate aqueous solution

8.2. Experimental method

Materials:

- 1) Liquids:
 - i. Distilled water, NaNO_3 ,
- 2) Solids:
 - i. Aluminium hydroxide $\text{Al}(\text{OH})_3$ (**S1**),
 - ii. Zirconium oxide (ZrO_2) (**S2**)
 - iii. Stainless steel – 316(SS) (**S3**)

Samples: slurry stimulants, ten repeats of the measurements for every sample.

- All solutions/suspensions were prepared according to proposed concentration ladder 0.1%, 0.5%, 1%, 2%, 3%, 5%, 7%, 10%, 15%, 20%, 25%, (30%, 37.5% for sodium nitrate, precipitation of the solid in the sample) apart from complex powder mixture samples where the concentration ladder was 1%, 2%, 4%, 6%, 8%, 10%.
- The suspensions were mixed during measurements using a magnetic stirrer.
- The only solutions when the stirrer hasn't been able to provide the even suspension during the measurements were the stainless steel ones and so the results from these samples are the most inconsistent.

Prepared samples:

- 1) Aqueous solution NaNO_3
- 2) Distilled water + aluminium hydroxide $\text{Al}(\text{OH})_3$,
- 3) Distilled water + zirconium oxide (ZrO_2)
- 4) Distilled water + stainless steel – 316(SS)
- 5) Solutions of NaNO_3^* + S1
- 6) Solutions of NaNO_3^* + mixtures of solids

*: All $\text{Na}(\text{NO}_3)_3$ Solutions were prepared at 13v% and a relative bulk density of 1.17



Parameters of the experiment: room temperature, the volume of the sample 350 ml.

8.3. The preparation of the solutions

An appropriate amount of the substance was dissolved in a medium, the volume of the solution 350ml, using the measures listed in Appendix 1.

8.4. Methodology

The measurements were carried out using the USS cell in the static and dynamic conditions with a low and Hi-C (aluminium hydroxide) concentration system. For purpose of the experiments the immersion transducers were used with no delay line. Every experiment consisted of 10 repeated measurements for each sample. The sample was then removed from the cell for disposal and new material was added. In the case of sodium nitrate solutions the spacing between transducers was equal 22 mm, for suspensions it was 10 mm. The system was calibrated using deionised water as a reference material.

Reflection measurements were carried out using probe in a dynamic condition and Hi-C system. For this experiment the probe containing transducer with the delay line was used. Ten measurements were collected for every sample. For all types of tested suspensions the spacing between transducer and reflector was kept as 4 mm. The system was calibrated using deionised water as a reference material.

Solids Loading Ladder Tests:

ITS performed two sets of solids loading ladder tests on the USS system in a bench-scale setup approved by FIU. The solids load ladder tests were performed on solutions with a mixture of FIU-provided solids consisting of aluminium hydroxide, zirconium oxide and stainless steel 316. The mixtures had varying bulk densities in the range of 1.15 to 1.55 g/cc. The solvent was either distilled, RO/DI water, or a sodium nitrate/water solution. Tests were performed on stirred solution, which provided sufficient mixing power to maintain the solids in suspension. The mixing process is documented in the test results report.



9. Results

9.1. Sodium nitrate solution

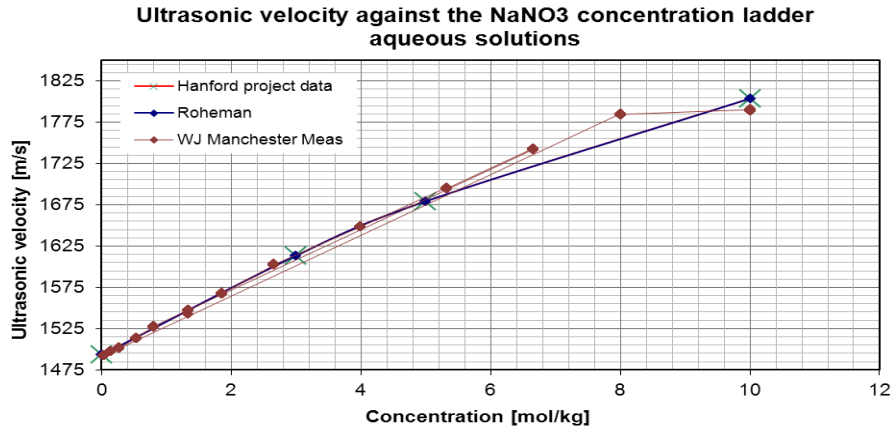


Figure 20. Comparison of ultrasonic velocity values from measurements of NaNO₃ aq solutions collected with the USS instrument in Manchester with the presented data in the Phase1, Hanford project and the data published by Rohman N., Mahiuddin S. [1].

The velocity data obtained (Figure 20.) gives the R² value relatively close to 1. The Hanford and Rohman data overlapped, each showing lower velocity values to the ones collected with the USS equipment in Manchester. The ultrasonic velocity values increase with the increased amount of NaNO₃ in the aqueous solutions. The initial set of experiments was carried out at room temperature in the range 22-25°C (digital reading from the software, not confirmed with independent source). It was noted that a change in the hydration process increases with the increasing temperature [2]

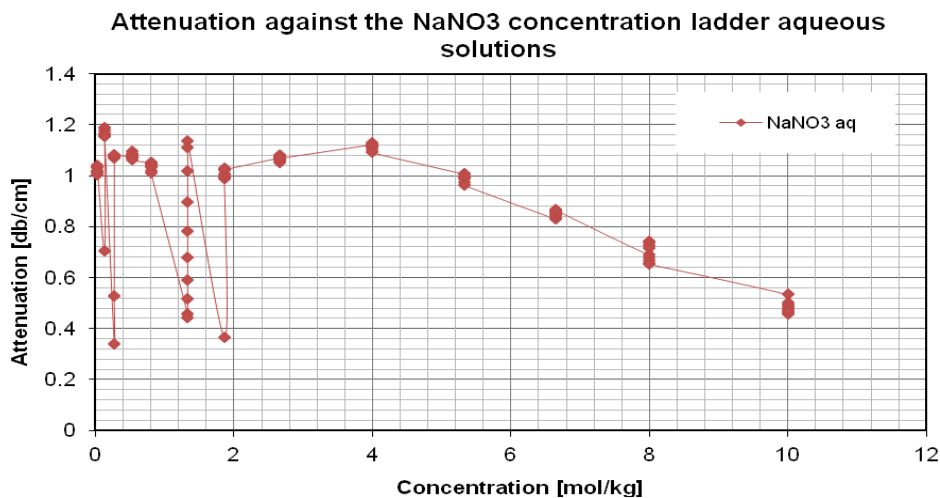


Figure 21 Attenuation values from measurements of NaNO₃ aq solutions collected by the USS instrument low concentration system.



The attenuation values of sodium nitrate are the attenuation values for a pure liquid and hence extremely low values. They would appear to be lower than water and hence are extremely difficult to measure as pure liquids. Water in particular is seen by many researchers to have too low a value to be measured accurately using through transmission techniques.

9.2. Solid suspensions

The next step involved preparation of the suspensions in liquid phase (deionised water) and the appropriate powder substance.

9.2.1. Suspension of aluminium hydroxide $Al(OH)_3$ in water

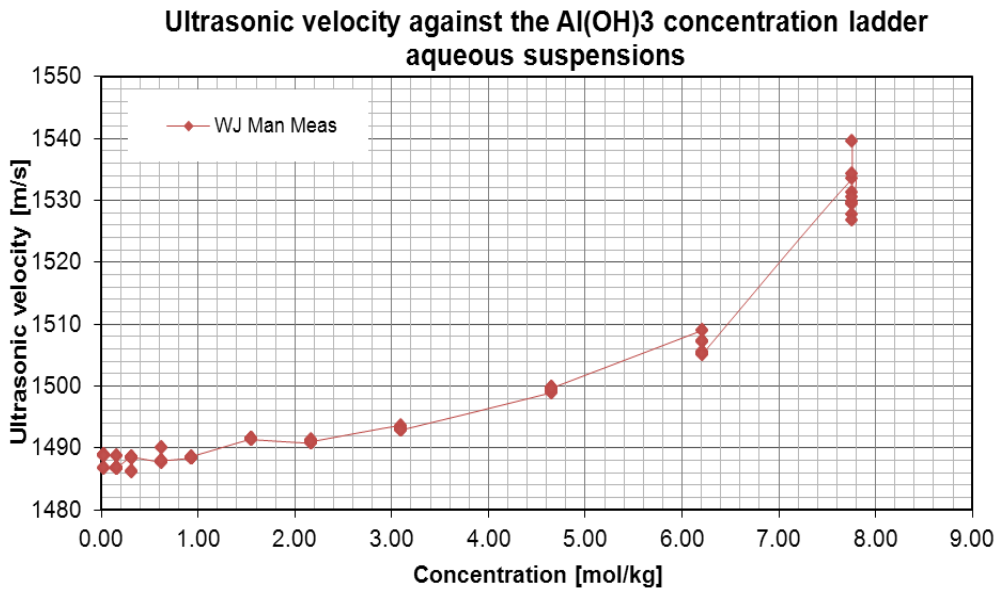


Figure 22 Ultrasonic velocity values from measurements of $Al(OH)_3$ aq suspensions collected by the USS instrument low concentration system.

The analysis of the collected measurements (Figure 22) showed the multiple scattering occurring in the samples with high concentration of aluminium hydroxide. The velocity values were increasing with the increase of the concentration of the solid in a mixture.

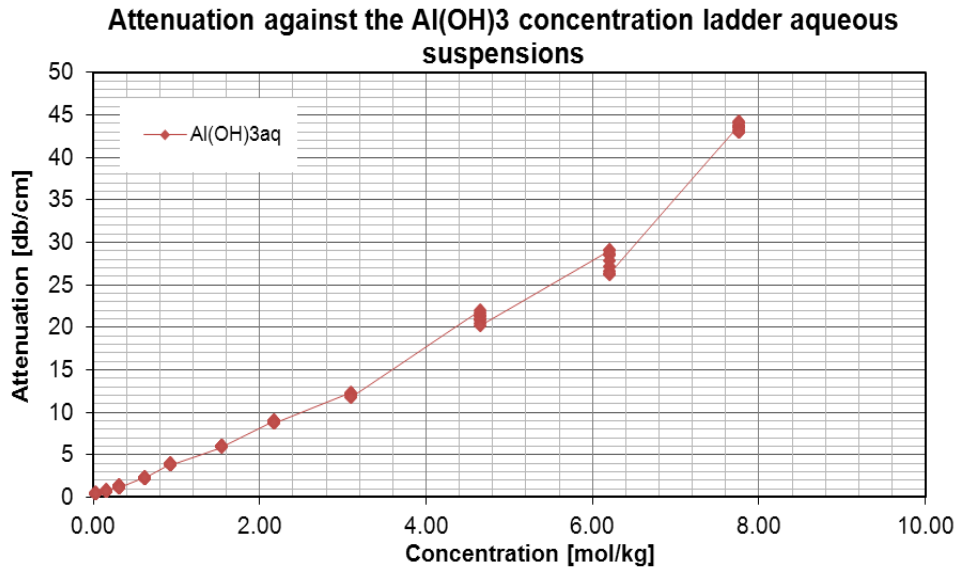


Figure 23 Attenuation values from measurements of Al(OH)₃ aq suspensions collected by the USS instrument low concentration system.

The attenuation results for aluminium hydroxide (Figure 23) show a linear increase with attenuation suggesting that at the measurement frequencies there is no non-linear effects due to particle – particle interactions and that within the measurement time scale there was no sedimentation. During the measurement sedimentation did not occur in this sample.

9.2.2. Suspension of zirconium oxide ZrO₂ in water

Zirconium dioxide samples were made up using the same concentration ladder as for aluminium hydroxide to compare the performance of the system with a high density material.

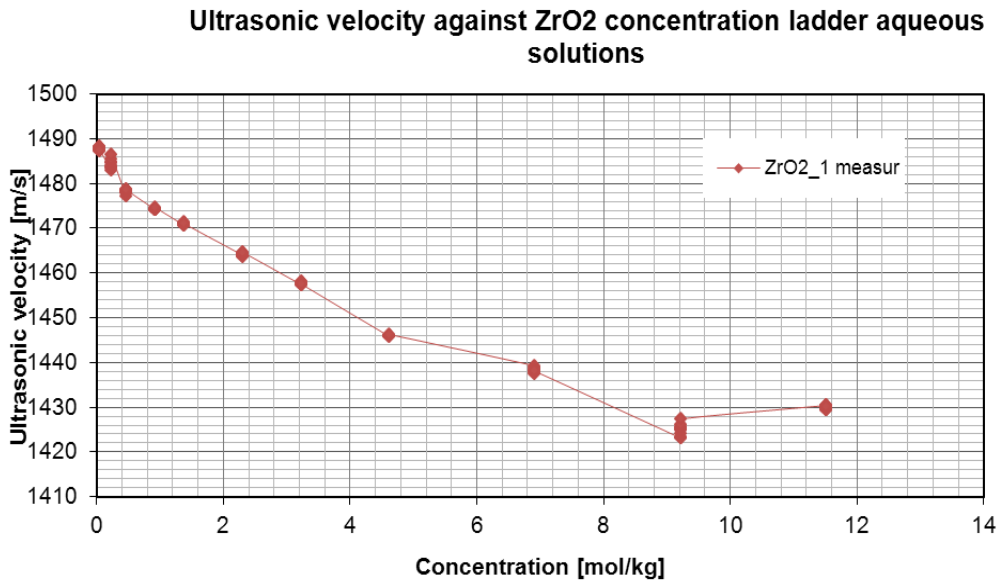


Figure 24 Ultrasonic velocity values from measurements of ZrO₂ aq suspensions collected by the USS instrument low concentration system.

9.2.3. Repeatability of the measurements

Further zirconium dioxide measurements are shown in Figure 25 and Figure 26 to demonstrate that the results are repeatable.

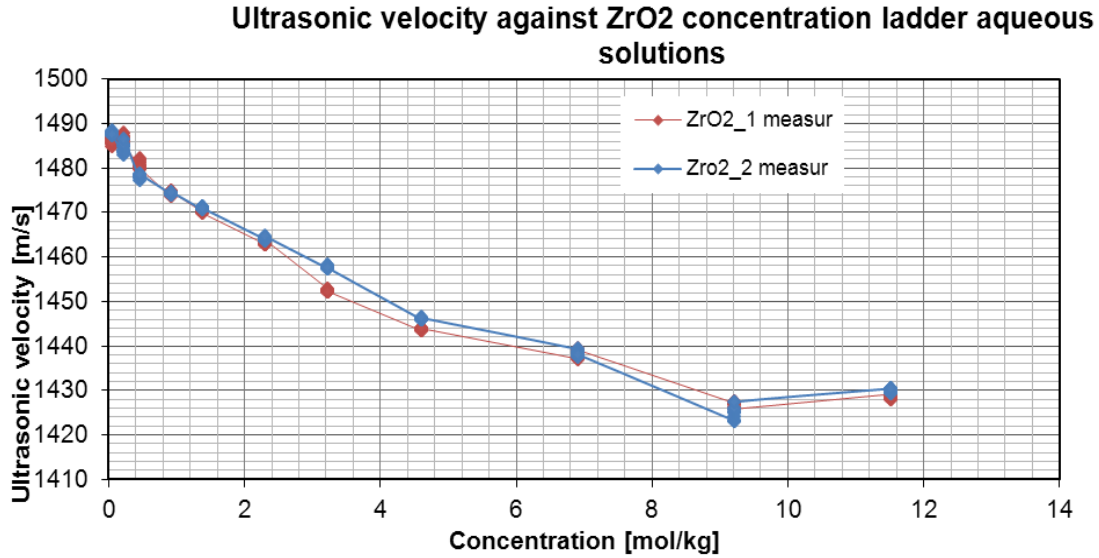


Figure 25 Repeatability of the ultrasonic measurements of ZrO₂ aq suspension collected by the USS instrument low concentration system.

The measurements collected from the suspensions of zirconium oxide in water showed an inversely proportional velocity to the concentration of the sample. The repeated experiment proved good reproducibility of the collected results on the low concentration system. Multiple scattering in the high concentrated samples was also observed.

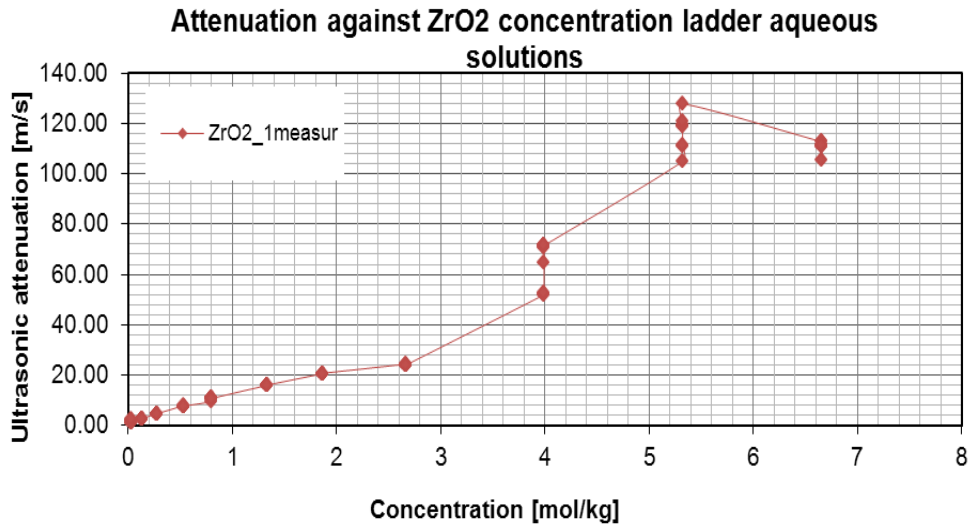


Figure 26 Attenuation measurements of ZrO₂ aq suspension collected by the USS instrument low concentration system.

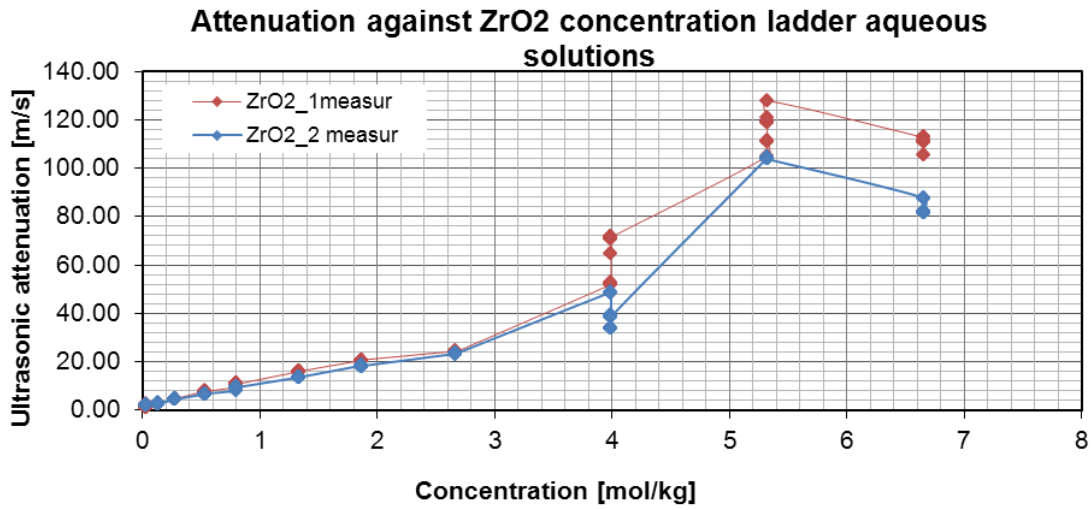


Figure 27 Repeatability of attenuation measurements of ZrO₂ aq suspension collected by the USS instrument low concentration system.

9.2.4. Suspension of stainless steel (SS316) in water

Velocity results

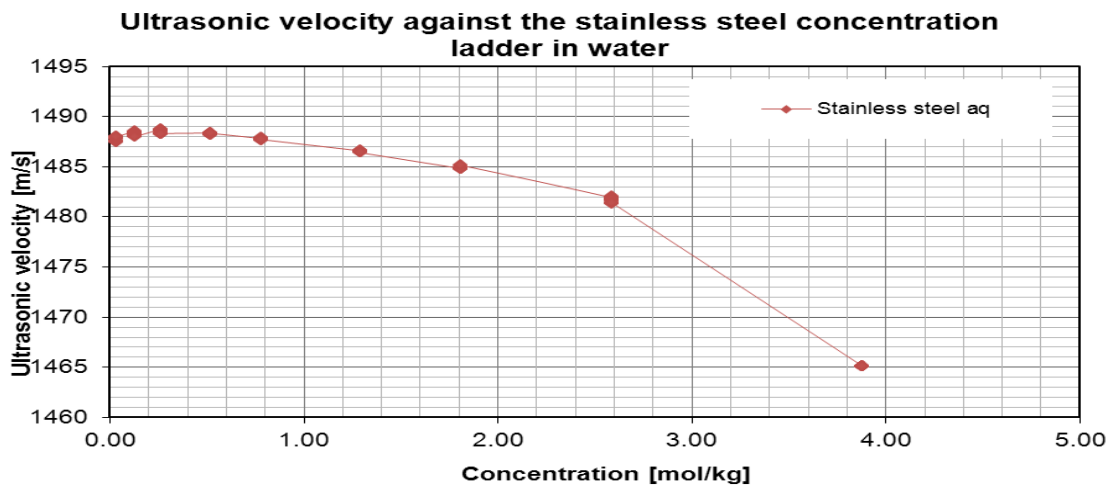


Figure 28 Velocity measurements of stainless steel aq suspension collected by the USS instrument low concentration system.



Attenuation results

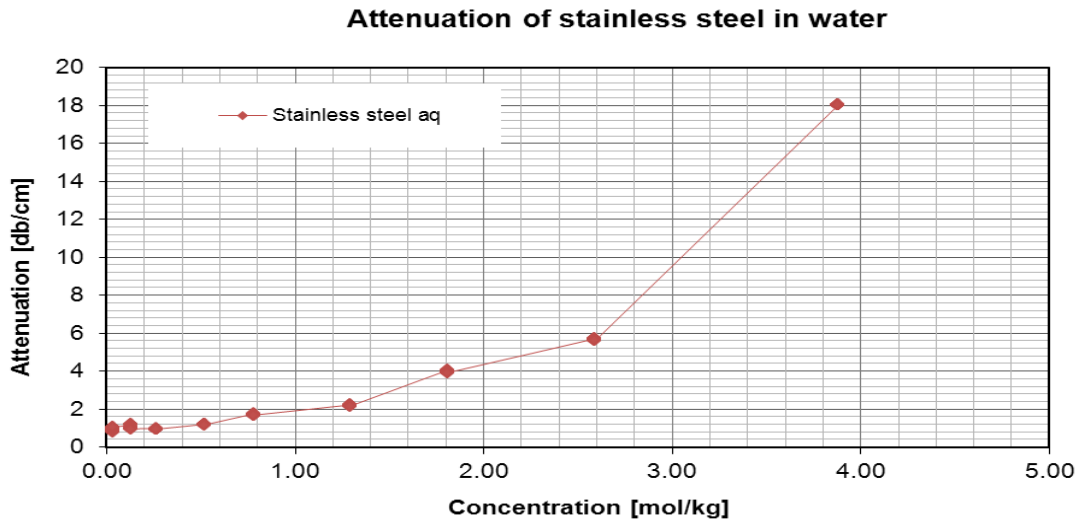


Figure 29 Attenuation measurements of stainless steel aq suspension collected by the USS instrument low concentration system.

9.2.5. Suspension of aluminum hydroxide in NaNO₃aq solution

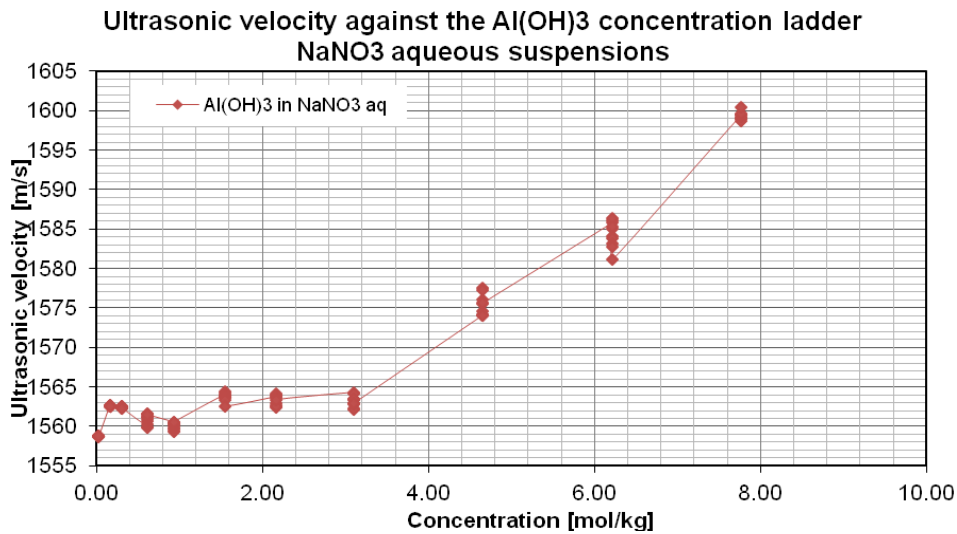


Figure 30 Ultrasonic velocity values from measurements of Al(OH)₃ in NaNO₃ aq suspensions collected by the USS instrument low concentration system.

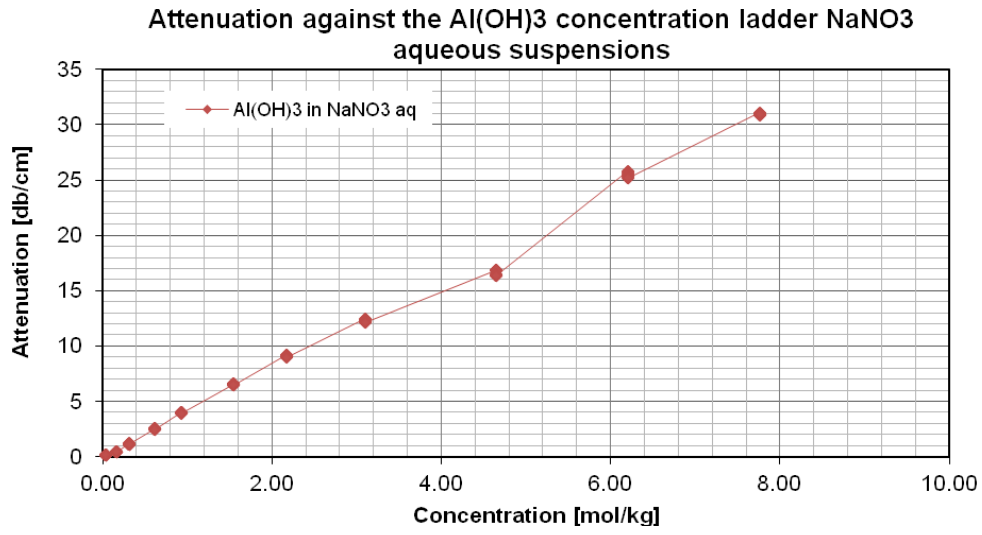


Figure 31 Attenuation values from measurements of Al(OH)₃ in NaNO₃ aq suspensions collected by the USS instrument low concentration system



9.2.6. Complex powder mixtures in NaNO₃ aqueous solution

The volume concentration of the Na(NO₃)₃ was 13% and the bulk density was 1.17.

Solids Content required concentrations per mixture	
Compound	vol % Total solids
Gibbsite	52
Zirconium Oxide	37
Stainless Steel	11
Combine Stimulants	100

Table 3 The volume percentage for a solids used to prepare the complex mixtures. Attenuation values from measurements of complex powders mixture in NaNO₃ aq suspensions collected by the USS instrument Hi-C concentration system.

Concentration of mixture [%]	Total mass [g]
1	14.69
2	29.39
4	60.04
6	93.92
8	125.24
10	160.09

Table 4 Total amount of mixed powders used to prepare the complex samples

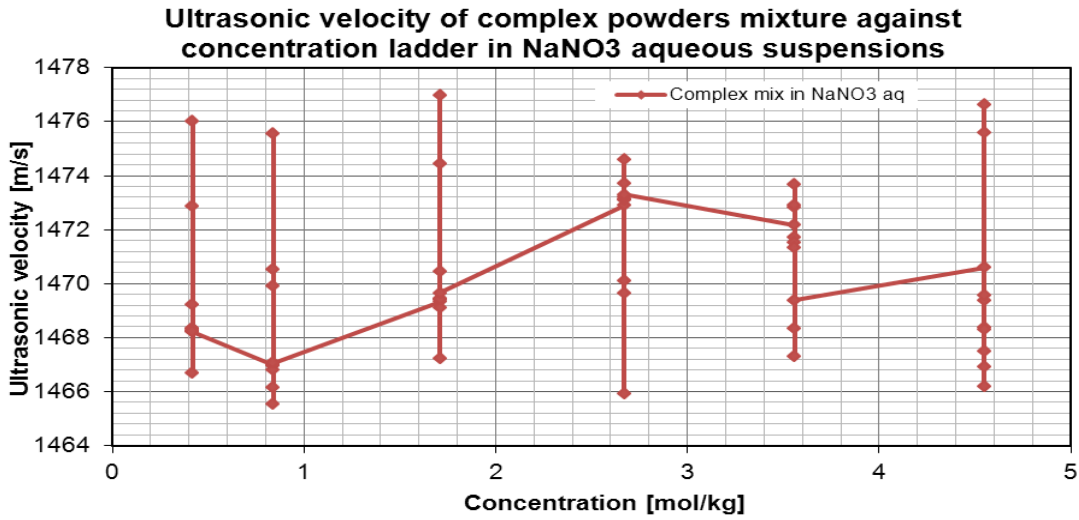


Figure 32 Ultrasonic velocity values from measurements of complex powders mixture in NaNO₃ aq suspensions collected by the USS instrument Hi-C concentration system

The bigger concentration of the sample the more tendency for particles to sediment, although the samples were mixed with the visible vortex on top of the surface.

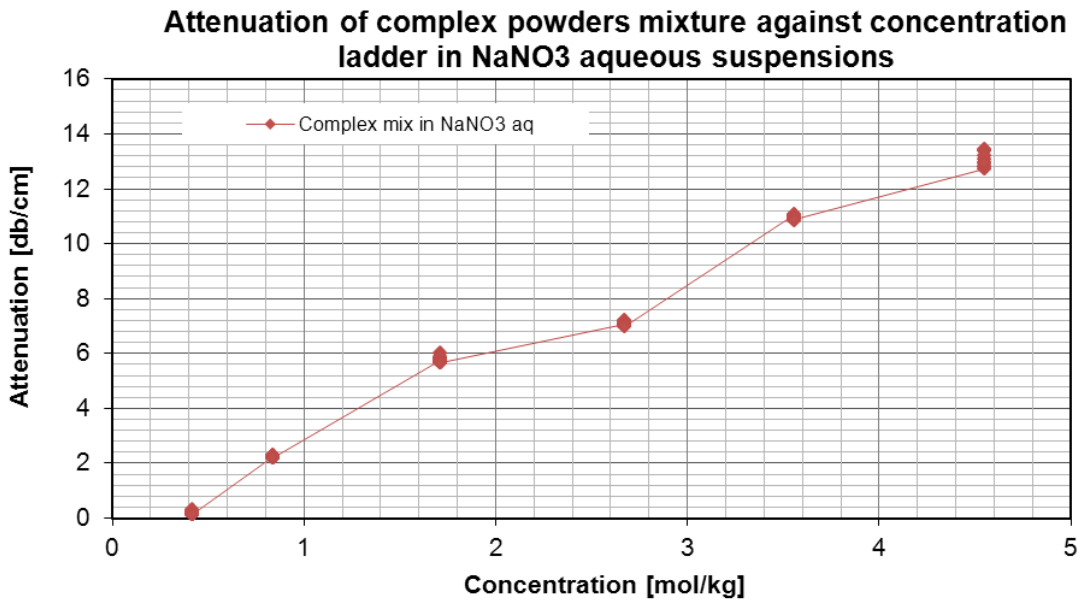


Figure 33 Attenuation values from measurements of complex powders mixture in NaNO₃ aq suspensions collected by the USS instrument Hi-C concentration system



9.3. Density measurement

9.3.1. Density of Na(NO)₃ solution

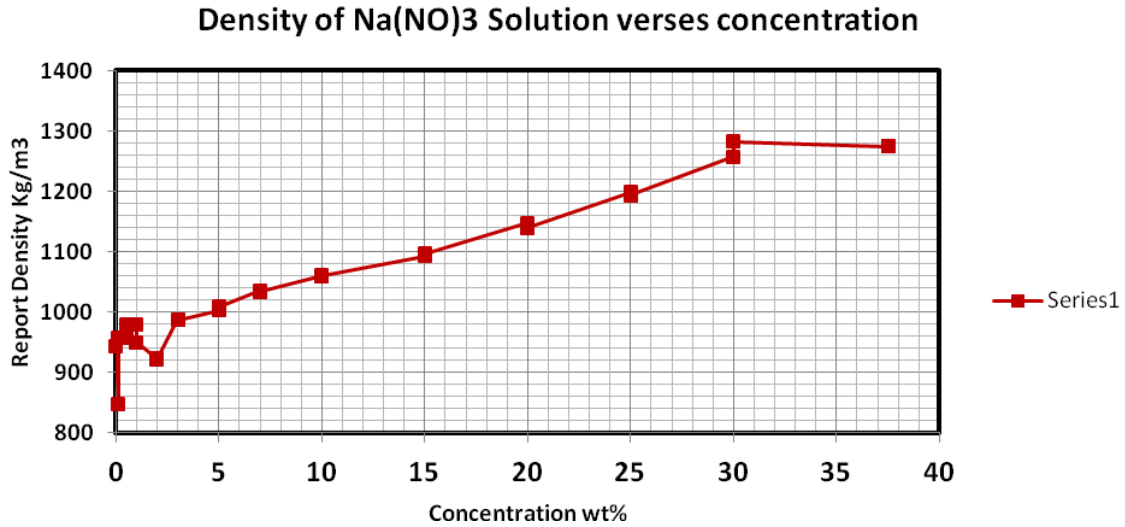


Figure 34 Density values from measurements NaNO₃ aq solutions collected by the USS instrument Hi-C concentration system

%volume	measured density	calculated density	vol average density	relative error [%]
0.1	1	1001.26	999.26	0.13
0.5	1.01	1006.31	1004.32	0.37
1	1.01	1012.61	1010.63	0.26
2	1.03	1025.22	1023.26	0.47
3	1.04	1037.83	1035.89	0.21
5	1.06	1063.05	1061.15	0.29
7	1.09	1088.27	1086.41	0.16
10	1.13	1126.1	1124.3	0.35
15	1.19	1189.15	1187.45	0.07
20	1.25	1252.2	1250.6	0.18
25	1.32	1315.25	1313.75	0.36
30	1.38	1379.23	1376.9	0.06
37.5	1.47	1474	1471.63	0.27

Table 4 Comparing measured density NaNO₃ aq solutions with calculated density

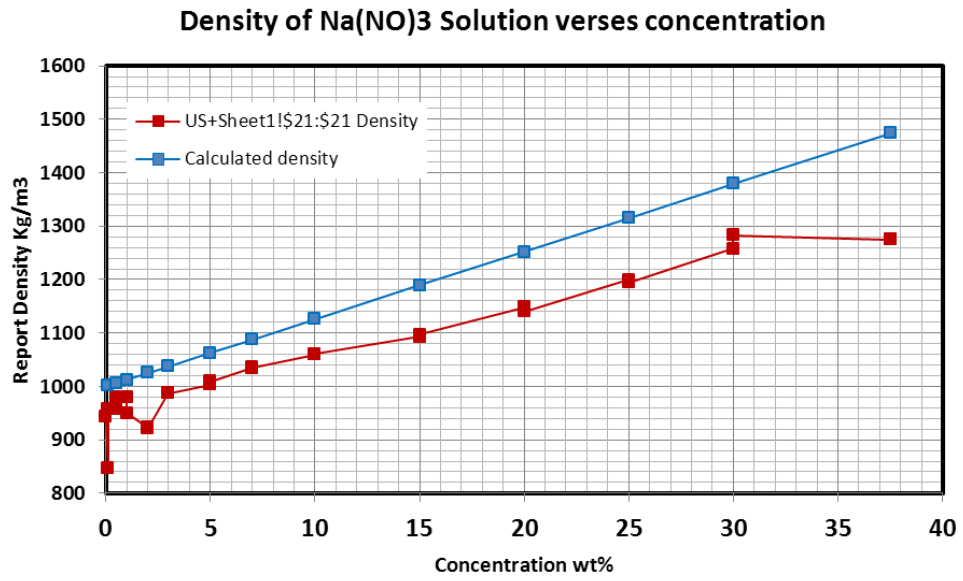


Figure 35 Density values from measurements NaNO₃ solution compared to the calculated value



9.3.2. Density of aluminium hydroxide suspension in Na(NO₃)

% volume	measured density	calculated density	relative error [%]
0.1	1.17	1.17166	0.14
0.5	1.18	1.17831	0.14
1	1.19	1.18661	0.29
2	1.2	1.20323	0.27
3	1.22	1.21984	0.01
5	1.25	1.25307	0.24
7	1.29	1.2863	0.29
10	1.34	1.33614	0.29
15	1.42	1.41921	0.06
20	1.5	1.50228	0.15
25	1.59	1.58535	0.29

Table 5 Comparing measured density of aluminium hydroxide in NaNO₃ aq suspensions with calculated density

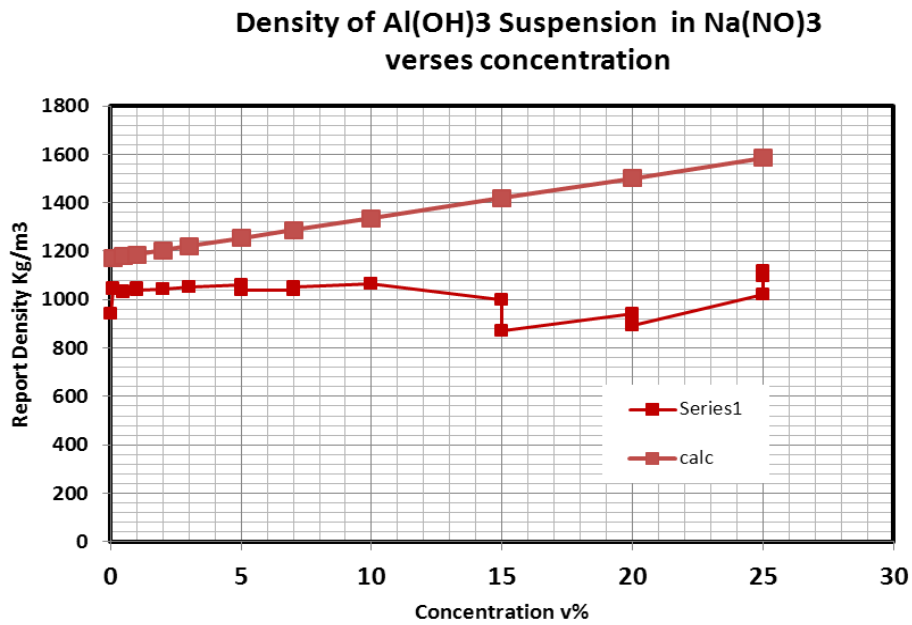


Figure 36 Density values for aluminium hydroxide in NaNO₃ aq suspensions plotted against calculated values



9.3.3. Density of aluminium hydroxide suspension in water

% volume	measured density	calculated density	relative error [%]
0.1	1	1001.42	0.14
0.5	1.01	1007.1	0.29
1	1.01	1014.2	0.41
2	1.03	1028.4	0.16
3	1.04	1042.6	0.25
5	1.07	1071	0.09
7	1.1	1099.4	0.05
10	1.14	1142	0.18
15	1.21	1213	0.25
20	1.28	1284	0.31
25	1.36	1355	0.37

Table 6 Comparing measured density of aluminium hydroxide in water suspensions with calculated density.

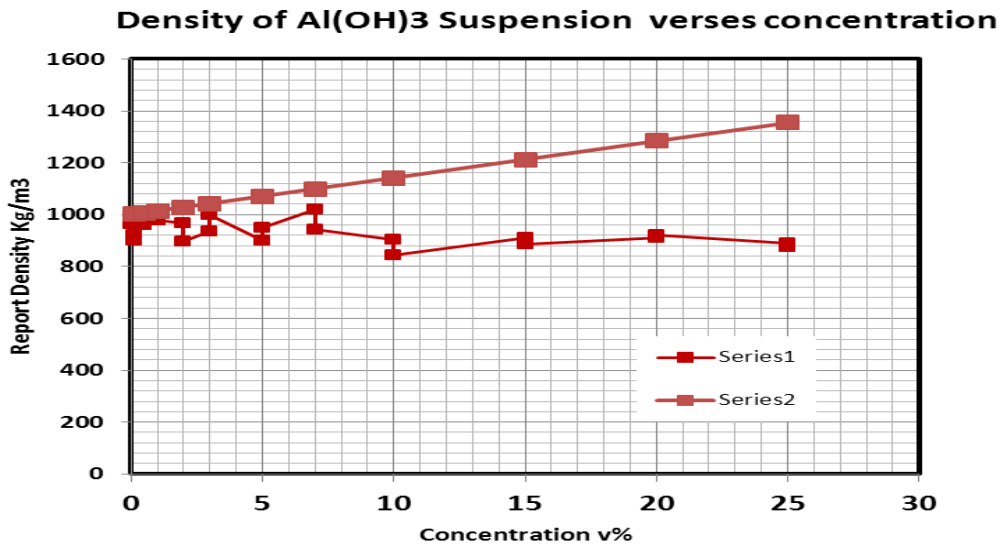


Figure 37 Density values for aluminium hydroxide in water suspensions plotted against calculated values



9.3.4. Density of complex mixture in Na(NO)₃

% volume	measured density	calculated density	relative error [%]
1	1.07	1200.398	10.86
2	1.024	1230.81	16.80
4	1.057	1294.207	18.33
6	1.047	1364.286	23.26
8	1.087	1429.074	23.94
10	1.101	1501.173	26.66

Table 7 Comparing measured density of complex mixture in Na(NO)₃ solution with calculated density.

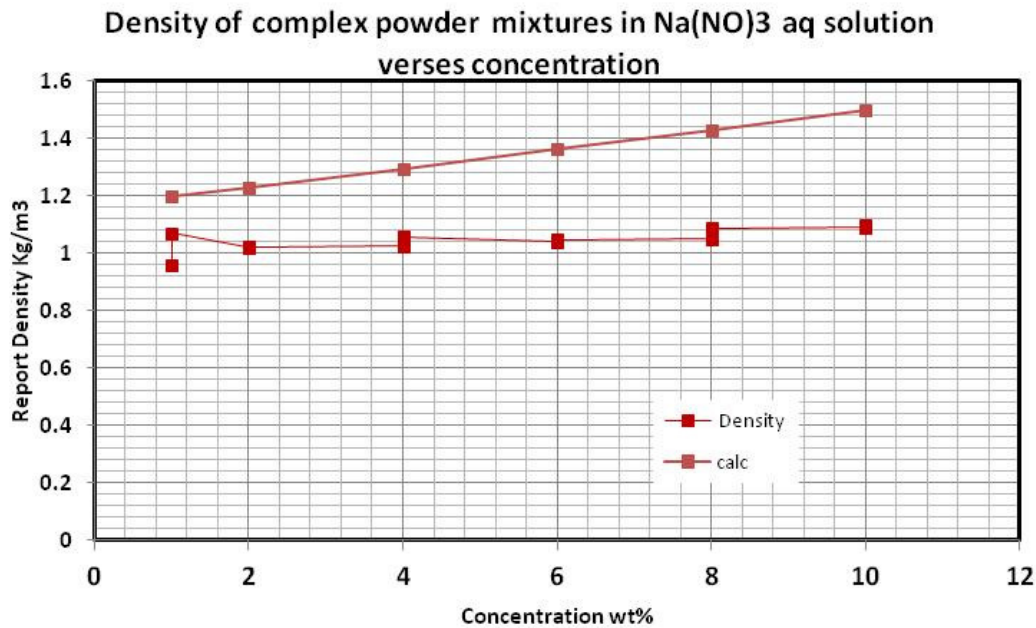


Figure 38 Density values for a complex mixture in Na(OH)₃ solution plotted against calculated values

9.4. Review of original data

When reviewing the original data we found that the increasing solids loading and hence change in acoustic velocity resulted in changes in the arrival time of the signal and a reduction in the magnitude of the pulse amplitude due to attenuation.

As the pulse arrival time moved, the pulse echo signal moved in relation to the other echoes in the system. With a sufficient shift in arrival time the pulse clashed with other echoes and was either lost or misdeteected. This was



especially problematic as the amplitude of the pulse was reduced and the pulse became the same order of magnitude as the noise. This caused the detector to falsely identify the wrong signal and hence the spectra become highly variable.

10. Discussion

10.1. Ultrasonic velocity of constituent materials

McClements and Povey provide a useful application of the first principles scattering model to the phenomenological method, by splitting the scattering coefficients of the ECAH model between the effective compressibility and effective density of the Urlick equations

$$\alpha_{exc} = -\frac{3\phi}{2k_c^2 r^3} \text{Re}[A_0 + 3A_1]$$

This allows the thermal and visco-inertial mechanisms to be related individually to the compressibility and density. Hence it is possible to show that the compressibility is linked to thermal effects

$$\kappa_{eff} = \kappa \cdot \left(1 - \frac{3i\phi A_0}{k_c^3 r^3} \right)$$

And density is linked to visco-inertial effects

$$\rho_{eff} = \rho \cdot \left(1 - \frac{9i\phi A_1}{k_c^3 r^3} \right)$$

$$\frac{\omega}{k^*} = \sqrt{\frac{1}{\kappa_{eff} \cdot \rho_{eff}}}$$

For less dense materials the thermal mechanism becomes important for both the attenuation and velocity components. This is demonstrated in the results where the aluminium hydroxide velocity has a negative velocity co-efficient with concentration, compared to the denser zirconium dioxide and stainless steel, which have positive co-efficients.

Hence, in the complex mixture, the effect of increasing stainless steel and zirconium dioxide concentration is offset by the increase in aluminium hydroxide concentration - the result being that acoustic velocity is not a linear function of overall solids loading.

10.2. The reflection coefficient

The reflection co-efficient method was proposed as a method of studying the density. However, experience showed that the data was very variable and drifted over time.



Plotting the value of the reflection coefficient against changes in density for a quartz delay line in a liquid (as shown below), and comparing the change with the change in velocity for a similar density change, shows that the magnitude of the relative change in reflection co-efficient is substantially smaller than the velocity.

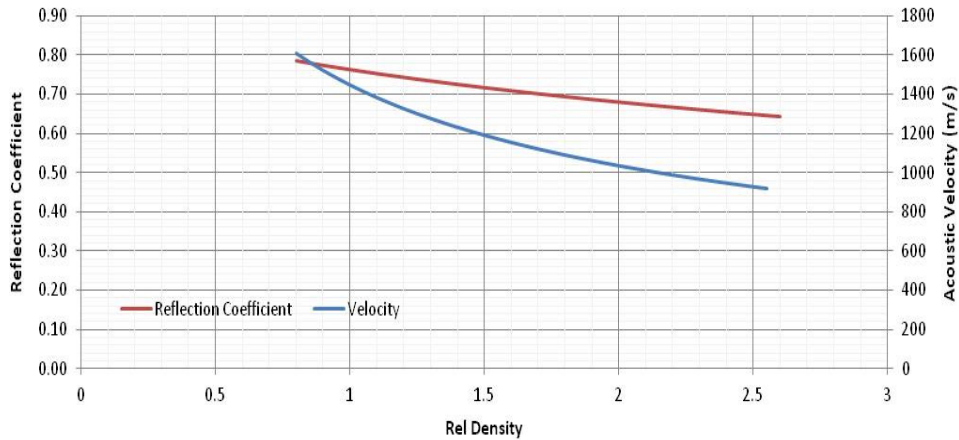


Figure 39 The Effect of changing liquid density on the reflection coefficient and ultrasound velocity

This, coupled with the high sensitivity of the technique to disturbances on the face of the transducer, ie bubbles etc, makes the method relatively noisy. It could be argued that using a fixed value of reflection co-efficient, and using the measured forward velocity to produce a rough measure of density, would be as accurate. Since the ultrasound velocity can be used to calculate solids loading directly through the equation this diminishes the usefulness of this technique.

11. Conclusions

- Using ultrasound to detect the composition of complex mixtures is complicated by the presence of materials in which the physical properties result in velocities which increase or decrease with concentration, hence effectively cancelling out the impact of overall changes in solids loading on the ultrasound velocity
- The reflection co-efficient measurement is very ill posed with a high degree of sensitivity to noise and a relative insensitivity to changes in density. The noise is due to the short range action of the wave in the liquid and the relative small integration area of the transducer
- Using multiple reflective echoes from the back face of the transducer delay line would improve this but requires more sophisticated measurement algorithms to identify the correct signals in the other echoes
- The diameter of the transducer has a large impact of the quality of the reflection measurement but the larger transducer used was still not large enough to have sufficient integration
- Ultrasonic through transmission measurements will always be better than pulse echo in the same sensor geometry



12. Recommendations for further work

- Further development of this method should concentrate on forward through transmission measurements with separate transmit and receive transducer
- The larger the transducers the better for the measurements
- Using a high concentration measurement method such as the Hi-C version of the Ultrasound Spectrometer would provide better data quality
- For reflectance measurements use a softer delay line material to reduce the impedance mismatch between the delay line and the liquid and hence increase the reflection co-efficient sensitivity to density changes.

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14. Appendix 1

Table 8 Composition of sodium nitrate solution

No of sample	Conc [v%]	Volume [ml]	Density [g/cm ³]	NaNO ₃ [v]	NaNO ₃ [g]	NaNO ₃ [mol]	NaNO ₃ [mol/kg]	mass	bulk density
1	0.10	350.00	2.26	0.35	0.79	0.01	0.03	350.44	1.00
2	0.50	350.00	2.26	1.75	3.96	0.05	0.13	352.21	1.01
3	1.00	350.00	2.26	3.50	7.91	0.09	0.27	354.41	1.01
4	2.00	350.00	2.26	7.00	15.83	0.19	0.53	358.83	1.03
5	3.00	350.00	2.26	10.50	23.74	0.28	0.80	363.24	1.04
6	5.00	350.00	2.26	17.50	39.57	0.47	1.33	372.07	1.06
7	7.00	350.00	2.26	24.50	55.39	0.65	1.86	380.89	1.09
8	10.00	350.00	2.26	35.00	79.14	0.93	2.66	394.14	1.13
9	15.00	350.00	2.26	52.50	118.70	1.40	3.99	416.20	1.19
10	20.00	350.00	2.26	70.00	158.27	1.86	5.32	438.27	1.25
11	25.00	350.00	2.26	87.50	197.84	2.33	6.65	460.34	1.32
12	30.00	350.00	2.26	105.24	237.97	2.80	8.00	482.73	1.38
13	37.50	350.00	2.26	131.56	297.46	3.50	10.00	515.90	1.47



Table 9 Composition of aluminium hydroxide suspensions in water

Density of Al(OH) ₃ [g/cm ³]	Al(OH) ₃ [g]	Al(OH) ₃ [mol]	AL(OH) ₃ [mol/kg]	bulk density²
2.42	0.85	0.01	0.03	1.00
2.42	4.24	0.05	0.16	1.01
2.42	8.47	0.11	0.31	1.01
2.42	16.94	0.22	0.62	1.03
2.42	25.41	0.33	0.93	1.04
2.42	42.35	0.54	1.55	1.07
2.42	59.29	0.76	2.17	1.10
2.42	84.70	1.09	3.10	1.14
2.42	127.05	1.63	4.65	1.21
2.42	169.40	2.17	6.21	1.28
2.42	211.75	2.71	7.76	1.36



Table 10 Composition of zirconium dioxide suspension

Density of ZrO2 [g/cm ³]	ZrO2 [g]	ZrO2 [mol]	ZrO2 mol/kg]	bulk density³
5.68	1.99	0.02	0.05	1.00
5.68	9.94	0.08	0.23	1.02
5.68	19.88	0.16	0.46	1.05
5.68	39.76	0.32	0.92	1.09
5.68	59.64	0.48	1.38	1.14
5.68	99.40	0.81	2.30	1.23
5.68	139.16	1.13	3.23	1.33
5.68	198.80	1.61	4.61	1.47
5.68	298.20	2.42	6.91	1.70
5.68	397.60	3.23	9.22	1.94
5.68	497.00	4.03	11.52	2.17



Table 11 Composition of stainless steel suspension

Density Stainless Steel 316 SS [g/cm ³]	Stainless Steel 316 SS [g]	Stainless steel [mol]	Stainless steel [mol/kg]	bulk density 4
7.96	2.79	0.01	0.03	1.01
7.96	13.93	0.04	0.13	1.03
7.96	27.86	0.09	0.25	1.07
7.96	55.72	0.18	0.50	1.14
7.96	83.58	0.26	0.76	1.21
7.96	139.30	0.44	1.26	1.35
7.96	195.02	0.62	1.76	1.49
7.96	278.60	0.88	2.52	1.70
7.96	417.90	1.32	3.78	2.04
7.96	557.20	1.76	5.04	2.39
7.96	696.50	2.20	6.30	2.74



Table 12 Composition of suspension of Al(OH)₃ in sodium nitrate aqueous solution

Density of NaNO ₃ solutions [g/cm ³]	Mass of the solution [g]	Al(OH) ₃ in NaNO ₃ solution [g]	bulk density 5
1.17	409.50	0.99	1.17
1.17	409.50	4.95	1.18
1.17	409.50	9.91	1.19
1.17	409.50	19.82	1.20
1.17	409.50	29.73	1.22
1.17	409.50	49.55	1.25
1.17	409.50	69.37	1.29
1.17	409.50	99.10	1.34
1.17	409.50	148.65	1.42
1.17	409.50	198.20	1.50
1.17	409.50	247.75	1.59



Table 13 Composition of Complex mixtures

Concentration of mixture [%]	Al(OH) ₃ [g]	ZrO ₂ [g]	SS 316 [g]	Density of NaNO ₃ aq	Mass of NaNO ₃ to prepare 350 ml aq solution d=1,17	bulk density 5
1	4.34	7.31	3.04	1.17	51.24g	1.200398
2	8.69	14.62	6.08	1.17	51.24g	1.23081
4	17.78	29.83	12.43	1.17	51.24g	1.294207
6	27.82	46.68	19.42	1.17	51.24g	1.364286
8	37.1	62.24	25.9	1.17	51.24g	1.429074
10	47.42	79.53	33.14	1.17	51.24g	1.501173

15. Appendix 2

Other information

The liquid density in AY-102 is approximately 1.17 g/mL (Hu 2007).

PNNL-18327 Estimate of the Distribution of Solids within Mixed Hanford Double-Shell Tank AZ-101: Implications for AY-102, is correct

ICD-19 specifies bulk density of < 1.5 g/mL However, it was decided that using a density of 1.55 g/ml will be sufficiently bounding because it is the sediment density which will typically exceed the feed HLW slurry density. This number is taken from PNNL-18327.

Representative liquid temperature is approximately 113°F (Hu and Barker 2003), and the dynamic viscosity of AY-102 liquid at that temperature is approximately 3 cP (TWINS data from Attachment B, Warrant 2001; see Onishi and Wells 2004). The sediment density is approximately 1.55 g/mL (Hu 2007),

The sediment density is approximately 1.55 g/mL (Hu 2007), and the UDS have a bulk density of 2.5 g/mL (Onishi and Wells 2004). With Equations (2.1) and (2.2), the UDS content in the sediment can be estimated at 29% by volume, 46% by mass.